

Drought – Year 2

CACHE CREEK OFF-CHANNEL AGGREGATE MINING PONDS – 2021 MERCURY MONITORING

FINAL Report

Monitoring and Report by

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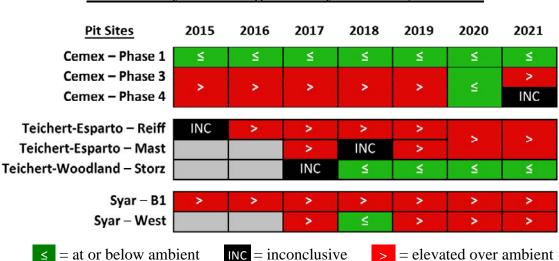
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SUMMARY OF THE 2021 MONITORING AND ITS FINDINGS

• This Fall 2021 monitoring was the seventh year of fish mercury testing (Year 7) for four offchannel aggregate mining ponds, adjacent to lower Cache Creek between Capay and Woodland: Cemex–Phase 1, Cemex–Phase 3-4, Teichert-Esparto–Reiff, and Syar–B1 ponds. The monitoring was initiated in 2015. Three other ponds were added to the monitoring program in 2017: Teichert-Esparto–Mast, Teichert-Woodland–Storz, and Syar–West ponds. For these ponds, 2021 was Year 5 of mercury monitoring. The monitoring is required by Section 10-5.517 of the Yolo County Code, which was revised and updated in December 2019. That Ordinance requires 5 years of annual pre-reclamation mercury monitoring for mining ponds, and then biannual monitoring for 10 years following reclamation to permanent water bodies. The fish monitoring includes new sampling each year and assessment of mercury levels in relation to comparable baseline fish data from Cache Creek.



Fish Mercury Monitoring Summary – All Sites, 2015-2021

- As summarized in the table above, the Cemex–Phase 3, Teichert-Esparto, Syar–B1, and Syar–West pits have had "three or more years out of five elevated over the ambient". The program requires that the County take certain steps following a third year in five of exceedance for a pit:
 - Require an additional <u>five years of fish mercury monitoring and water column profiling</u>. This pattern will continue until a lake is found to be at or under the ambient for a five-year period; the regulations also allow the County to require continued monitoring during mining. Comparison monitoring during this time will also be conducted at control/reference sites.
 - Require <u>Expanded Analysis</u> including expanded water column profiling of all relevant water quality parameters (multiple times per year rather than a single time per year) and one-time

bottom sediments analysis. Expanded analyses, as set out in the Ordinance, began in 2018 and are reported separately – see summary status tables of these activities at the end of this section.

- Once the reports are completed, the County will notify individual operators of results in individual ponds that require <u>Lake Management Plans (LMPs</u>). The information in the fish monitoring, water column profiling, and bottom sediments reports will then be used to identify mercury control methods to reduce fish mercury levels and prepare required LMPs.
- <u>Implementation of the LMP</u> is required within three years of completion of the expanded monitoring. Management controls may differ for different pits based on site conditions; and may differ during mining, while idle, and post-mining. LMPs may be multi-part or phased to reflect this. Fish monitoring and water column profiling will continue, per the regulations, for a minimum of five more years. Required periodic analysis of ambient conditions will also continue.
- For environmental mercury, fish consumption is by far the most significant exposure route for people and wildlife. Fish also provide an accurate measure of relative mercury exposure levels over time, and for comparison between ponds and Cache Creek. For these reasons, the mercury monitoring program for Yolo County aggregate mining ponds focuses on fish.
- A variety of collection techniques were used to obtain samples of the fish found in each of these ponds, including seines, gill nets, baited setlines, dip nets, and angling. Large, angling-sized fish were tested individually for fillet muscle mercury, relevant to human consumption. Small, young, "biosentinel" fish were analyzed whole-body, relevant to wildlife consumption and inter-annual comparisons, in replicate multiple-individual composite samples.
- Samples of both large and small fish of multiple species, as available and as possible under severe drought conditions, were collected from the seven identified ponds (in their current configurations). A total of 142 adult, angling-sized fish (mainly bass) were sampled individually for fillet muscle mercury analysis in this 2021 monitoring. Additionally, a total of 537 small, young, biosentinel fish were split into 87 multi-individual, whole fish composite samples by site, species, and size. These were also analyzed for mercury.
- The new 2021 data are compared with results from 2015-2020, and with the most closely corresponding 'baseline' and historic fish collections conducted previously in Cache Creek (from the stretch of creek within the planning and aggregate-mining area). As in prior years, the ponds sampled in 2021 were found to show distinct, individual mercury signatures that were broadly consistent across the different fish types tested.
- Cemex Phase 1 Pond: Twenty adult Largemouth Bass were sampled, and multiple composite samples were taken of young-of-year Mosquitofish, juvenile Largemouth Bass, and juvenile Green Sunfish. The Cemex–Phase 1 Pond remained one of the lowest mercury ponds of the sites being monitored. Concentrations in 2021 were statistically similar to or lower than corresponding baseline Cache Creek samples of same or similar species and sizes. The Phase 1 Pond was therefore not found to be "elevated in three or more years of five" and did not trigger seasonal water column profiling and consideration of mercury management. However, the

overall low mercury status of this pond, and the fish mercury trends over the years monitored in relation to operations changes, made it a key comparison for management insights for the elevated ponds. It was chosen as a control/reference site, as required for the "expanded analysis" parts of the monitoring, and has been part of that work since 2018.

- Cemex Phase 3 Pond: The former Phase 3-4 Pond was divided into two separate parts in late 2020, with active mining continuing in the eastern, Phase 4 Pond and the western, Phase 3 subbasin becoming an isolated, relatively undisturbed pond. Mercury monitoring continued this year in both the Phase 3 and Phase 4 Ponds. In 2021, twenty adult Largemouth Bass were sampled from Phase 3, and multiple composite samples were taken of young-of-year Mosquitofish, juvenile Green Sunfish, and juvenile Largemouth Bass. Except for the Mosquitofish samples, all of the 2021 Phase 3 fish were elevated in mercury on average. As the former combined Phase 3-4 Pond was found to be "elevated for three or more years of five" over creek baselines (2015-2019, all 5 years), that triggered the addition of "expanded analysis" and development of a mercury management plan. Expanded analysis work began in the Phase 3-4 Pond in 2018, with seasonal water column profiling of a range of relevant constituents and testing of bottom sediments, and is presented in accompanying reports. After Phase 3-4 was split into two parts, this work continued in both ponds in 2021.
- Cemex Phase 4 Pond: In 2021, twenty adult Largemouth Bass were sampled from Phase 4 (the active-mining pond), and multiple composite samples were taken of young-of-year Mosquitofish, juvenile Green Sunfish, and juvenile Largemouth Bass. <u>Overall mercury results were 'inconclusive' in 2021, with some samples relatively elevated and others, including the key adult bass, similar to or below baseline comparisons. However, as the former combined Phase 3-4 Pond was found to be "elevated for three or more years of five" over creek baselines (2015-2019, all 5 years), that triggered the addition of "expanded analysis" and, following a period of data gathering, development of a mercury management plan. Expanded analysis work began in the Phase 3-4 Pond in 2018, with seasonal water column profiling of a range of relevant constituents and testing of bottom sediments, and is presented in accompanying reports. After Phase 3-4 was split into two parts, this work continued in both ponds in 2021. Future status of Phase 4 will depend on cummulative monitoring results.
 </u>
- Teichert Esparto Pond: The previously separate Reiff and Mast Ponds were combined by Teichert into a single large Esparto Pond in early 2020, by excavating parts of dividing levees. Monitoring continued in the combined pond. With the second year drought conditions of 2021, the Esparto Pond began the sampling season interconnected but dried down into disconnected, shallow basins by summer. Samples of adult Largemouth Bass (20), White Catfish (6), and Common Carp (2) were taken from the central basin. Small, young-of-year fish were sampled from two basins; these included Mosquitofish, juvenile Largemouth Bass, juvenile Green Sunfish, and Red Shiners. Average mercury levels in all sample types showed increases in 2021, particularly in the small, young fish which are most representative of recent conditions. This site remained highly elevated in mercury in 2021. All of the fish sample types averaged significantly higher mercury than corresponding Cache Creek baseline samples. Similar results from previous years in the Reiff and Mast ponds triggered the collection of additional information ("expanded analysis") to help guide development of a mercury management plan.

Water column profiling and collection of bottom sediment samples began in May 2018 and are the subject of accompanying reports.

- Teichert-Woodland Storz Pond: Drought conditions dried this pond down into two disconnected sub-basins in 2021. A sample of 18 small adult bass (the main size present) and 2 larger bass was taken. Small, young-of-year fish samples included multiple composites of Mosquitofish, juvenile Largemouth Bass, and juvenile Bluegill Sunfish. Mosquitofish were taken from both sub-basins to check for potential localized differences; no difference was found. The adult bass and all of the small fish samples remained in the lower range of mercury levels for this site, which has consistently been the first or second lowest mercury pond in the monitoring program. Relative to Cache Creek comparison data, Storz Pond continued to rank as "not elevated over baseline" in 2021 and is not flagged for expanded analysis or management planning. One-time-per-year routine water profiling was added to the monitoring in 2020, following recent revisions of the mining ordinance. With five years now with low mercury status, this site can shift to a schedule of fish and water testing once each two years rather than annually.
- Syar B1 Pond: Drought conditions affected both Syar Ponds significantly in 2020-2021, dropping water levels to far below normal. In addition to impacts on the ponds, it also made access and some fish collections difficult or not possible. Sixteen adult Largemouth Bass were sampled, one adult Bluegill Sunfish and for the first time here, four Channel Catfish. Young-of-year small fish were sampled with multiple composites of Mosquitofish, juvenile Largemouth Bass, and juvenile Bluegill Sunfish. Fish mercury was up in 2021 over the previous year for most sample types but remained significantly lower than the peak levels found here in 2015-2016. Despite the relatively lower levels in recent years, B1 Pond fish mercury in 2021 remained significantly higher on average than most baseline Cache Creek comparisons, most importantly in the Largemouth Bass. Because of the overall status of the B1 Pond as "elevated over baseline in three or more years of five" (all years since 2015), water column profiling and collection of bottom sediments was started here in 2018, in support of the development of a lake management plan. That work is detailed in accompanying reports.
- Syar West Pond: Drought conditions affected both Syar Ponds significantly in 2020-2021, dropping water levels to far below normal. In addition to impacts on the ponds, it also made access and some fish collections difficult or not possible. Thirteen adult Largemouth Bass were sampled. Young-of-year small fish samples included multiple composites of Mosquitofish, juvenile Bluegill Sunfish, and juvenile Largemouth Bass. Fish mercury was up in 2021 over the previous year for most sample types; most importantly, the adult bass samples averaged higher than the bass from all previous years. Syar-West fish mercury in 2021 was significantly higher on average than most baseline Cache Creek comparisons, as in 2017, 2019, and 2020. Because of the overall status of the West Pond as "elevated over baseline in three or more years of five" as of 2020, expanded analysis and development of a lake management plan is required. Expanded analyses have in fact been conducted here since 2018, as a second control/reference site. This pond is far deeper than the other ponds and is representative of the range of final depths projected at several of the sites. With elevated fish mercury status as of 2020, this work will help in the development of a lake management plan.

Status of Other Components of the Mercury Monitoring Program

Pit Sites	2015	2016	2017	2018	2019	2020	2021
Cemex – Phase 1 (control)				- ✓	√		- ✓
Cemex – Phase 3				1	~	1	✓
Cemex – Phase 4				, in the second s		, i i i i i i i i i i i i i i i i i i i	 ✓
Teichert-Esparto – Reiff				\checkmark	✓	1	1
Teichert-Esparto – Mast							
Teichert-Woodland – Storz							(1x)
					-		
Syar – B1				\checkmark	\checkmark	\checkmark	\checkmark
Syar – West (control to 2020)					\checkmark	 ✓ 	1

<u>Water Column Profiling</u> (elevated sites and controls)

Bottom Sediment Collections (single event, elevated sites and controls)

Pit Sites	2015	2016	2017	2018	2019	2020	2021
Cemex – Phase 1 (control)					(
Cemex – Phase 3-4				٧	(
Teichert-Esparto – Reiff				~			
Teichert-Esparto – Mast	î de la centre de la						
Teichert-Woodland – Storz							
						_	
Syar – B1				v			
Syar – West (deep control)				۷			

Reports Completed

Report	2015	2016	2017	2018	2019	2020	2021
Fish Mercury Monitoring	Final	Final	Final	Final	Final	Draft	Draft
Water Column Profiling			· · · · · · · · · · · · · · · · · · ·	Final	Final	Draft	Draft
Bottom Sediments (1x)				Fir	nal		

INTRODUCTION

This monitoring was conducted for Yolo County in the summer and fall of 2021, to provide ongoing fish mercury information from a set of aggregate mining ponds located adjacent to lower Cache Creek. The monitoring was triggered by Section 10.5.517 of the Yolo County Surface Mining Reclamation Ordinance (Yolo County Code), which was enacted originally in 1996. Earlier reports (2015-2018) have gone into detail about the County's history with the mercury issue, placing the first years of monitoring into context with the 1996 Ordinance. In December 2019, the County adopted a comprehensive update to the Cache Creek Area Plan (CCAP), which included a full revision of this code section (Yolo County Code 2019), incorporating new findings and issues identified since 1996. Mercury monitoring and reporting since then, including this 2021 report, complies with the updated ordinance requirements. The complete 2019 Ordinance is attached, as Appendix A, at the end of this report. Below, in this introduction, parts that most directly affect this fish mercury monitoring program are excerpted and discussed. Ordinance text is shown in **bold italics**, with discussion in plain text.

Yolo County, CA Code of Ordinances, Sec. 10-5.517 Dec 2019 Revision – Mercury Bioaccumulation in Fish.

As part of each approved long-term mining plan involving wet pit mining to be reclaimed to a permanent pond, lake, or water feature, the operator shall maintain, monitor, and report to the Director according to the standards given in this section. Requirements and restrictions are distinguished by phase of operation as described below.

(a) MERCURY PROTOCOLS. The Director shall issue and update as needed "Lower Cache Creek Off-Channel Pits Mercury Monitoring Protocols" (Protocols), which shall provide detailed requirements for mercury monitoring activities. The Protocols shall include procedures for monitoring conditions in each pit lake, and for monitoring ambient mercury level in the lower Cache Creek channel within the CCAP planning area, as described below.

Mercury Protocols for these tasks were developed before the monitoring program began in 2015 and were followed through 2019. The protocols were revised, expanded, and updated (Slotton 2021) to support the 2019 revision of the County Code Ordinance.

(b) AMBIENT MERCURY LEVEL. The determination of the ambient or "baseline" fish mercury level shall be undertaken by the County every ten years in years ending in 0. This analysis shall be undertaken by the County for use as a baseline of comparison for fish

mercury testing conducted in individual wet mining pits.

The most recent creek sampling targeted to the aggregate mining zone was conducted in 2011 and 2012 (Slotton and Ayers 2013). Data from other earlier studies that coincidentally fell within the planning area have also been used for comparisons. Another full Cache Creek Baseline set of fish collections will be conducted some time in the next few years, as possible dependent on the presence of sufficient water and fish in the creek in these drought times.

(c) PIT MONITORING.

(1) Mining Phase (including during idle periods as defined in SMARA). The operator shall monitor fish and water column profiles in each pit lake once every year during the period generally between September and November for the first five (5) years after a pit lake is created. Fish monitoring should include sport fish where possible, together with other representative species that have comparison samples from the creek and/or other monitored ponds. Sport fish are defined as predatory, trophic level four fish such as bass, which are likely to be primary angling targets and have the highest relative mercury levels. The requirements of this subsection apply to any pit lake that is permanently wet and navigable by a monitoring vessel. This monitoring began in 2015, at four aggregate mining ponds: Cemex–Phase 1, Cemex–Phase 3-4, Teichert-Esparto–Reiff, and Syar–B1. Three other ponds were added to the monitoring program in 2017: Teichert-Esparto-Mast, Teichert-Woodland-Storz, and Syar–West ponds. In 2020, Reiff and Mast were combined into a single Teichert-Esparto Pond and the Cemex Phase 3-4 Pond was separated into two distinct ponds (Phase 3 and Phase 4), making seven total ponds monitored at this time. An important focus of the monitoring has been largemouth bass, which are present in most of the ponds.

If, in the initial five (5) years after the pit lake is created, the applicable response threshold identified in subsection (e) is exceeded in any three (3) of five (5) monitoring years, the operator shall, solely at their own expense, undertake expanded analysis pursuant to subsection (f) and preparation of a lake management plan pursuant to subsection (g).

Earlier in the monitoring program, three of the ponds were found to have fish mercury above baseline creek comparison levels in three or more years: Cemex–Phase 3-4, Teichert-Esparto–Reiff, and Syar–B1. Beginning in 2018, "expanded analysis" testing was initiated at these three ponds, and also at Cemex–Phase 1 as a lower mercury control/reference site (see (f)(1)) and at Syar–West as a deep pond control. The Syar–West Pond has since shifted into the elevated category, requiring this work. With the recent splitting of the Cemex–Phase 3-4 pond into two distinct ponds, both have been included during 2021. The expanded analyses have included one-time sediment testing and seasonal, ongoing water column profiling as specified in Section (f). This work has been in the data gathering stage. Findings are intended to now guide the preparation of realistic lake management plans.

For future, post-mining years: monitoring and potential lake management requirements:

- (2) <u>Reclamation Phase</u>. No monitoring is required after mining has concluded, during the period that an approved reclamation plan is being implemented, provided reclamation is completed within the time specified by SMARA or the project approval, whichever is sooner.
- (3) <u>Post-Reclamation Phase</u>. After reclamation is completed, the operator shall monitor fish and water column profiles in each pit lake at least once every two (2) years during the period of September-November for ten (10) years following reclamation. Monitoring shall commence in the first calendar year following completion of reclamation activities. If fish monitoring results from the post-reclamation period exceed the applicable response threshold described in subsection (e) or, for ponds that have implemented mitigation management, results do not exhibit a general decline in mercury levels, the operator shall, solely at their own expense, undertake expanded analysis pursuant to subsection (f) and preparation of a lake management plan pursuant to subsection (g).
- (4) <u>Other Monitoring Obligation</u>. If monitoring conducted during both the mining and post-reclamation phase did not identify any exceedances of the ambient mercury level for a particular pit lake, and at the sole discretion of the Director no other relevant factors substantially support that continued monitoring is merited, the operator shall have no further obligations.

(e) RESPONSE THRESHOLDS.

(1) <u>Fish Consumption Advisory</u>. If at any time during any phase of monitoring the pit lake's average sport fish tissue mercury concentration exceeds the Sport Fish Water Quality Objective (as of 2019, the level was 0.2 mg/kg), the operator shall post fish consumption advisory signs at access points around the lake and around the lake perimeter. Catch-and-release fishing may still be allowed. The sites have been posted. Catch and release fishing has been common at the Syar.

The sites have been posted. Catch and release fishing has been common at the Syar ponds and not at the others.

(2) <u>Mining Phase Results</u>. If, during the mining phase of monitoring, the pit lake's average fish tissue mercury concentration exceeds the ambient mercury level for any three (3) of five (5) monitoring years, annual monitoring shall continue for an additional five (5) years, and the operator shall undertake expanded analysis pursuant to subsection (f) and preparation of a lake management plan pursuant to subsection (g).

Fish monitoring has continued at all pit lakes found to be elevated in fish mercury, and at those still in the initial five years of testing. As noted above in (c)(1), expanded analysis is in progress at the identified elevated mercury sites and control/reference ponds, gathering data to help develop lake management plans.

For future, post-mining years: monitoring and potential lake management requirements:

(3) <u>Post-Reclamation Phase Results</u>. If during the first ten (10) years of the post-reclamation phase of monitoring, the pit lake's average fish tissue mercury concentration exceeds the ambient mercury level for any three (3) of five (5) monitoring years, biennial monitoring shall continue for an additional ten (10) years, and the operator shall undertake expanded analysis pursuant to subsection(f) and preparation of a lake management plan pursuant to subsection (g).

(f) EXPANDED ANALYSIS.

(1) <u>General</u>. If, during the mining or post-reclamation phase, any pit lake's average fish tissue mercury concentration exceeds the ambient mercury level for any three (3) years, the operator shall undertake expanded analyses. The analysis shall include expanded lake water column profiling (a minimum of five profiles per affected wet pit lake plus one or more nonaffected lakes for control purposes) conducted during the warm season (generally May through October) in an appropriate deep profiling location for each pit lake. The following water quality parameters shall be collected at regular depth intervals, from surface to bottom of each lake, following protocols identified in subsection (a): temperature, dissolved oxygen, conductivity, pH and oxidation-reduction potential (ORP), turbidity or total suspended solids, dissolved organic matter, and algal density by Chlorophyll or Phycocyanin. The initial analysis shall also include one-time collections of fine grained (clay/silt) bottom sediments from a minimum of six well distributed locations for each affected lake, and from one or more non-affected lakes for control purposes, to be analyzed for mercury and organic content.

The current expanded analysis work is guided by these directions. Data gathering on these various, potentially important parameters began in 2018; 2021 data are presented in the companion report on water profiling.

(2) Scope of Analysis. The purpose of the expanded analyses is to identify and assess potential factors linked to elevated methylmercury production and/or bioaccumulation in each pit lake. In addition to the analyses described in subsection (f)(1) above, the analysis should also consider such factors as: electrical conductivity, bathymetry (maximum and average depths, depth-to-surface area ratios, etc.), and trophic status indicators (concentrations, Secchi depth, chlorophyll a, fish assemblages, etc.). Additional types of testing may be indicated and appropriate if initial results are inconclusive.

These suggestions are all being followed in the expanded analysis work.

(3) <u>Use of Results</u>. The results of the expanded analyses undertaken pursuant to this subsection shall be used to inform the preparation of a lake management plan described below under subsection (g).

As noted above, this work has been in the data gathering stage. Findings are intended to now guide the preparation of realistic lake management plans. This, and future

management and monitoring activities, are described in these final Ordinance excerpt sections:

(g) LAKE MANAGEMENT ACTIVITIES.

- (1) <u>General</u>. If monitoring conducted during the mining or post-reclamation phases triggers the requirement to undertake expanded analysis and prepare and implement a lake management plan, the operator shall implement lake management activities designed by a qualified aquatic scientist or equivalent professional acceptable to the Director, informed by the results of subsection (f). Options for addressing elevated mercury levels may include (A) and/or (B) below at the Director's sole discretion and at the operator's sole expense.
 - (A) <u>Lake Management Plan</u>. Prepare a lake management plan that provides a feasible, adaptive management approach to reducing fish tissue mercury concentrations to at or below the ambient mercury level. Potential mercury control methods could include, for example: addition of oxygen to or physical mixing of anoxic bottom waters; alteration of water chemistry (modify pH or organic carbon concentration); and/or removal or replacement of affected fish populations. The lake management plan may be subject to external peer review at the discretion of the Director. Lake management activities shall be appropriate to the phase of the operation (e.g., during mining or post-reclamation). The Lake Management Plan shall include a recommendation for continued monitoring and reporting. All costs associated with preparation and implementation of the lake management plan shall be solely those of the operator. Upon acceptance by the Director, the operator shall immediately implement the plan. The lake management plan shall generally be implemented within three years of reported results from the expanded analyses resulting from subsection (f). If lake management does not achieve acceptable results and/or demonstrate declining mercury levels after a maximum of three years of implementation, at the sole discretion of the Director, the operator may prepare an alternate management plan with reasonable likelihood of mitigating the conditions.
 - (B) <u>Revised Reclamation Plan</u>. As an alternative to (A), or if (A) does not achieve acceptable results and/or demonstrate declining mercury levels after a maximum of three years of implementation, at the sole discretion of the Director, the operator shall prepare and submit revisions to the reclamation plan (including appropriate applications and information for permit amendment) to fill the pit lake with suitable fill material to a level no less than five (5) feet above the average seasonal high groundwater level, and modify the end use to agriculture, habitat, or open space at the discretion of the Director, subject to Article 6 of the Mining Ordinance and/or Article 8 of the Reclamation Ordinance as may be applicable.

(2) IMPLEMENTATION OBLIGATIONS.

- (A) If a lake management plan is triggered during the mining or post-reclamation phase and the subsequent lake management activities do not achieve acceptable results and/or demonstrate declining mercury levels, the operator may propose different or additional measures for consideration by the Director and implementation by the operator, or the Director may direct the operator to proceed to modify the reclamation plan as described in subsection (g)(1)(B).
- (B) Notwithstanding the results of monitoring and/or lake management activities during the mining phase, the operator shall, during the post-reclamation phase, conduct the required ten years of biennial monitoring.
- (C) If monitoring conducted during the post-reclamation phase identifies three monitoring years of mercury concentrations exceeding the ambient mercury level, the operator shall implement expanded analyses as in subsection (f), to help prepare and implement a lake management plan and associated monitoring.
- (D) If subsequent monitoring after implementation of lake management activities, during the post-reclamation phase, demonstrates levels of fish tissue mercury at or below the ambient mercury level for any three monitoring years (i.e., the management plan is effective), the operator shall be obligated to continue implementation of the plan and continue monitoring, or provide adequate funding for the County to do both, in perpetuity.

As fish have been found to be the most straightforward, clear measure of methylmercury exposure and bioaccumulation in aquatic systems, this monitoring focuses on fish. All seven of the currently identified ponds (Table A, Figure A) were monitored for fish mercury in 2021. Five of the ponds have been monitored since 2015 and, for them, this was Year 7 of sampling: Cemex–Phase 1, Cemex Phase 3 and Phase 4 (formerly Phase 3-4), Teichert–Esparto (formerly Reiff and Mast), and Syar–B1. Two additional ponds were added to the monitoring in 2017; for these, 2021 was Year 5: Teichert-Woodland–Storz and Syar–West. Both large and small fish samples of multiple species, as available, were collected and analyzed from each of the ponds.

The purpose of this report is to present the new 2021 fish mercury data from the tested aggregate mining ponds and, for each pond, to compare levels to similar baseline samples taken from the planning area of Cache Creek in 2011-2012 and in earlier studies. A key objective is to help the mining operators and Yolo County determine if specific pond sites are falling below, at, or above

fish mercury concentrations found in adjacent Cache Creek. This will help guide future reclamation and, if necessary, pond management.

The factors that influence the production of methylmercury and its uptake by fish are complex and can change from one year to the next, often leading to a range of fish mercury levels over time rather than some absolute value. Because of this, the Ordinance states that multiple years of data are needed to make assessments. Therefore, another objective is to compare this year's data (2021) with monitoring results found at the same sites in the previous monitoring years (2015-2020).

In the sections below we will discuss the methods used, followed by a presentation of the 2021 fish mercury data, by individual pond site. Each data table is accompanied by a matching figure with the same number that graphically shows the information. For each site, we first present the analytical results from each individual large fish sample and each small fish composite sample. Then we show the new data in reduced form (means, error bars, etc.) for each sample type and compare to 2015-2020 same-site findings and the most closely comparable historic creek data. For creek comparisons, we are focusing on historic data specifically from the planning / aggregatemining section of the creek, roughly between River Mile 28 (below the Capay diversion dam) and River Mile 15 (app. 1 km below County Road 94B). In particular, these include the 2011 Baseline collections from River Mile 15 (RM15), RM20, and RM28, which were conducted specifically to provide comparable samples for the pond monitoring. In the data tables and figures, the 2011 Baseline comparison data are highlighted with bold text and outlines. Additional historic sampling that was coincidentally done within the planning region of Cache Creek includes a project around the Cache Creek Nature Preserve in 2000-2006 (RM15 and RM17 small fish) and a CalFed 1998-2000 UC Davis study of the entire Cache Creek watershed that included some fish collections in the study zone.

After individual reporting sections for each pond, a final data section consolidates summary results for each fish type, from all the sites and baseline creek samples for easy comparison. In the Discussion/Conclusions, the available pond data to-date are placed into the context of the updated Yolo County Ordinance, with next steps and recommendations. Appendix A includes the full text of the new Ordinance. Appendix B shows photos of the Fall 2021 fish mercury monitoring work.

Table A. Wet Pits Subject to Annual Mercury Monitoring

(modified from Yolo County Exhibit C)

Operator	Site	Pit	Year Mining Crossed Water Table (app)	End Reclamation Plan	Year Monitoring Began	Monitoring Year in Fall 2019
Cemex	Madison	Phase 1	< 1996	Lake and habitat	2015	Year 7
Cemex	Madison	Phase 3-4 (Phase 3 and	≤ 2002 d Phase 4 were sept	Lake and habitat arated in late 2020)	(2015-2020)	
Cemex Cemex	Madison Madison	Phase 3 Phase 4	Ĩ	Lake and habitat Lake and habitat		(Year 7) (Year 7)
Teichert	Esparto	Reiff	≤ 2002	Lake and habitat	(2015-2019)	
Teichert	Esparto	Mast (Reiff and M	2007-2008 last were combined	Lake and habitat <i>into one pond in 202</i>	(2017-2019) 20)	
Teichert	Esparto	Esparto		-		(Year 7)
Teichert	Woodland	Storz	2010-2011	Lake and habitat	2017	Year 5
Syar	Madison	B1	≤ 2002	Lake and habitat	2015	Year 7
Syar	Madison	West	≤ 2002	Lake and habitat	2017	Year 5

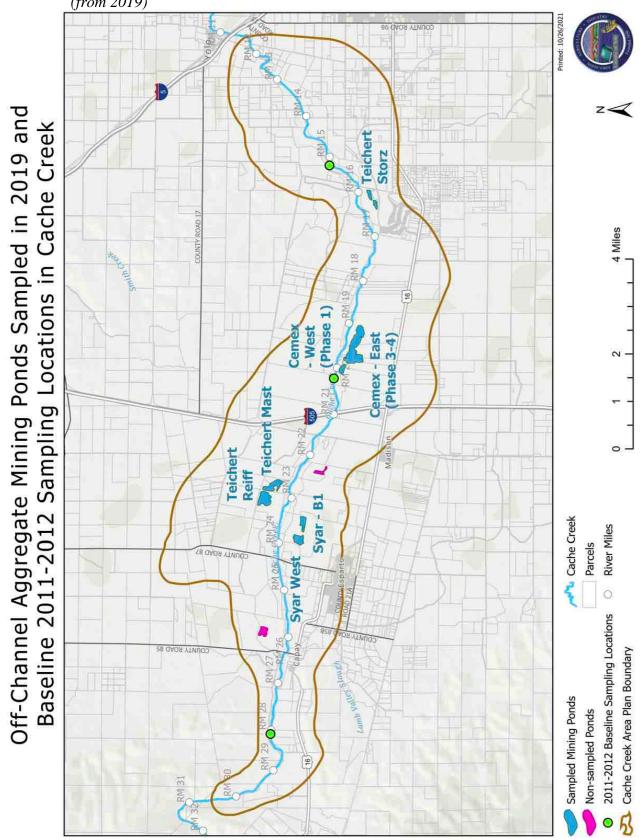


Figure A. Map of aggregate mining ponds and Cache Creek baseline monitoring locations (*from 2019*)

METHODS

Field sampling was coordinated with staff of the three mining companies: Teichert, Cemex, and Syar. Access ramps for boat launching were constructed at the Cemex ponds. We used our sampling boat to move around those ponds and collect the fish. This was not possible at the Syar and Teichert sites, which were heavily impacted by the drought, making the shorelines inaccessible to heavy equipment.

The fish samples were taken with a variety of techniques. Adult fish were collected with gill nets in a range of mesh sizes, with baited set lines laid at the bottom of ponds (catfish), by angling (bass) and, as a last resort in ponds with no boat or shore access due to the drought, by spearfishing. Small, young fish samples were collected with a variety of seines and hand nets.

Large fish were field identified, weighed and measured, and sampled for mercury analysis using a non-destructive biopsy technique we developed that allows us to return the fish back to the water in good condition (Slotton et al. 2002). In this technique, numbered sample vials are preweighed, empty, to 0.0001 g accuracy. In the field, several scales are removed from each fish on the left side above the lateral line and a small biopsy sample of app. 0.200 g (about the size of a small raisin) is taken of the fillet muscle. The sample is carefully placed into a pre-weighed vial. Vials are closed with sealing screw-tops and stored on ice in a protective vial box. Later, at the laboratory, the vials with sample pieces are again weighed and the exact weight of each sample is determined by subtracting the empty tube weight.

Small young-of-year fish, in contrast, were sacrificed for analyses, analyzed whole. Small fish were field identified, cleaned and sorted by species, bagged in labeled freezer-weight, zip-close bags with air removed, and transported on ice to the laboratory. Fish were then weighed, measured, and assembled into composite groupings of similar-sized fish for each size class. Each composite sample was frozen in doubled freezer-weight bags with water surrounding and air removed, a technique our group has found to maintain natural moisture levels through the freezing process, something that can be a major problem for small fish samples (Slotton et al.

2015). Pre-analytical processing included weighing and measuring the fish in each composite group and drying the sample to constant weight in a laboratory oven at 55 °C. Solids percentage was calculated during this process, through sequential weighings of empty weigh pans, pans with wet sample, and pans with dry sample. Dried samples were later homogenized to fine powders using a laboratory grinder.

Large fish fillet muscle samples were analyzed for mercury directly, on a wet (fresh) weight basis. Small fish composite samples were analyzed whole body, homogenized into dry powders for consistency, as described above. Dry weight results were converted to original wet/fresh weight concentrations using the calculated percentage solids values. Mercury analyses were conducted with a direct mercury analyzer system (Milestone DMA-80 evo), using EPA Method 7473.

Extensive Quality Assurance / Quality Control (QA/QC) samples were included in all analytical runs and tracked with control charts. These included, for each 20 field samples: 3 method blanks, 3 standard reference materials with certified levels of mercury, 2 aqueous calibration samples, a laboratory duplicate, a spiked field sample, a spike duplicate, and 3 continuing calibration samples. For initial machine calibration (stable for over a month), an extensive calibration was performed each month, using at least 15 aqueous calibration solutions, each run in duplicate. Calibration stability was tested each analytical run with the many aqueous and solid reference samples. QA/QC Results for this project were all well within control limits.

2021 FISH MERCURY MONITORING RESULTS

1. CEMEX – PHASE 1 POND

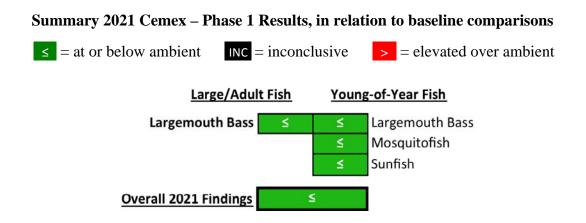


(Google Earth 10/21/20)

1. CEMEX – PHASE 1 POND (Tables and Figures 1-6)

Summary

Twenty adult Largemouth Bass were sampled, and multiple composite samples were taken of young-of-year Mosquitofish, juvenile Largemouth Bass, and juvenile Green Sunfish. Despite some relative ups and downs over the years, the Cemex–Phase 1 Pond remained one of the lowest in fish mercury of the ponds being monitored. Concentrations in 2021 were statistically similar to or lower than corresponding baseline Cache Creek samples of same or similar species and sizes. The Phase 1 Pond was therefore not found to be "elevated in three or more years of five" and did not trigger seasonal water column profiling and consideration of mercury management. However, the overall low mercury status of this pond, and the interesting changes over the years monitored in relation to operations changes, made it a key comparison for management insights for the elevated ponds. It was chosen as a control/reference site, as required for the "expanded analysis" parts of the monitoring, and has been part of that work since 2018.



This pond is the older of the 2 current Cemex ponds, dating from the 1990s. It is located just south of Cache Creek and east of Highway 505. The Phase 1 Pond is an oval shaped bowl that is approximately 400 m long and 150 m wide. In 2021, depths ranged narrowly between 5.3 and 5.8 m (17-19 feet). Phase 1 and the other Cemex ponds had their water levels maintained during the drought, presumably by adding water through the season. This pond went through some changes

over the years of monitoring. Active mining was underway in 2015. In 2016 there was little or no mining in the pond itself, but it continued to receive the silt and clay slurry effluent of the general plant operations, so the water was very turbid. In 2017, active mining was on hold at both Cemex ponds, so there was less slurry effluent to the Phase 1 Pond. Since 2018, active mining resumed at the Phase 3-4 Pond, continuing since 2020 in the Phase 4 portion, with process slurry effluent discharging to the Phase 1 Pond and generally keeping this shallow pond turbid. This (2021) was Year 7 of monitoring at this site.

We sampled the pond with a range of techniques, and were able to obtain samples of the fish species available. Large, angling-sized fish taken were: 20 Largemouth Bass (*Micropterus salmoides*). Despite fishing effort for other species, they have not been found since 2018. In previous years, we routinely took several Channel Catfish (*Ictalurus punctatus*) and White Catfish (*Ameiurus catus*). We suspect that these may have been fished out of the system (not by us; we always return biopsy-sampled fish back to the ponds in good condition). The small fish present were Mosquitofish (*Gambusia affinis*, 1-2"), juvenile Green Sunfish (1-2"), and juvenile Largemouth Bass (3-4"). Three to four multi-individual composite samples, as available, were analyzed from each of these small fish species, for 11 total composites.

In total, this added up to 20 large fish muscle samples and 11 composite small fish samples, 31 separate fish mercury samples, analyzed from the Cemex–Phase 1 Pond in the Fall 2021 monitoring. The fish metrics and analytical results from each individual large bass muscle sample can be seen in Table 1 and, graphically, in Figure 1. Then, the new data are shown in reduced form (means, error bars, etc) and compared to 2015-2020 results and the most closely comparable historic creek data (Table 2 and Figure 2). Results from the composite samples of small, young-of-year fish are similarly presented in Tables and Figures 3-6.

Large, Angling-sized Fish

Largemouth Bass (Tables/Figures 1 and 2)

The 2021 bass samples ranged in size between 224 and 302 mm (about 9-12"), with a single much larger fish (393 mm, 15.5"). Adult Bass represent the top predator fish in this region and will typically have the highest mercury levels at any given site. Mercury concentrations generally increased with fish size, as is typical; the large fish had the highest level (0.863 ppm). The 2021 bass samples had fillet muscle mercury ranging from 0.169-0.863 ppm, averaging 0.381 ppm. This was very similar to last year (2020: 0.352 ppm) and statistically similar to all of the previous collections here. As compared to baseline/historic samples from Cache Creek, the 2020 Phase 1 adult Largemouth Bass were not elevated in mercury; they were lower than 6 of 7 similar baseline/historic sample sets from Cache Creek; the difference was statistically significant for three of the comparisons. As noted in previous reports, the Phase 1 bass were among the lower mercury top predator fish samples we have collected in California, across many studies. Although the overall concentrations remained relatively low, the changes seen between 2015 and 2021 provide evidence of some of the factors influencing fish mercury exposure in the aggregate mining ponds. The changes in bass mercury uptake corresponded to changes in mining practices at this site: from active mining plus slurry inputs, to slurry only, to no mining or slurry and, after 2018, back to slurry inputs.

Small, Young Fish

Juvenile Largemouth Bass (Tables/Figures 3 and 4)

Juvenile bass were scarce here in 2021. We collected 9 individual fish. These were divided into four size-class composite samples of 2-3 fish each. These whole-body composites had uniformly low mercury levels of 0.085-0.111 ppm, with a mean of 0.096 ppm. Levels have been consistently low at this pond, across the seven years monitored to this point. Within this range of relatively low juvenile bass mercury concentrations, the 2021 set were in the mid-range of the long-term data. They were statistically lower in 2021 (0.096 ppm) than 2017 (0.146 ppm) and 2019 (0.114 ppm), statistically similar to collections from 2016 (0.094 ppm) and 2020 (0.104

ppm), and statistically higher than samples from 2015 (0.044 ppm) and 2018 (0.068 ppm, single fish). Within each year, variability was very low, allowing statistical differentiations between years and with comparative creek samples. <u>Relative to baseline Cache Creek comparison fish, the 2021 Phase 1 juvenile Largemouth Bass were, on average, not elevated</u>; they were significantly lower in mercury than the River Mile 28 set and significantly above the River Mile 15 set.

Mosquitofish (Tables/Figures 3 and 5)

Mosquitofish were sampled with four ascending size-class sets of 8-10 individuals each. These multiple-fish composites had whole-body mercury ranging from 0.077-0.153 ppm, averaging 0.115 ppm. The seven-year trend of mean levels showed a gradual increase between 2015 and 2017 (0.075-0.135 ppm), a significant decline in 2018 (0.083 ppm), and a gradual increase since then (0.096-0.115 ppm). The 2021 set were significantly higher in mercury than the 2015 and 2018 fish, and statistically similar to the samples from all the other years. As in previous years and at the other sites, this species was more variable within each year than the juvenile bass, showing an increase in mercury with Mosquitofish size. This broadened statistical confidence intervals of the means, leading to more overlap statistically. <u>Relative to the creek baseline comparisons, the 2021 Phase 1 Mosquitofish were not elevated;</u> mean mercury was statistically similar to both River Mile 15 comparisons (0.100-0.103 ppm) and significantly lower than the River Mile 17 sets (0.178 ppm).

Juvenile Green Sunfish (Tables/Figures 3 and 6)

The juvenile Green Sunfish composites had whole-body mercury ranging from 0.077-0.111 ppm, averaging 0.090 ppm, nearly identical to the last two years (0.089 ppm). This species, collected since 2017, was generally consistent with the other two small fish species: highest levels were seen in 2017 (0.118 ppm), lowest in 2018 (0.035 ppm), and a relative increase to an intermediate level in 2019-2021 (0.089-0.090 ppm). These broad changes were statistically significant, though all were relatively low levels. <u>As compared to the creek baseline samples, the 2021 Phase 1 juvenile Green Sunfish were not elevated</u>; levels were statistically similar to two of five comparisons and significantly lower than three.

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Fish Species	Fish Tot (mm)	al Length (inches)	Fish ' (g)	Weight (lbs)	Muscle Mercury $(\mu g/g = ppm, wet wt)$
Largemouth Bass	224	8.8	150	0.3	0.169
Largemouth Bass	228	9.0	165	0.4	0.181
Largemouth Bass	237	9.3	160	0.4	0.244
Largemouth Bass	254	10.0	168	0.4	0.235
Largemouth Bass	255	10.0	202	0.4	0.294
Largemouth Bass	257	10.1	215	0.5	0.282
Largemouth Bass	260	10.2	205	0.5	0.204
Largemouth Bass	265	10.4	255	0.6	0.246
Largemouth Bass	268	10.6	262	0.6	0.24 8
Largemouth Bass	272	10.7	230	0.5	0.525
Largemouth Bass	274	10.8	225	0.5	0.526
Largemouth Bass	276	10.9	225	0.5	0.508
Largemouth Bass	277	10.9	270	0.6	0.379
Largemouth Bass	283	11.1	280	0.6	0.469
Largemouth Bass	284	11.2	288	0.6	0.418
Largemouth Bass	284	11.2	282	0.6	0.504
Largemouth Bass	293	11.5	302	0.7	0.430
Largemouth Bass	297	11.7	360	0.8	0.376
Largemouth Bass	302	11.9	342	0.8	0.513
Largemouth Bass	393	15.5	655	1.4	0.863

Table 1. Cemex – Phase 1 Pond: Individual large fish sampled, 2021

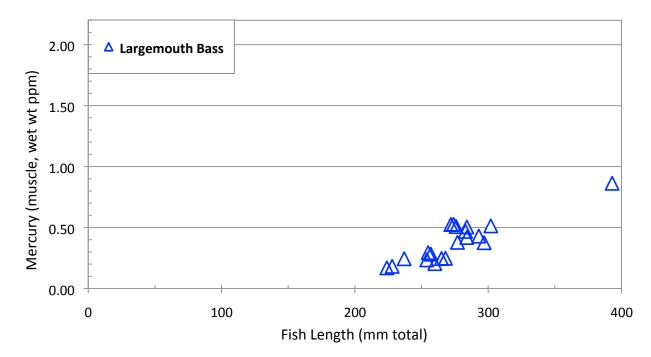
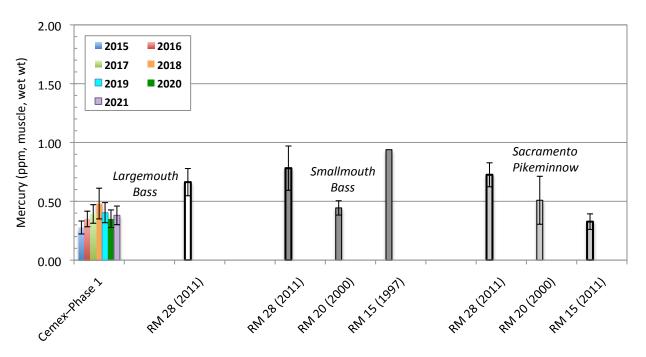


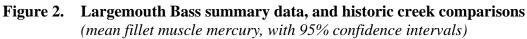
Figure 1.Cemex – Phase 1 Pond: Large fish sampled, 2021
(fillet muscle mercury in individual fish)

Table 2. Largemouth Bass summary data, and historic creek comparisons

(mean fillet muscle mercury, with 95% confidence intervals)

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (μg/g ppm, wet w	
Cemex – Phase 1	2015	18	305	393	0.278	± 0.055
Cemex – Phase 1	2016	20	313	383	0.350	± 0.066
Cemex – Phase 1	2017	17	299	357	0.393	± 0.079
Cemex – Phase 1	2018	20	298	331	0.481	± 0.131
Cemex – Phase 1	2019	20	280	247	0.404	± 0.085
Cemex – Phase 1	2020	20	267	253	0.352	± 0.075
Cemex – Phase 1	2021	20	274	262	0.381	± 0.079
Largemouth Bass River Mile 28	2011	0				
	2011	9	199	137	0.663	± 0.116
Smallmouth Bass	2011	9	199	137	0.663	± 0.116
Smallmouth Bass River Mile 28	2011	9 7	199 265	137 326	0.663 0.782	± 0.116 ± 0.188
		-				
River Mile 28	2011	7	265	326	0.782	± 0.188
River Mile 28 River Mile 20	2011 2000 1997	7 7	265 234	326 183	0.782 0.444	± 0.188
River Mile 28 River Mile 20 River Mile 15	2011 2000 1997	7 7	265 234	326 183	0.782 0.444	± 0.188
River Mile 28 River Mile 20 River Mile 15 Sacramento Pikeminno	2011 2000 1997	7 7 2	265 234 383	326 183 780	0.782 0.444 0.939	± 0.188 ± 0.061





Small, Young Fish Samples (note lower concentration scales)

Table 3. Cemex – Phase 1 Pond: Small Fish Sampled, 2021

(multi-individual, whole body composite samples) 'n' = number: number of individual fish per composite

Fish Species	n (indivs. in comp)	Av. Fish (mm)	Length (inches)	Av. Fisl (g)	h Weight (oz)	Whole-Body Mercury $(\mu g/g = ppm, wet wt)$
Largemouth Bass (juv)	2	71	2.8	4.90	0.17	0.088
Largemouth Bass (juv)	3	76	3.0	5.95	0.21	0.085
Largemouth Bass (juv)	2	84	3.3	7.68	0.27	0.111
Largemouth Bass (juv)	2	85	3.3	8.72	0.31	0.101
Green Sunfish (juv)	8	28	1.1	0.23	0.008	0.077
Green Sunfish (juv)	9	32	1.2	0.84	0.029	0.081
Green Sunfish (juv)	1	40	1.6	5.95	0.210	0.111
Mosquitofish	10	27	1.1	0.19	0.007	0.077
Mosquitofish	10	32	1.3	0.35	0.012	0.094
Mosquitofish	10	37	1.5	0.54	0.019	0.137
Mosquitofish	8	41	1.6	0.86	0.031	0.153

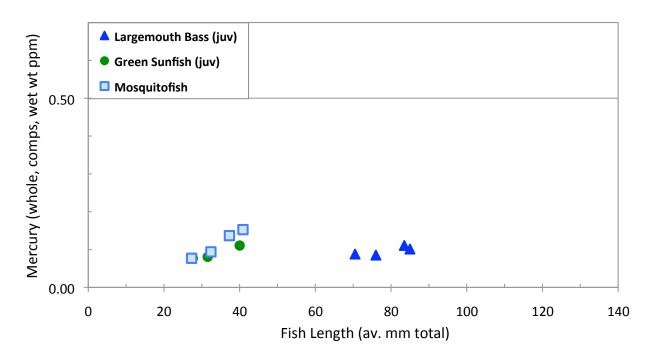


Figure 3. Cemex – Phase 1 Pond: Small, young fish sampled, 2021 (mercury in whole-body, multi-individual composite samples)

Table 4.Juvenile Largemouth Bass summary data, and historic creek comparisons
(means of multiple whole-body, multi-individual composite samples)
'n' = number: number of composite samples; number of individual fish per composite

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	$\mathbf{Hg} \ (\mu g/g = ppm, wet wt)$	Std. Error
Largemouth Bass	(juveniles)						
Cemex – Phase 1	2015	4	8	109	17	0.044	± 0.004
Cemex – Phase 1	2016	4	3	102	17	0.094	± 0.006
Cemex – Phase 1	2017	4	2	117	22	0.146	± 0.011
Cemex – Phase 1	2018	1	1	78	6	0.068	
Cemex – Phase 1	2019	4	4-5	106	17	0.114	± 0.007
Cemex – Phase 1	2020	5	2-4	100	13	0.104	± 0.008
Cemex – Phase 1	2021	4	2-3	79	7	0.096	± 0.006
Historic/Baseline D	ata						
River Mile 28	2011	4	3-5	75	6	0.142	± 0.013
River Mile 15	2011	3	1	93	10	0.050	± 0.014

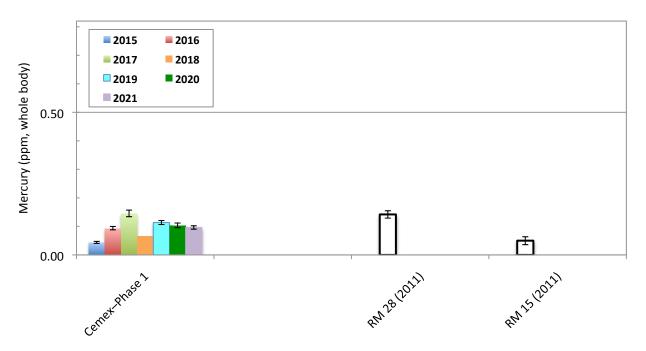


Figure 4. Juvenile Largemouth Bass summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 5. Mosquitofish summary data, and historic creek comparisons

(means of multiple whole-body, multi-individual composite samples) 'n' = number: number of composite samples; number of individual fish per composite

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Mosquitofish							
Cemex – Phase 1	2015	4	10	39	0.6	0.075	± 0.008
Cemex – Phase 1	2016	4	10	34	0.4	0.093	± 0.019
Cemex – Phase 1	2017	4	10	33	0.4	0.135	± 0.019
Cemex – Phase 1	2018	4	6-10	34	0.5	0.083	± 0.016
Cemex – Phase 1	2019	4	10	34	0.5	0.096	± 0.024
Cemex – Phase 1	2020	4	12	35	0.5	0.102	± 0.021
Cemex – Phase 1	2021	4	8-10	34	0.5	0.115	± 0.018
Historic/Baseline De	ata						
River Mile 17	2000-2002	13	5-30	26-47	0.2-1.1	0.178	± 0.020
River Mile 15	2000-2002	10	5-30	26-47	0.2-1.0	0.100	± 0.018
River Mile 15	2000 2002	4	1-10	37	0.7	0.103	± 0.024

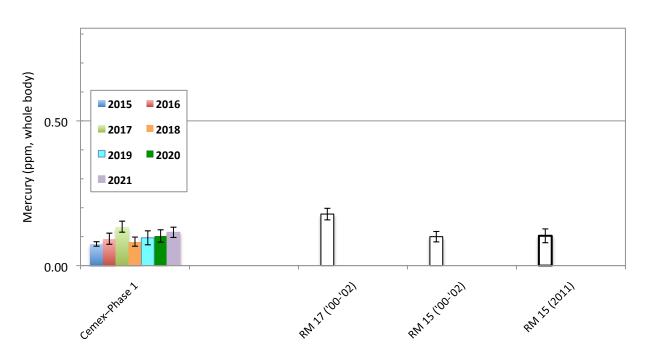


Figure 5. Mosquitofish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 6. Juvenile Green Sunfish summary data, and historic creek comparisons

(means of multiple whole-body, multi-individual composite samples) 'n' = number: number of composite samples; number of individual fish per composite

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Green Sunfish (ju	veniles)						
Cemex – Phase 1	2017	4	8-10	47	1.9	0.118	± 0.023
Cemex – Phase 1	2018	4	2	51	2.1	0.035	± 0.009
Cemex – Phase 1	2019	4	2-10	44	1.7	0.089	± 0.011
Cemex – Phase 1	2020	3	1-3	50	2.7	0.089	± 0.009
Cemex – Phase 1	2021	3	1-9	33	2.3	0.090	± 0.011
Historic/Baseline L	Data						
River Mile 28	2011	4	4	53	2.8	0.139	± 0.007
River Mile 20	2011	4	4	58	3.4	0.084	± 0.002
River Mile 17	2000-2002	8	5-10	41-90	1-6	0.169	± 0.013
River Mile 15	2000-2002	8	4-8	40-87	1-6	0.117	± 0.005
River Mile 15	2011	4	4-5	56	3.1	0.086	± 0.009

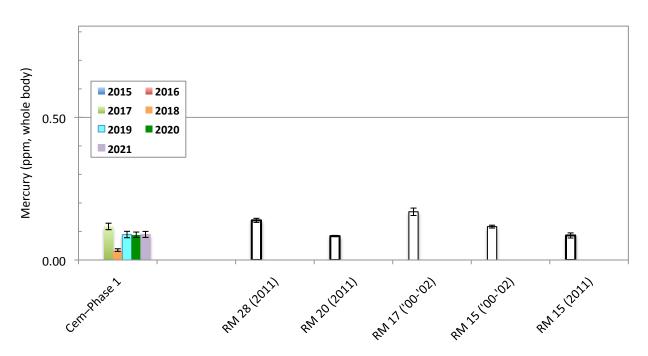


Figure 6. Juv. Green Sunfish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

2. CEMEX – PHASE 3 POND

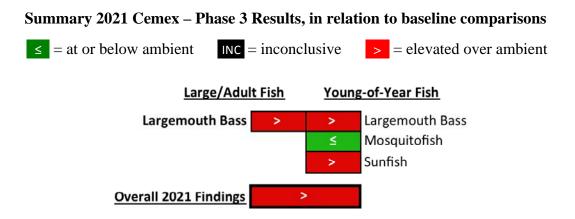


(Google Earth 10/21/2020)

2. CEMEX – PHASE 3 POND (Tables and Figures 7-12)

Summary

The former Phase 3-4 Pond was divided into two separate parts in late 2020, with active mining continuing in the eastern, Phase 4 Pond and the western, Phase 3 sub-basin becoming an isolated, relatively undisturbed pond. Mercury monitoring continued this year in both the Phase 3 and Phase 4 Ponds. In 2021, twenty adult Largemouth Bass were sampled from Phase 3, and multiple composite samples were taken of young-of-year Mosquitofish, juvenile Green Sunfish, and juvenile Largemouth Bass. Except for the Mosquitofish samples, all of the 2021 Phase 3 fish were elevated in mercury on average. As the former combined Phase 3-4 Pond was found to be "elevated for three or more years of five" over creek baselines (2015-2019, all 5 years), that triggered the addition of "expanded analysis" and, following a period of data gathering. development of a mercury management plan. Expanded analysis work began in the Phase 3-4 Pond in 2018, with seasonal water column profiling of a range of relevant constituents and testing of bottom sediments, and is presented in accompanying reports. After Phase 3-4 was split into two parts, this work continued in both ponds in 2021.



The former Phase 3-4 Pond was divided into two separate parts in late 2020, with active mining continuing in the eastern, Phase 4 Pond and the western, Phase 3 sub-basin becoming an isolated, relatively undisturbed pond. Mercury monitoring continued this year in both the Phase 3 and Phase 4 Ponds. Like the other Cemex ponds, Phase 3 is located just south of Cache Creek and

east of Highway 505. It lies between the Phase 1 Pond to the west and the newly separated Phase 4 Pond to the east. The Phase 3 Pond is a rectangular water body that is approximately 700 m long (0.7 km) and 450 m wide. Maximum depth ranged narrowly in 2021, despite drought conditions (8.7-10.1 m, 29-33 feet). This pond was left dormant in 2021, with regard to mining and other earth moving. Riparian and aquatic vegetation rapidly colonized, creating new habitat. Continuing from testing in the previously combined Phase 3-4 Pond, this (2021) was Year 7 of monitoring.

We sampled the pond during day and twilight conditions with a range of techniques, and collected useful samples of most of the fish species present. These included individual fillet muscle samples of 20 Largemouth Bass (*Micropterus salmoides*) across the range of sizes present. The small fish available were juvenile Green Sunfish (*Lepomis cyanellus*, 1-2"), and Mosquitofish (*Gambusia affinis*, 1-2"), each sampled with 3-4 composites, for 10 total composites.

In total, 20 large fish muscle samples and 10 small fish composite samples, 30 separate mercury samples, were analyzed from the Cemex–Phase 3 Pond in the Fall 2021 monitoring. The fish metrics and analytical results from each individual large bass muscle sample can be seen in Table 7 and, graphically, in Figure 7. Then, the new data are shown in reduced form (means, error bars, etc) and compared to 2015-2020 results and the most closely comparable historic creek data (Table 8 and Figure 8). Results from the composite samples of small, young-of-year fish are similarly presented in Tables and Figures 9-12.

Large, Angling-sized Fish

Largemouth Bass (Tables/Figures 7 and 8)

The 2021 bass samples ranged fairly narrowly in size between 330 and 387 mm (about 13-15"); this pond contains a healthy population of mainly larger bass than in the other ponds. The 2021 samples had fillet muscle mercury ranging from 0.539-1.280 ppm, averaging 0.971 ppm. This was up significantly from 2020 (0.656 ppm) in the combined Phase 3-4 Pond and was higher than all but the highest average found there (2017: 1.093 ppm), which it was statistically similar to. In relation to comparable baseline Cache Creek samples, the 2021 Phase 3 bass were elevated in

mercury; the average level was higher than all 7 historical comparisons. The difference was statistically significant for 5 of 7. The 2021 Phase 3 bass average (0.971 ppm) was also significantly higher than the corresponding data from the recently separated, actively mined Phase 4 (0.782 ppm). The bass mercury results from both ponds are shown together in Figure 7(b). Part of the reason for the difference in average mercury levels is that the fish available for collection in the Phase 3 Pond were larger/older on average, all 330-387 mm and 445-740 g in size. In Phase 4, the bass averaged smaller/younger, with about half the fish in a similar size range as in Phase 3 (332-358 mm, 365-550 g), but the rest of the fish in the smaller range of 275-330 mm and 265-430 g. The Phase 3 part of the former Phase 3-4 combined pond was left relatively undisturbed over the last few years, with active mining shifting to the Phase 4 section. Aquatic and riparian vegetation established there, creating steadily improving fish habitat. The catch results indicate that this prime habitat may have been taken over by dominant, largest bass, with many of the smaller bass excluded into the less desirable, active mining habitat of Phase 4. When the size versus mercury data from both ponds are plotted together in Figure 7(b), they blend evenly into each other.

Small, Young Fish

Juvenile Largemouth Bass (Tables/Figures 9 and 10)

As in last few years, juvenile bass were very scarce. There is tremendous predation pressure on them here from the thriving adult bass population. Three individuals were taken though, as compared to just 0-1 in each of the previous three years, allowing some statistical comparison. The three 106-115 mm (4-5") fish had whole body mercury of 0.275-0.329 ppm, averaging 0.302 ppm, in the historic upper range for this site. This was higher than last year's single individual (0.144 ppm) and statistically higher than the 2017 samples (0.249 ppm). It was lower than the 2015, 2016, and 2019 samples – statistically lower than the 2016 set (0.372 ppm). <u>Relative to creek baseline comparisons, the 2021 Phase 3 juvenile Largemouth Bass were elevated in mercury</u>; except for the single fish sample of last year (0.144 ppm), all of the juvenile bass data from this site have been far above the baseline creek samples (0.050-0.142 ppm). The 2021 Phase 3 samples were consistent with that trend.

Juvenile Green Sunfish (Tables/Figures 9 and 11)

Like the juvenile bass, young sunfish were scarce, but we were able to collect 12 individuals, which were split into 3 size-class composites of 2-5 fish each. The samples had whole-body mercury ranging from 0.118-0.214 ppm and averaging 0.159 ppm. This was up significantly from the combined Phase 3-4 Pond in 2020 (0.117 ppm), to a level about mid-range for these samples there since 2017. It remained significantly lower than concentrations in the first monitoring years of 2015-2016 (0.233-0.275 ppm). <u>Compared to baseline juvenile Green Sunfish mercury from Cache Creek, Phase 3 Pond fish in 2021 were elevated on average; they were statistically higher than three creek comparisons and statistically similar to two.</u>

Mosquitofish (Tables/Figures 9 and 12)

We were able to collect strong composite samples of Mosquitofish in 2021, as in all previous years in the combined Phase 3-4 Pond. Four size-class composites of 10 fish each were analyzed. The 2021 samples had whole-body mercury ranging from 0.105-0.201 ppm, averaging 0.137 ppm. This was up from the previous year in the combined pond (2020: 0.112 ppm), though the increase was not statistically significant. Levels remained lower than in all other monitoring years before 2020 (0.157-0.286 ppm), significantly lower than three of the five. <u>Relative to the baseline Cache Creek comparison samples, the 2021 Phase 3 Mosquitofish were, on average, not elevated;</u> statistically lower in mercury than one of three comparisons, similar to one, and higher than one.

Fish	Fish Tot	al Length	Fish V	Weight	Muscle Mercury
Species	(mm)	(inches)	(g)	(lbs)	$(\mu g/g = ppm, wet wt)$
Largemouth Bass	330	13.0	480	1.1	1.006
Largemouth Bass	337	13.3	445	1.0	0.938
Largemouth Bass	339	13.3	510	1.1	0.973
Largemouth Bass	340	13.4	505	1.1	1.168
Largemouth Bass	344	13.5	510	1.1	0.559
Largemouth Bass	345	13.6	545	1.2	0.588
Largemouth Bass	350	13.8	510	1.1	0.880
Largemouth Bass	360	14.2	620	1.4	0.539
Largemouth Bass	360	14.2	625	1.4	0.996
Largemouth Bass	362	14.3	555	1.2	1.016
Largemouth Bass	363	14.3	555	1.2	1.049
Largemouth Bass	364	14.3	605	1.3	1.099
Largemouth Bass	366	14.4	470	1.0	1.280
Largemouth Bass	366	14.4	595	1.3	1.073
Largemouth Bass	367	14.4	610	1.3	1.010
Largemouth Bass	370	14.6	650	1.4	1.089
Largemouth Bass	371	14.6	635	1.4	1.133
Largemouth Bass	372	14.6	640	1.4	1.026
Largemouth Bass	375	14.8	740	1.6	0.950
Largemouth Bass	387	15.2	695	1.5	1.055

Table 7. Cemex – Phase 3 Pond: Individual large fish sampled, 2021

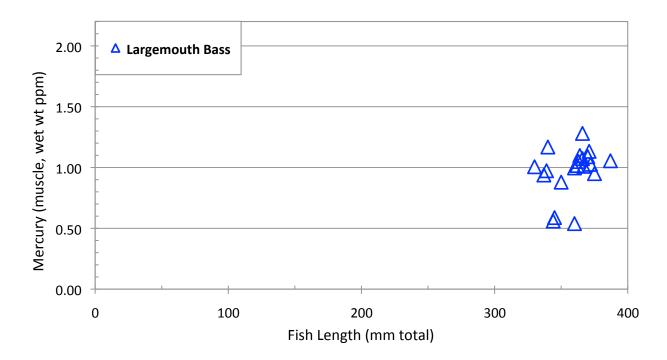


Figure 7. Cemex – Phase 3 Pond: Large fish sampled, 2021 (*fillet muscle mercury in individual fish*)

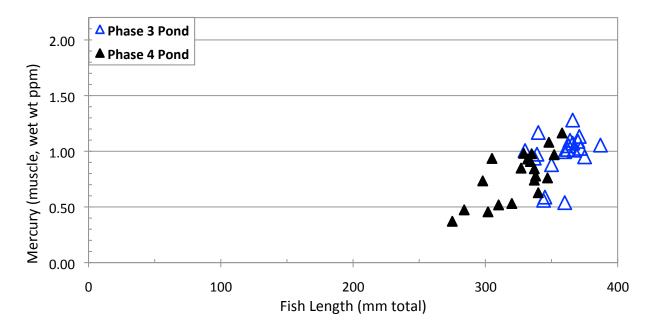
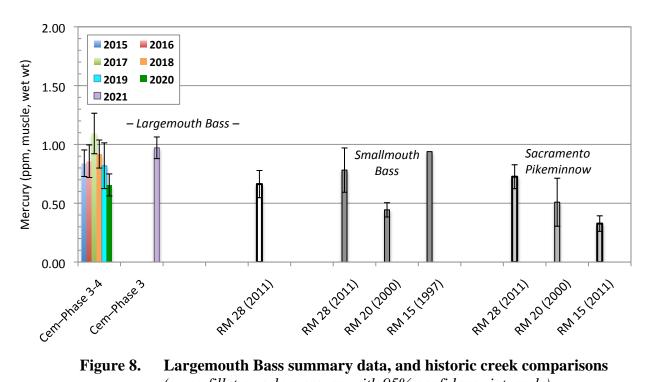
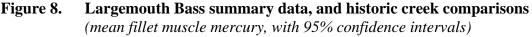


Figure 7(b). Cemex – Phase 3 AND Phase 4 Bass comparison, 2021 (*fillet muscle mercury in individual fish*)

Table 8. Largemouth Bass summary data, and historic creek comparisons

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (μg/g ppm, wet wt	
Cemex – Phase 3-4	2015	20	344	526	0.840	± 0.113
Cemex – Phase 3-4	2016	20	344	557	0.858	± 0.139
Cemex – Phase 3-4	2017	20	334	479	1.093	± 0.172
Cemex – Phase 3-4	2018	20	331	463	0.918	± 0.119
Cemex – Phase 3-4	2019	20	312	402	0.819	± 0.195
Cemex – Phase 3-4	2020	20	310	399	0.656	± 0.094
Cemex – Phase 3	2021	20	358	575	0.971	± 0.092
	ta (comparal	ble predatory	species)			
Historic/Baseline Da Largemouth Bass River Mile 28	ta (comparal 2011	ble predatory 9	, species) 199	137	0.663	± 0.116
<i>Largemouth Bass</i> River Mile 28		1 2	1 /	137	0.663	± 0.116
Largemouth Bass River Mile 28 Smallmouth Bass		1 2	1 /	137 326	0.663 0.782	± 0.116 ± 0.188
Largemouth Bass	2011	9	199			
Largemouth Bass River Mile 28 Smallmouth Bass River Mile 28	2011 2011	9 7	199 265	326	0.782	± 0.188
Largemouth Bass River Mile 28 Smallmouth Bass River Mile 28 River Mile 20	2011 2011 2000 1997	9 7 7	199 265 234	326 183	0.782 0.444	± 0.188
Largemouth Bass River Mile 28 Smallmouth Bass River Mile 28 River Mile 20 River Mile 15	2011 2011 2000 1997	9 7 7	199 265 234	326 183	0.782 0.444	± 0.188
Largemouth Bass River Mile 28 Smallmouth Bass River Mile 28 River Mile 20 River Mile 15 Sacramento Pikeminnow	2011 2011 2000 1997	9 7 7 2	199 265 234 383	326 183 780	0.782 0.444 0.939	± 0.188 ± 0.061





Small, Young Fish Samples (note lower concentration scales)

Table 9. Cemex – Phase 3 Pond: Small Fish Sampled, 2021

(multi-individual, whole body composite samples) 'n' = number: number of individual fish per composite

Fish Species	n (indivs. in comp)	Av. Fish Len (mm) (inch	8	h Weight (oz)	Whole-Body Mercury $(\mu g/g = ppm, wet wt)$
Largemouth Bass (juv)	1	106 4.	2 10.9	0.38	0.329
Largemouth Bass (juv)	1	108 4.	3 10.8	0.38	0.275
Largemouth Bass (juv)	1	115 4.	5 14.1	0.50	0.301
Green Sunfish (juv)	5	29 1.	1 0.45	0.016	0.118
Green Sunfish (juv)	5	33 1.	3 0.79	0.028	0.146
Green Sunfish (juv)	2	38 1.	5 0.49	0.017	0.214
Mosquitofish	10	28 1.	1 0.21	0.007	0.105
Mosquitofish	10	33 1.	3 0.32	0.011	0.111
Mosquitofish	10	37 1.	5 0.53	0.019	0.130
Mosquitofish	10	42 1.	7 0.77	0.027	0.201

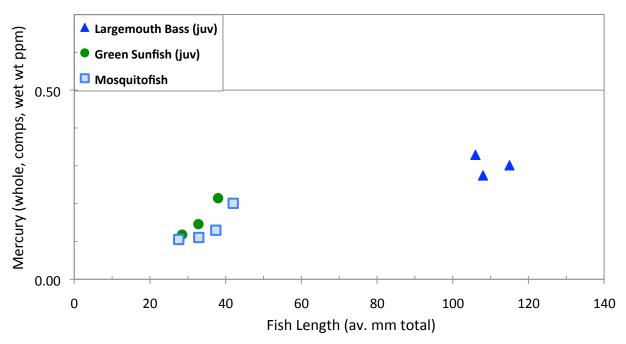


Figure 9. Cemex – Phase 3 Pond: Small, young fish sampled, 2021 (mercury in whole-body, multi-individual composite samples)

Table 10. Juvenile Largemouth Bass summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Largemouth Bass (j	juveniles)						
Cemex – Phase 3-4	2015	4	7	108	16	0.334	± 0.052
Cemex – Phase 3-4	2016	4	2	114	18	0.372	± 0.053
Cemex – Phase 3-4	2017	4	2-3	108	16	0.249	± 0.033
Cemex – Phase 3-4	2018	(no sample	s)				
Cemex – Phase 3-4	2019	1	1	125	23	0.336	
Cemex – Phase 3-4	2020	1	1	124	23	0.144	
Cemex – <u>Phase 3</u>	2021	3	1	110	12	0.302	± 0.016
Historic/Baseline Da	ita						
River Mile 28	2011	4	3-5	75	6	0.142	± 0.013
River Mile 15	2011	3	1	93	10	0.050	± 0.014
		·	-				

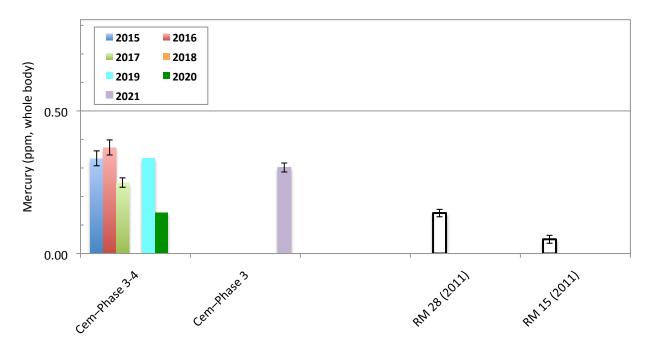


Figure 10. Juvenile Largemouth Bass summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 11. Juvenile Green Sunfish summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Green Sunfish (juv	eniles)						
Cemex – Phase 3-4	2015	4	10	47	1.8	0.275	± 0.022
Cemex – Phase 3-4	2016	4	4-5	49	2.0	0.233	± 0.026
Cemex – Phase 3-4	2017	4	2-6	36	0.7	0.150	± 0.051
Cemex – Phase 3-4	2018	4	1	34	0.5	0.112	± 0.020
Cemex – Phase 3-4	2019	4	10	43	1.6	0.185	± 0.016
Cemex – Phase 3-4	2020	4	1-12	38	0.9	0.117	± 0.024
Cemex – <u>Phase 3</u>	2021	3	2-5	33	0.6	0.159	± 0.029
Historic/Baseline D	ata						
River Mile 28	2011	4	4	53	2.8	0.139	± 0.007
River Mile 20	2011	4	4	58	3.4	0.084	± 0.002
River Mile 17	2000-2002	8	5-10	41-90	1-6	0.169	± 0.013
River Mile 15	2000-2002	8	4-8	40-87	1-6	0.117	± 0.005
River Mile 15	2011	4	4-5	56	3.1	0.086	± 0.009

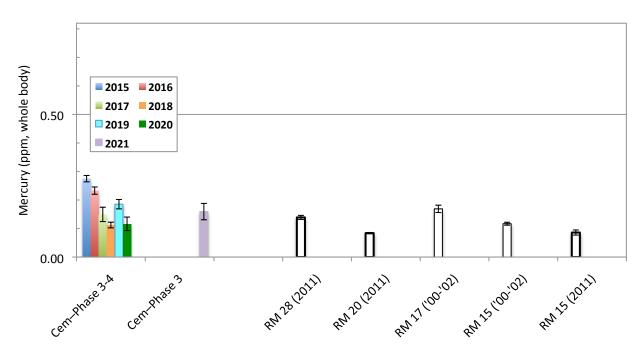


Figure 11. Juv. Green Sunfish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 12. Mosquitofish summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Mosquitofish							
Cemex – Phase 3-4	2015	4	10	37	0.6	0.228	± 0.029
Cemex – Phase 3-4	2016	4	10	37	0.6	0.157	± 0.019
Cemex – Phase 3-4	2017	4	6-10	34	0.5	0.286	± 0.035
Cemex – Phase 3-4	2018	4	3-10	34	0.5	0.203	± 0.021
Cemex – Phase 3-4	2019	4	10	35	0.6	0.183	± 0.029
Cemex – Phase 3-4	2020	4	3-12	33	0.4	0.112	± 0.018
Cemex – <u>Phase 3</u>	2021	4	10	35	0.5	0.137	± 0.022
Historic/Baseline Do	ita						
River Mile 17	2000-2002	13	5-30	26-47	0.2-1.1	0.178	± 0.020
River Mile 15	2000-2002	10	5-30	26-47	0.2-1.0	0.100	± 0.018
River Mile 15	2011	4	1-10	37	0.7	0.103	± 0.024

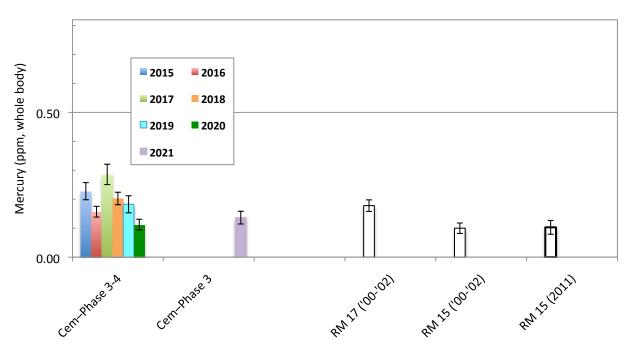


Figure 12. Mosquitofish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

3. CEMEX – PHASE 4 POND

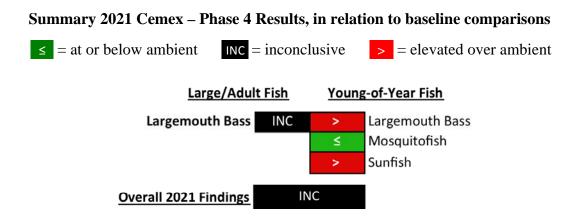


(Google Earth 10/21/2020)

3. CEMEX – PHASE 4 POND (Tables and Figures 13-18)

Summary

The former Phase 3-4 Pond was divided into two separate parts in late 2020, with active mining continuing in the eastern, Phase 4 Pond and the western, Phase 3 Pond left dormant. Mercury monitoring continued this year in both the Phase 3 and Phase 4 Ponds. In 2021, twenty adult Largemouth Bass were sampled from Phase 4, and multiple composite samples were taken of young-of-year Mosquitofish, juvenile Green Sunfish, and juvenile Largemouth Bass. <u>Overall mercury results were 'inconclusive' in 2021</u>, with some samples relatively elevated and others, including the key adult bass, similar to or below baseline comparisons. However, as the former combined Phase 3-4 Pond was found to be "elevated for three or more years of five" over creek baselines (2015-2019, all 5 years), that triggered the addition of "expanded analysis" and, following a period of data gathering, development of a mercury management plan. Expanded analysis work began in the Phase 3-4 Pond in 2018, with seasonal water column profiling of a range of relevant constituents and testing of bottom sediments, and is presented in accompanying reports. After Phase 3-4 was split into two parts, this work continued in both ponds in 2021. Future status of Phase 4 will depend on cumulative monitoring results.



The former Phase 3-4 Pond was divided into two separate parts in late 2020, with active mining continuing in the eastern, Phase 4 Pond and the western, Phase 3 sub-basin becoming an isolated, relatively undisturbed pond. Mercury monitoring continued this year in both the Phase 3 and

Phase 4 Ponds. Like the other Cemex ponds, Phase 4 is located just south of Cache Creek and east of Highway 505. It lies east of the newly separated Phase 3 Pond, which is east of the Phase 1 Pond. The Phase 4 Pond is app. 850 m long (0.85 km) and 450 m wide. Maximum depth ranged narrowly in 2021 (8.2-9.4 m, 27-31 feet), despite drought conditions. This pond was mined actively throughout 2021. Continuing from testing in the previously combined Phase 3-4 Pond, this (2021) was Year 7 of monitoring.

We sampled the pond during day and twilight conditions with a range of techniques, and collected useful samples of most of the fish species present. These included individual fillet muscle samples of 20 Largemouth Bass (*Micropterus salmoides*) across the range of sizes present. The small fish available were juvenile Largemouth Bass (3-5"), juvenile Green Sunfish (*Lepomis cyanellus*, 1-2"), and Mosquitofish (*Gambusia affinis*, 1-2"), each sampled with 4 composites, for 12 total composite samples.

In total, 20 large fish muscle samples and 12 small fish composite samples, 32 separate mercury samples, were analyzed from the Cemex–Phase 4 Pond in the Fall 2021 monitoring. The fish metrics and analytical results from each individual large bass muscle sample can be seen in Table 13 and, graphically, in Figure 13. Then, the new data are shown in reduced form (means, error bars, etc) and compared to 2015-2020 results and the most closely comparable historic creek data (Table 14 and Figure 148). Results from the composite samples of small, young-of-year fish are similarly presented in Tables and Figures 15-18.

Large, Angling-sized Fish

Largemouth Bass (Tables/Figures 13 and 14)

The Phase 4 Pond adult Largemouth Bass samples had fillet muscle mercury ranging from 0.371-1.164 ppm, averaging 0.782 ppm. This was up (not significantly) from 2020 (0.656 ppm) in the formerly combined Phase 3-4 Pond, but remained lower than all previous years there (2015-2019: 0.819-1.093 ppm); the difference was statistically significant for the highest comparison (2018: 1.093 ppm). The Phase 4 average bass mercury (0.782 ppm) was significantly lower than the average in the recently separated Phase 3 (0.971 ppm). As discussed above in the section on the Phase 3 bass, this was at least partly due to the generally smaller/younger fish sizes in the Phase 4 Pond relative to the Phase 3 fish; the largest, most dominant bass appear to have excluded many of the smaller bass from the preferred habitat of Phase 3 and into the less desirable active mining Phase 4 section. With full separation of the two Phases, this left smaller/younger average sizes in Phase 4 in the short term. As shown in the Phase 3 section and again here in Figure 13(b), when the bass size versus mercury data from both ponds are plotted together, they blend fairly evenly into each other. In relation to comparable baseline Cache Creek samples, the 2021 Phase 4 bass were in the borderline, 'inconclusive' range for mercury; the average level was similar to the higher baseline samples; statistically similar to 5 of 7 historical comparisons and significantly higher than 2 of 7.

Small, Young Fish

Juvenile Largemouth Bass (Tables/Figures 15 and 16)

We were able to collect a more extensive set of juvenile bass (12 total) from the recently separated Phase 4 Pond in 2021 than in Phase 3 (3 fish) or from the combined Phase 3-4 Pond throughout 2018-2020 (0-1 fish/yr). The 12 fish were divided into four size-class composite samples of 2-4 fish each, with the sets spanning the 77-120 mm (3-5") size range. The four composites ranged in mercury from 0.135-0.205 ppm, averaging 0.160 ppm. This was significantly lower than in the 2021 Phase 3 Pond (0.302 ppm) and the historic record from the combined Phase 3-4 Pond throughout 2015-2019. In comparison with the Cache Creek baseline samples, though, the 2021 Phase 4 juvenile Largemouth Bass were, on average, still elevated; the elevation was significant for one of the comparisons and statistically overlapping the other.

Juvenile Green Sunfish (Tables/Figures 15 and 17)

Twenty-three individuals were taken, which were split into 4 size-class composites of 3-10 fish each. The samples had whole-body mercury ranging from 0.109-0.203 ppm and averaging 0.160 ppm. This was up significantly from the combined Phase 3-4 Pond in 2020 (0.117 ppm), to a level about mid-range for these samples there since 2017. It remained significantly lower than concentrations in the first monitoring years of 2015-2016 (0.233-0.275 ppm). In contrast with the

juvenile bass in the newly separated ponds, the juvenile Green Sunfish samples remained nearly identical in both sides in 2021. <u>Compared to baseline samples from Cache Creek, Phase 4 Pond</u> <u>juvenile sunfish in 2021 were elevated on average</u>; they were significantly higher in mercury than three creek comparisons and statistically similar to two.

Mosquitofish (Tables/Figures 15 and 18)

Four size-class composites of 6-10 fish each were analyzed. The 2021 samples had whole-body mercury ranging from 0.086-0.157 ppm, averaging 0.122 ppm. This was up somewhat from the previous year in the combined pond (2020: 0.112 ppm); the increase was not statistically significant. Levels were statistically lower than in all other monitoring years before 2020 (0.157-0.286 ppm). They were also lower than in comparable samples from the 2021 Phase 3 Pond (0.137 ppm), though not significantly. <u>Relative to the baseline Cache Creek comparison samples, the 2021 Phase 4 Mosquitofish were not elevated</u>; they were statistically lower in mercury than the River Mile 17 sample sets (0.178 ppm) and statistically overlapping (similar to) the two sets from River Mile 15 (0.100-0.103 ppm).

Fish	Fish Tot	al Length	Fish V	Weight	Muscle Mercury
Species	(mm)	(inches)	(g)	(lbs)	$(\mu g/g = ppm, wet wt)$
Largemouth Bass	275	10.8	265	0.6	0.371
Largemouth Bass	284	11.2	265	0.6	0.474
Largemouth Bass	298	11.7	350	0.8	0.734
Largemouth Bass	302	11.9	380	0.8	0.456
Largemouth Bass	305	12.0	305	0.7	0.936
Largemouth Bass	310	12.2	385	0.8	0.519
Largemouth Bass	320	12.6	390	0.9	0.531
Largemouth Bass	327	12.9	430	0.9	0.849
Largemouth Bass	329	13.0	415	0.9	0.981
Largemouth Bass	332	13.1	365	0.8	0.931
Largemouth Bass	334	13.1	455	1.0	0.908
Largemouth Bass	335	13.2	455	1.0	0.978
Largemouth Bass	337	13.3	525	1.2	0.741
Largemouth Bass	337	13.3	450	1.0	0.843
Largemouth Bass	338	13.3	415	0.9	0.778
Largemouth Bass	340	13.4	470	1.0	0.628
Largemouth Bass	347	13.7	465	1.0	0.761
Largemouth Bass	348	13.7	550	1.2	1.081
Largemouth Bass	352	13.9	505	1.1	0.969
Largemouth Bass	358	14.1	515	1.1	1.164

Table 13. Cemex – Phase 4 Pond: Individual large fish sampled, 2021

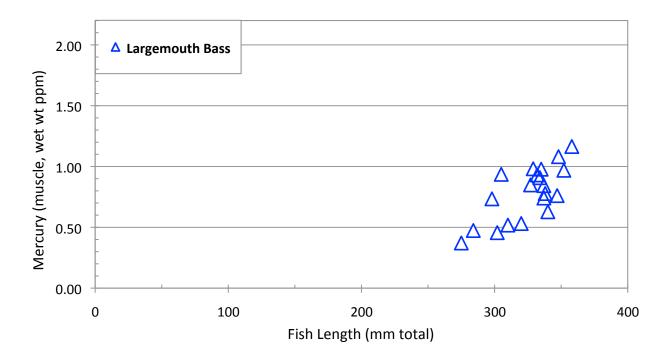


Figure 13. Cemex – Phase 4 Pond: Large fish sampled, 2021 (fillet muscle mercury in individual fish)

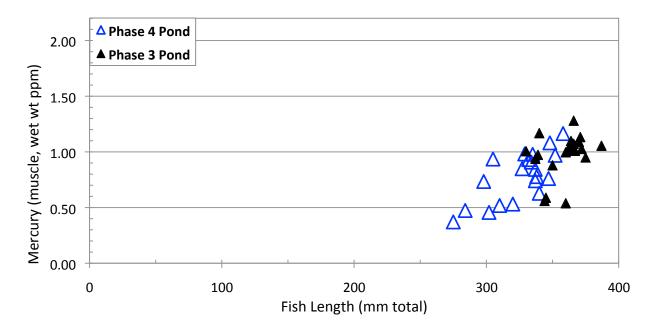
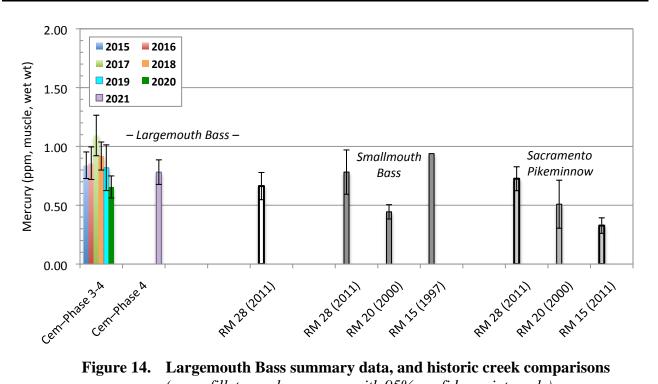
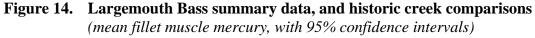


Figure 13(b). Cemex – Phase 4 AND Phase 3 Bass comparison, 2021 (*fillet muscle mercury in individual fish*)

Table 14. Largemouth Bass summary data, and historic creek comparisons

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (μg/g ppm, wet w	
Cemex – Phase 3-4	2015	20	344	526	0.840	± 0.113
Cemex – Phase 3-4	2016	20	344	557	0.858	± 0.139
Cemex – Phase 3-4	2017	20	334	479	1.093	± 0.172
Cemex – Phase 3-4	2018	20	331	463	0.918	± 0.119
Cemex – Phase 3-4	2019	20	312	402	0.819	± 0.195
Cemex – Phase 3-4	2020	20	310	399	0.656	± 0.094
Cemex – Phase 4	2021	20	325	418	0.782	± 0.105
Historic/Baseline Da Largemouth Bass	ta (comparal	ble predatory	species)			
Historic/Baseline Da Largemouth Bass River Mile 28	ta (comparal 2011	ble predatory 9	species) 199	137	0.663	± 0.116
<i>Largemouth Bass</i> River Mile 28		1 2	. ,	137	0.663	± 0.116
Largemouth Bass River Mile 28 Smallmouth Bass	2011	9	199			
<i>Largemouth Bass</i> River Mile 28		1 2	. ,	137 326 183	0.663 0.782 0.444	± 0.188
Largemouth Bass River Mile 28 Smallmouth Bass River Mile 28	2011 2011	9 7	199 265	326	0.782	
Largemouth Bass River Mile 28 Smallmouth Bass River Mile 28 River Mile 20	2011 2011 2000 1997	9 9 7 7	199 265 234	326 183	0.782 0.444	± 0.188
Largemouth Bass River Mile 28 Smallmouth Bass River Mile 28 River Mile 20 River Mile 15	2011 2011 2000 1997	9 9 7 7	199 265 234	326 183	0.782 0.444	± 0.188
Largemouth Bass River Mile 28 Smallmouth Bass River Mile 28 River Mile 20 River Mile 15 Sacramento Pikeminnow	2011 2011 2000 1997	9 7 7 2	199 265 234 383	326 183 780	0.782 0.444 0.939	± 0.188 ± 0.061





Small, Young Fish Samples (note lower concentration scales)

Table 15. Cemex – Phase 4 Pond: Small Fish Sampled, 2021

(multi-individual, whole body composite samples) 'n' = number: number of individual fish per composite

Fish Species	n (indivs. in comp)	Av. Fis (mm)	h Length (inches)	Av. Fisl (g)	h Weight (oz)	Whole-Body Mercury $(\mu g/g = ppm, wet wt)$
Largemouth Bass (juv)	4	77	3.0	6.00	0.21	0.150
Largemouth Bass (juv)	4	85	3.4	7.63	0.27	0.150
Largemouth Bass (juv)	2	95	3.7	10.8	0.38	0.135
Largemouth Bass (juv)	2	120	4.7	21.6	0.76	0.205
Green Sunfish (juv)	10	34	1.3	0.22	0.008	0.152
Green Sunfish (juv)	6	41	1.6	0.49	0.017	0.109
Green Sunfish (juv)	4	48	1.9	0.48	0.017	0.203
Green Sunfish (juv)	3	54	2.1	0.93	0.033	0.176
Mosquitofish	10	28	1.1	0.20	0.007	0.086
Mosquitofish	10	32	1.3	0.37	0.013	0.116
Mosquitofish	9	37	1.5	0.53	0.019	0.130
Mosquitofish	6	42	1.6	0.79	0.028	0.157

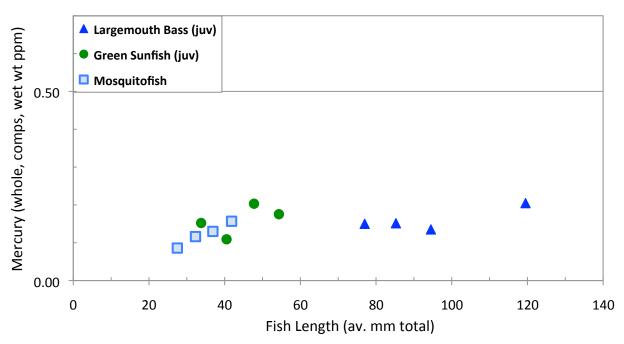


Figure 15. Cemex – Phase 4 Pond: Small, young fish sampled, 2021 (mercury in whole-body, multi-individual composite samples)

Table 16. Juvenile Largemouth Bass summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

 $\mathbf{n}' = number: number of composite samples; number of individual fish per composite$

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Largemouth Bass (j	uveniles)						
Cemex – Phase 3-4	2015	4	7	108	16	0.334	± 0.052
Cemex – Phase 3-4	2016	4	2	114	18	0.372	± 0.053
Cemex – Phase 3-4	2017	4	2-3	108	16	0.249	± 0.033
Cemex – Phase 3-4	2018	(no sample	s)				
Cemex – Phase 3-4	2019	1	1	125	23	0.336	
Cemex – Phase 3-4	2020	1	1	124	23	0.144	
Cemex – <u>Phase 4</u>	2021	4	2-4	94	12	0.160	± 0.015
Historic/Baseline Da	uta						
River Mile 28	2011	4	3-5	75	6	0.142	± 0.013
River Mile 15	2011	3	1	93	10	0.050	± 0.014

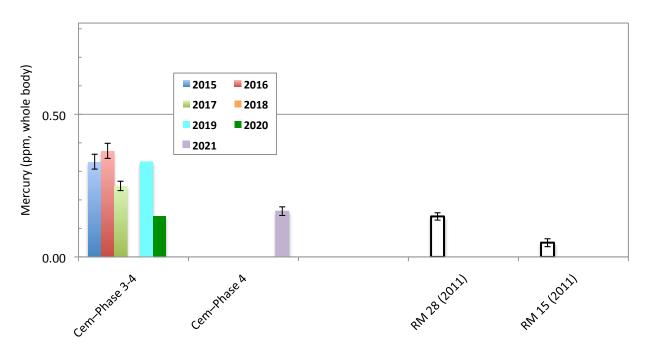


Figure 16. Juvenile Largemouth Bass summary data, and historic creek comparisons *(means of multiple whole-body, multi-individual composite samples)*

Table 17. Juvenile Green Sunfish summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Green Sunfish (juv	veniles)						
Cemex – Phase 3-4	2015	4	10	47	1.8	0.275	± 0.022
Cemex – Phase 3-4	2016	4	4-5	49	2.0	0.233	± 0.026
Cemex – Phase 3-4	2017	4	2-6	36	0.7	0.150	± 0.051
Cemex – Phase 3-4	2018	4	1	34	0.5	0.112	± 0.020
Cemex – Phase 3-4	2019	4	10	43	1.6	0.185	± 0.016
Cemex – Phase 3-4	2020	4	1-12	38	0.9	0.117	± 0.024
Cemex – Phase 4	2021	4	3-10	44	0.5	0.160	± 0.020
Historic/Baseline D	ata						
River Mile 28	2011	4	4	53	2.8	0.139	± 0.007
River Mile 20	2011	4	4	58	3.4	0.084	± 0.002
River Mile 17	2000-2002	8	5-10	41-90	1-6	0.169	± 0.013
River Mile 15	2000-2002	8	4-8	40-87	1-6	0.117	± 0.005
River Mile 15	2011	4	4-5	56	3.1	0.086	± 0.009

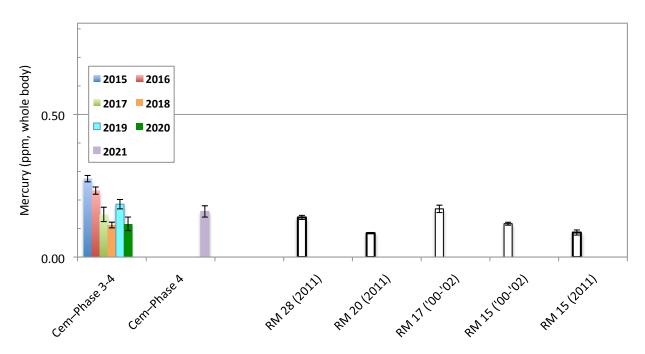


Figure 17. Juv. Green Sunfish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 18. Mosquitofish summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Mosquitofish							
Cemex – Phase 3-4	2015	4	10	37	0.6	0.228	+0.029
Cemex – Phase 3-4	2016	4	10	37	0.6	0.157	± 0.029 ± 0.019
Cemex – Phase 3-4	2017	4	6-10	34	0.5	0.286	± 0.035
Cemex – Phase 3-4	2018	4	3-10	34	0.5	0.203	± 0.021
Cemex – Phase 3-4	2019	4	10	35	0.6	0.183	± 0.029
Cemex – Phase 3-4	2020	4	3-12	33	0.4	0.112	± 0.018
Cemex – <u>Phase 4</u>	2021	4	6-10	35	0.5	0.122	± 0.015
Historic/Baseline Do	ita						
River Mile 17	2000-2002	13	5-30	26-47	0.2-1.1	0.178	± 0.020
River Mile 15	2000-2002	10	5-30	26-47	0.2-1.0	0.100	± 0.018
River Mile 15	2011	4	1-10	37	0.7	0.103	± 0.024

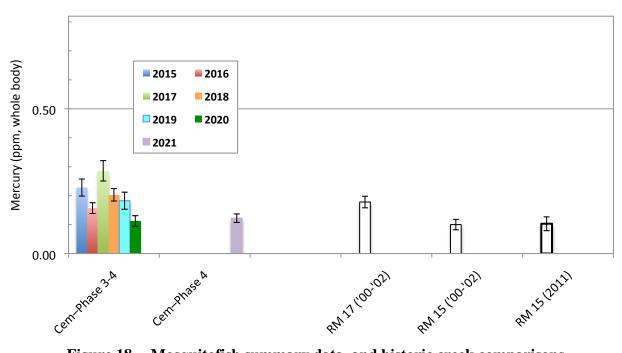


Figure 18. Mosquitofish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

4. TEICHERT – ESPARTO POND

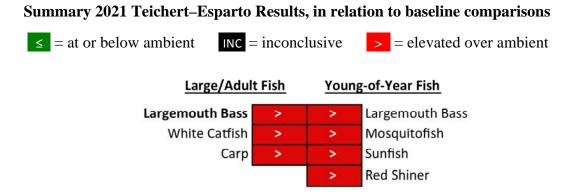


(Google Earth 10/21/2020)

4. TEICHERT – ESPARTO POND (Tables and Figures 19-27)

Summary

The previously separate Reiff and Mast Ponds were combined by Teichert into a single large Esparto Pond in early 2020, by excavating parts of dividing levees. Monitoring continued in the combined pond. With the second year drought conditions of 2021, the Esparto Pond began the sampling season interconnected but dried down into disconnected, shallow basins by summer. Samples of adult Largemouth Bass (20), White Catfish (6), and Common Carp (2) were taken from the central basin. Small, young-of-year fish were sampled from two basins; these included Mosquitofish, juvenile Largemouth Bass, juvenile Green Sunfish, and Red Shiners. Average mercury levels in all sample types showed increases in 2021, particularly in the small, young fish – which are most representative of recent conditions. <u>This site remained highly elevated in mercury in 2021</u>. All of the fish sample types averaged significantly higher mercury than corresponding Cache Creek baseline samples. Similar results from previous years in the Reiff and Mast ponds triggered the collection of additional information ("expanded analysis") to help guide development of a mercury management plan. Water column profiling and collection of bottom sediment samples began in May 2018 and are the subject of accompanying reports.



Overall 2021 Findings

>

The site is located at Teichert's Esparto Facility, just north of Cache Creek and west of Highway 505, between 505 and County Road 87. Mining began here in or before 2002. Active mining has been sporadic over the years, but aggregate processing was mostly continuous; slurry returns have typically kept the pond water very turbid/opaque. The combined Esparto Pond is approximately 1100 m long (1.1 km) and 300-500 m wide; drought reduced it into three smaller sub-basins within that footprint. Maximum depths dropped from 10.4 m (34 ft) in October 2020 to 9.4 m (31 ft) in May 2021 to 4.3 m (14 ft) in October 2021. Continuing from previous testing in Reiff and Mast Ponds, this (2021) was Year 7 of monitoring.

We sampled the pond(s) as possible, using a range of gear over multiple days. The fish collected included, for large, angling-sized fish, samples of 20 Largemouth Bass (*Micropterus salmoides*), 6 White Catfish (*Ameiurus catus*), and 2 Common Carp (*Cyprinus carpio*). Small fish samples included juvenile Largemouth Bass (3-5"), juvenile Bluegill Sunfish (*Lepomis macrochirus*, 1-2"), Red Shiners (*Cyprinella lutrensis*), and Mosquitofish (*Gambusia affinis*, 1-2"). We collected 3-4 multi-individual composite samples from each of these 4 species from the Reiff basin (13 total). Small fish were also used to address the re-separation of the basins in the drought, with potentially different conditions in the separate ponds. In the former Mast Northwest basin, we were able to obtain closely comparable (to the Reiff basin samples) sets of Mosquitofish and Red Shiners, with 3 multi-individual composite samples each.

In total, this added up to 28 large fish muscle samples and 19 young, small fish composites, or 47 separate mercury samples analyzed from the drought-separated basins of the Esparto Pond in the Fall 2021 monitoring. The fish metrics and analytical results from each individual large fish muscle sample can be seen in Table 19 and, graphically, in Figure 19. Then, for each large fish species taken, the new data are shown in reduced form (means, error bars, etc) and compared to 2015-2020 results and the most closely comparable historic creek data (Tables and Figures 20-22). Results from the composite samples of small, young-of-year fish are similarly presented in Tables and Figures 23-27.

Large, Angling-sized Fish

Largemouth Bass (Tables/Figures 19 and 20)

We took a sample of 20 bass in 2021, as they became prevalent here and drought conditions hampered our ability to collect the other species. The bass present in 2021 were smaller on average than last year, ranging in size mostly from 210-328 mm (8-13") and 85-345 g (0.2-0.8 lbs), with two larger fish both 463 mm (14.3") and 470-520 g (1.0-1.1 lbs). Mercury concentrations increased with size, as is normal, with highest levels in the larger fish. The average mercury, across the full size range, was 1.648 ppm. Despite the smaller average size/age of the 2021 bass, mercury was up from 2020 (1.238 ppm) and was statistically similar to the peak levels seen in 2017-2018 (1.679-1.997 ppm). This site again had the highest bass mercury level among the currently monitored ponds. As compared to baseline data from Cache Creek, the Esparto Pond 2021 adult Largemouth Bass remained elevated; significantly above all comparisons.

The relative increase in bass mercury from last year (2020) was apparent across all fish sizes. In Figure 19(b), bass from 2021 are compared to the set from 2020. The 2021 fish were higher in mercury at every comparable size class. Additionally, the normal increase in mercury with size followed a steeper trend line in 2021. As we discussed in previous reports, we think that a shift in food web structure may have played a role, particularly the prevalence or absence of the small prey species Red Shiner. Bass first began showing up in collections here in 2017. At the time, Red Shiners were the dominant small fish. For whatever reason, these shiners were extremely high in mercury for such small/young fish, averaging up to 0.695 ppm. The initial peak bass mercury years of 2017-2018 coincided with the presence of shiners. As the bass numbers and sizes increased, the Red Shiners decreased drastically in abundance. The growing bass apparently targeted the high-mercury Red Shiners as their choice food, reducing them to very low numbers and reaching very high mercury levels themselves in the process. This apparently left only lowermercury foods for the bass. With the rebounding population of Red Shiners this year, again with very high mercury levels for small fish (see next section), bass mercury is apparently shooting up to levels similar to the last time Red Shiners were prevalent here. Time will tell how this settles, but this may be a prime example of how changes to food web structure can change mercury transfer to top predators like bass. This may be something to consider for future mercuryremediation plans – at any of the ponds.

White Catfish (Tables/Figures 19 and 21)

Six adult White Catfish were taken in 2021, in the size range of 387-473 mm (15-19") and 820-1250 g (1.8-2.7 lbs), larger on average than earlier samples. Muscle mercury ranged between 0.215 and 0.894 ppm, averaging 0.501 ppm. This was up from last year (2020: 0.408 ppm), but statistically similar to that year and 2019 (0.637 ppm). Even with the larger fish collected in 2021 and the rise in mercury from 2020, the average catfish mercury levels across the last three years have all been significantly lower than peak average levels found in 2016-2017 (0.996-1.287 ppm). However, <u>relative to Cache Creek comparison data</u>, the Esparto Pond 2021 White Catfish mercury levels remained elevated.

Carp (Tables/Figures 19 and 22)

Normal carp collections were closed off by drought conditions in 2021. Two adult Carp were taken in the large size range of 545-590 mm (21-23") and 2100-3050 g (4.6-6.7 lbs). Muscle mercury levels were a divergent 0.825 ppm in the larger fish and an extremely elevated 2.379 ppm in the other. The 2021 average of these two samples was 1.602 ppm, far above the averages from prior years. With just two very divergent samples, we cannot apply our normal statistics for comparisons, but the very high level in the smaller of the fish sampled, in particular, is consistent with the elevated 2021 mercury trends seen in the other species. In comparison with baseline Cache Creek samples, the Esparto Pond 2021 carp were elevated in mercury. The large 2021 carp had mercury levels similar to and even higher than the co-occurring bass. As we have noted in previous reports, this seems odd at first, as carp typically feed lower on the food chain (on lower-mercury food items) than the top-predator bass and would be expected to accumulate less mercury. Last year, we confirmed that this was probably due to differences in age; Largemouth Bass were fairly recent colonizers of this pond complex, first appearing in significant numbers in 2017. The resident carp, though, were clearly much older fish than the young population of bass, giving them many more years to slowly accumulate higher mercury than would be found in carp the same age as the bass. Last year, much smaller, younger carp were collected here for the first time. Carp of a similar age as the co-occurring bass had much lower mercury than the bass; still significantly above Creek baselines though.

Small, Young Fish

Mosquitofish (Tables/Figures 23 and 24)

Mosquitofish have been difficult to collect in some years, but we were able to obtain multiindividual composite samples in 2021 from both the Reiff and Mast Northwest basins of the site. Three sets of 12 Mosquitofish each were sampled from the Reiff basin, and three sets of 3-10 fish each from the Mast Northwest side. Sizes broadly matched previous collections here and at other sites through the years. Mercury in the three Reiff composite sets ranged from 0.526-0.638 ppm, averaging 0.579 ppm. This was dramatically up, more than double the already elevated levels found last year in the combined pond (2020: 0.267 ppm) and in Reiff Pond alone before 2020 (0.094-0.262 ppm). The increase was highly significant statistically. However, the Mosquitofish samples from the central (Mast NW) basin averaged 0.207 ppm, in the range seen over the previous five years in the combined pond and in Reiff alone. The 2021 Mast NW Basin samples were in fact significantly lower in mercury than the 2018 and 2020 sets. In relation to Cache <u>Creek comparison samples though, the 2021 concentrations from the Mast NW Basin were, on</u> <u>average, elevated</u>; they were statistically similar to one of three baseline sets and still significantly higher than two of the three. <u>The much higher mercury 2021 Mosquitofish from the Reiff Basin</u> <u>were significantly elevated, far above all the baseline creek comparisons</u>.

Red Shiner (*Tables/Figures 23 and 25*)

Red Shiners were not found in sufficient numbers for sampling in 2019 or 2020. The growing population of Largemouth Bass had preferentially consumed them down to very scarce numbers. But since the combining of the Reiff and Mast Basins in 2020, they have rebounded to collectible densities. Twelve Shiners were taken in the Reiff Basin and nine in the Mast NW Basin. Each set was divided into three size-class composite samples spanning the same size range (40-60 mm). The Mast NW samples ranged from 0.486 to an astounding (for such small, young fish) 1.379 ppm, averaging 0.951 ppm. The Reiff basin samples were even higher (significantly) on average, all in the range of 1.257-1.373 ppm and averaging 1.310 ppm. Compared to Cache Creek baseline samples, all of the Esparto Ponds Red Shiners in 2021 were significantly, highly elevated in mercury.

Juvenile Largemouth Bass (Tables/Figures 23 and 26)

Young-of-year Bass were very scarce in our collections this year; 4 individuals were taken from the Reiff Basin and none from Mast NW. The 4 young bass were analyzed individually. These samples had whole-body mercury ranging from 0.981-1.163 ppm, averaging 1.064 ppm. Like the Red Shiners, this was extremely high and was significantly above the already elevated levels found in prior years. <u>Relative to baseline juvenile bass comparison data from Cache Creek, 2021</u> juvenile Largemouth Bass mercury in the Reiff Basin continued to be elevated; significantly, far higher in mercury than the two creek sample sets available: River Mile 28 (0.142 ppm) and River Mile 15 (0.050 ppm).

Juvenile Sunfish (Tables/Figures 23 and 26)

Juvenile Green Sunfish were taken here in 2015 and 2018-2020 but, in 2021, the only sunfish present in our collections were three individuals of Bluegill Sunfish from the Reiff Basin. These were analyzed individually, whole-body like the other composites. At these small sizes, the different sunfish species are functionally equivalent and comparable in their mercury accumulation; the 2021 Bluegill Sunfish samples are presented with the previous Green Sunfish data in the tables and figures. The 2021 juvenile Bluegill had whole-body mercury at 0.451-0.508 ppm, averaging 0.479 ppm. This was significantly higher (about double) than the previous range for young sunfish here (0.187-0.252 ppm). <u>As compared to Cache Creek baseline comparison samples, the 2021 Reiff juvenile sunfish were significantly, far higher in mercury than all of the baseline sets (0.084-0.169 ppm).</u>

Fish	Fish Tot	al Length	Fish '	Weight	Muscle Mercury
Species	(mm)	(inches)	(g)	(lbs)	$(\mu g/g = ppm, wet w$
Largemouth Bass	210	8.3	85	0.2	1.175
Largemouth Bass	222	8.7	90	0.2	1.149
Largemouth Bass	235	9.3	130	0.3	1.091
Largemouth Bass	240	9.4	135	0.3	1.309
Largemouth Bass	249	9.8	155	0.3	1.365
Largemouth Bass	253	10.0	170	0.4	1.667
Largemouth Bass	255	10.0	150	0.3	1.279
Largemouth Bass	258	10.2	160	0.4	1.588
Largemouth Bass	260	10.2	175	0.4	1.051
Largemouth Bass	260	10.2	150	0.3	1.622
Largemouth Bass	265	10.4	180	0.4	1.559
Largemouth Bass	269	10.6	160	0.4	1.715
Largemouth Bass	270	10.6	205	0.5	1.612
Largemouth Bass	280	11.0	215	0.5	1.602
Largemouth Bass	302	11.9	245	0.5	1.802
Largemouth Bass	308	12.1	242	0.5	1.635
Largemouth Bass	312	12.3	300	0.7	2.380
Largemouth Bass	328	12.9	345	0.8	2.539
Largemouth Bass	363	14.3	470	1.0	2.346
Largemouth Bass	363	14.3	520	1.1	2.472
White Catfish	387	15.2	820	1.8	0.445
White Catfish	408	16.1	940	2.1	0.894
White Catfish	422	16.6	980	2.2	0.399
White Catfish	425	16.7	1,050	2.3	0.666
White Catfish	470	18.5	1,250	2.8	0.215
White Catfish	473	18.6	1,220	2.7	0.388
Carp	545	21.5	2,100	4.6	2.379
Carp	590	23.2	3,050	6.7	0.825

Table 19. Teichert – Esparto Pond: Individual large fish sampled, 2021

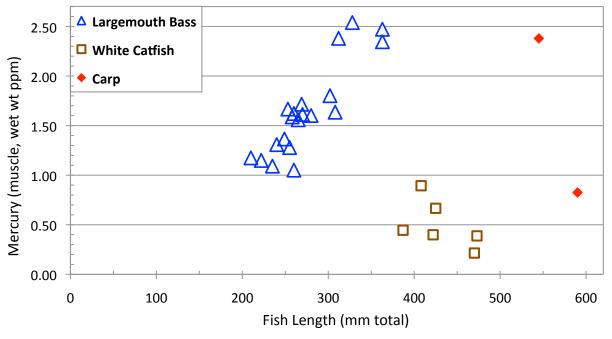


Figure 19. Teichert – Esparto Pond: large fish sampled, 2021 (*fillet muscle mercury in individual fish*)

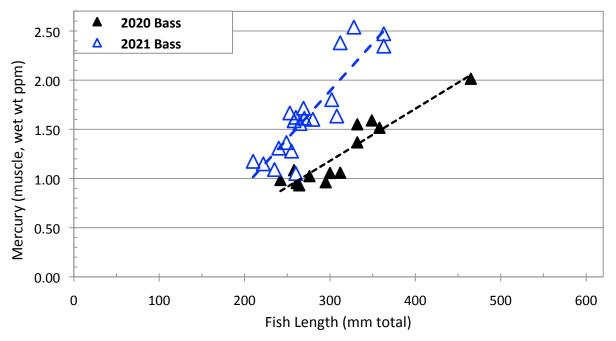
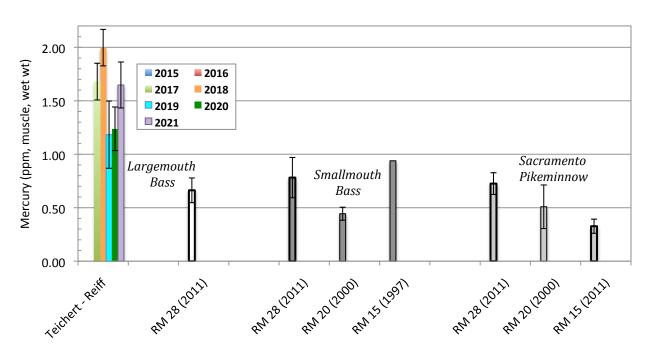
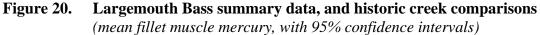


Figure 19(b). Teichert – Esparto 2020/2021 Largemouth Bass mercury comparison (*fillet muscle mercury in individual fish*)

Table 20. Largemouth Bass summary data, and historic creek comparisons

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (µg/g ppm, wet wt)	
Teichert-Esparto – Reiff	2017	5	189	78	1.679	± 0.180
Teichert-Esparto – Reiff	2018	10	251	181	1.997	± 0.170
Teichert-Esparto – Reiff	2019	10	295	353	1.183	± 0.314
Teichert – Esparto	2020	13	311	453	1.238	± 0.204
Teichert – Esparto (Mast NW)	2021	20	275	214	1.648	± 0.215
Historic/Baseline Data (c Largemouth Bass	•		•			
River Mile 28	2011	9	199	137	0.663	± 0.116
River Mile 28 Smallmouth Bass	2011	9	199	137	0.663	± 0.116
	2011 2011	9 7	199 265	137 326	0.663 0.782	± 0.116 ± 0.188
Smallmouth Bass		-				
Smallmouth Bass River Mile 28	2011	7	265	326	0.782	± 0.188
Smallmouth Bass River Mile 28 River Mile 20	2011 2000	7 7	265 234	326 183	0.782 0.444	± 0.188
Smallmouth Bass River Mile 28 River Mile 20 River Mile 15	2011 2000	7 7	265 234	326 183	0.782 0.444	± 0.188
Smallmouth Bass River Mile 28 River Mile 20 River Mile 15 Sacramento Pikeminnow	2011 2000 1997	7 7 2	265 234 383	326 183 780	0.782 0.444 0.939	± 0.188 ± 0.061





Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (μ g/g = ppm, wet wt)	= 95% C.I.
White Catfish						
Teichert-Esparto – Reiff	2015	20	347	658	0.737	± 0.156
Teichert-Esparto – Reiff	2016	20	297	341	0.996	± 0.153
Teichert-Esparto – Reiff	2017	16	355	677	1.287	± 0.197
Teichert-Esparto – Reiff	2018	_				
Teichert-Esparto – Reiff	2019	10	337	535	0.637	± 0.134
Teichert – Esparto	2020	10	369	742	0.408	± 0.123
Teichert – Esparto (Mast NW)	2021	6	431	1,043	0.501	± 0.253
Historic/Baseline Data						
Channel Catfish						
Rumsey	2000	1	411	565	0.225	
River Mile 28	2011	5	239	102	0.229	± 0.102
River Mile 20	2000	1	368	380	0.225	
River Mile 03	1997	10	336	304	0.174	± 0.019

Table 21. White Catfish summary data, and historic creek comparisons

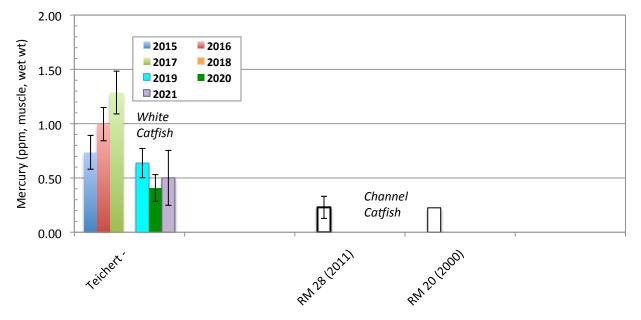


Figure 21. White Catfish summary data, and historic creek comparisons (mean fillet muscle mercury, with 95% confidence intervals)

Table 22. Carp summary data, and historic creek comparisons

Site	Site Year		6		Av Length (mm total)	Av Weight (grams)	Av Hg (μ g/g =95°ppm, wet wt)C.		
Carp									
Teichert-Esparto – Reiff	2015	2	421	918	0.351				
Teichert-Esparto – Reiff	2016	5	430	975	0.854	± 0.387			
Teichert-Esparto – Reiff	2017	9	481	1,499	1.122	± 0.321			
Teichert-Esparto – Reiff	2018	_							
Teichert-Esparto – Reiff	2019	9	483	1,475	0.988	± 0.279			
Teichert – Esparto	2020	7	381	1,086	0.636	± 0.334			
Teichert – Esparto (Mast NW)	2021	2	568	2,575	1.602				
Historic/Baseline Data (n Sacramento Sucker	iost com	parable spec	cies available)						
Rumsey	2000	6	328	396	0.198	± 0.113			
River Mile 20	2000	5	253	174	0.154	± 0.034			
River Mile 15	2011	8	276	231	0.143	± 0.011			
River Mile 08	2000	4	319	336	0.339				
River Mile 03	1997	5	343	402	0.263	± 0.068			

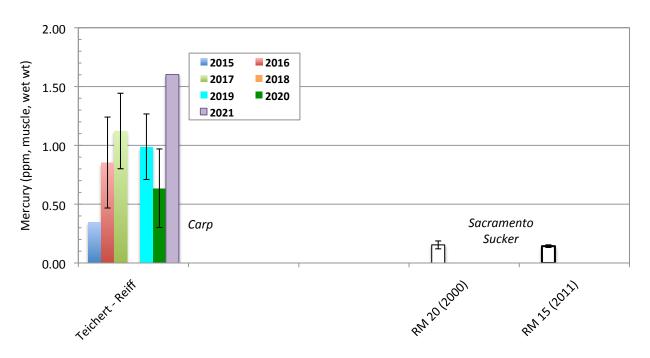


Figure 22. Carp summary data, and historic creek comparisons *(mean fillet muscle mercury, with 95% confidence intervals)*

Small, Young Fish Samples (note lower concentration scales)

Table 23. Teichert – Esparto Ponds: Small Fish Sampled, 2021

(multi-individual, whole body composite samples) '**n**' = number: number of individual fish per composite

Fish Species	n (indivs. in comp)	Av. Fish (mm)	Length (inches)	Av. Fish (g)	Weight (OZ)	Whole-Body Mercury $(\mu g/g = ppm, wet wt)$
Reiff (west side basin)						
Largemouth Bass (juv)	1	86	3.4	8.15	0.29	1.099
Largemouth Bass (juv)	1	94	3.7	9.95	0.35	1.014
Largemouth Bass (juv)	1	97	3.8	10.8	0.38	1.163
Largemouth Bass (juv)	1	113	4.4	18.9	0.67	0.981
Bluegill Sunfish (juv)	1	36	1.4	0.71	0.025	0.477
Bluegill Sunfish (juv)	1	38	1.5	0.98	0.034	0.451
Bluegill Sunfish (juv)	1	43	1.7	1.80	0.063	0.508
Mosquitofish	12	33	1.3	0.42	0.015	0.526
Mosquitofish	12	37	1.5	0.75	0.026	0.572
Mosquitofish	12	41	1.6	1.03	0.036	0.638
Red Shiners	4	47	1.9	1.26	0.044	1.373
Red Shiners	4	55	2.2	1.89	0.067	1.257
Red Shiners	4	59	2.3	2.15	0.076	1.298
Mast NW (central basin	n)					
Mosquitofish	10	27	1.1	0.23	0.008	0.216
Mosquitofish	10	32	1.3	0.40	0.014	0.175
Mosquitofish	3	39	1.5	0.79	0.028	0.229
Red Shiners	5	43	1.7	0.77	0.027	0.486
Red Shiners	3	51	2.0	1.30	0.046	0.987
Red Shiners	1	60	2.4	2.22	0.078	1.379

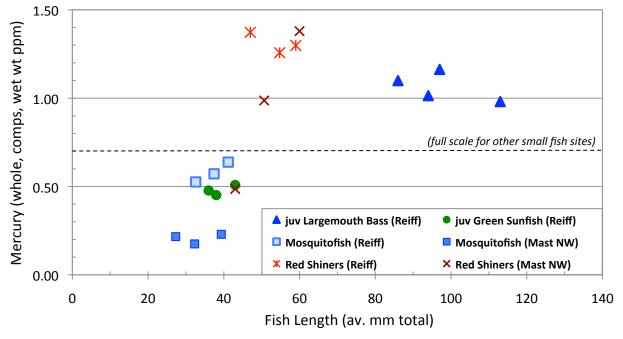


Figure 23. Teichert – Esparto Ponds: small, young fish sampled, 2021 (mercury in whole-body, multi-individual composite samples)

Small, Young Fish Samples (note lower concentration scales)

Table 24. Mosquitofish summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Mosquitofish							
Teichert-Esparto – Reiff	2015	4	12	38	0.6	0.094	± 0.005
Teichert-Esparto – Reiff	2016	4	10	36	0.5	0.212	± 0.021
Teichert-Esparto – Reiff	2017	_	_	_	_	_	
Teichert-Esparto – Reiff	2018	4	10	35	0.5	0.262	± 0.026
Teichert-Esparto – Reiff	2019	4	5-10	33	0.4	0.222	± 0.041
Teichert – <u>Esparto</u>	2020	4	1-12	37	0.7	0.267	± 0.033
Teichert – Esparto (Reiff)	2021	3	12	37	0.7	0.579	± 0.033
Teichert – Esparto (Mast N	W) 2021	3	3-10	33	0.5	0.207	± 0.016
Historic/Baseline Data	ļ						
River Mile 17	2000-2002	13	5-30	26-47	0.2-1.1	0.178	± 0.020
River Mile 15	2000-2002	10	5-30	26-47	0.2-1.0	0.100	± 0.018
River Mile 15	2011	4	1-10	37	0.7	0.103	± 0.024

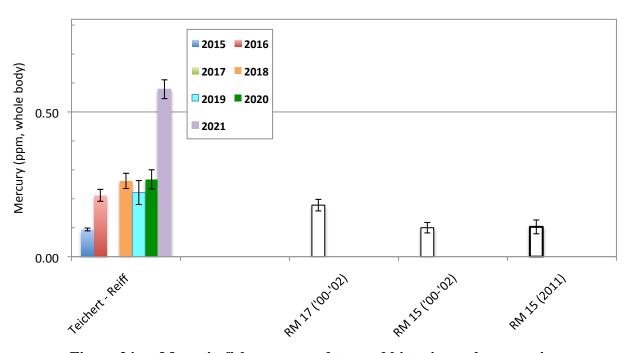


Figure 24. Mosquitofish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 25. Red Shiner summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Red Shiners							
Teichert-Esparto – Reiff	2015	4	10	50	1.3	0.152	± 0.009
Teichert-Esparto – Reiff	2016	4	10	47	1.1	0.412	± 0.042
Teichert-Esparto – Reiff	2017	4	10	49	1.1	0.695	± 0.070
Teichert-Esparto – Reiff	2018	4	10	45	0.8	0.556	± 0.031
Teichert-Esparto – Reiff	2019 (Shi	ners not fou	und in 2019	or 2020)			
Teichert – Esparto (Reiff)	2021	3	4	54	1.8	1.310	± 0.034
Teichert – Esparto (Mast NW	7) 2021	3	1-5	51	1.4	0.951	± 0.259
Historic/Baseline Data							
River Mile 28	2011	4	10	48	1.0	0.242	± 0.018
River Mile 20	2000	3	9	42	0.6	0.166	± 0.002
River Mile 17	2000-2002	11	6-15	27-58	0.2-1.8	0.225	± 0.023
River Mile 15	1997	3	19	37	0.5	0.159	± 0.014
	2000-2002	13	6-12	30-60	0.2-2.0	0.131	± 0.005

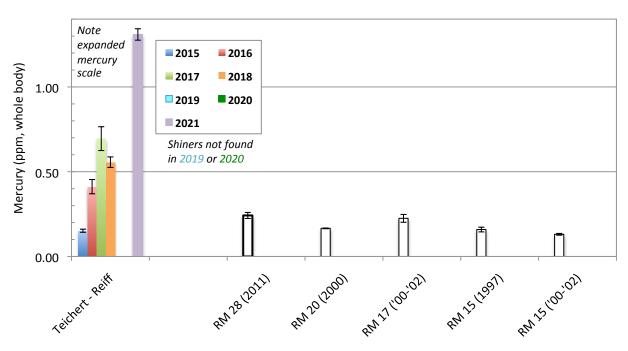


Figure 25. Red Shiner summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 26. Juvenile Largemouth Bass summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Largemouth Bass (juy	veniles)						
Teichert-Esparto – Reiff	2015	_	_				
Teichert-Esparto – Reiff	2016	_	_				
Teichert-Esparto – Reiff	2017	4	1-2	137	32	0.798	± 0.094
Teichert-Esparto – Reiff	2018	4	4-6	111	17	0.445	± 0.069
Teichert-Esparto – Reiff	2019	4	5	107	15	0.297	± 0.010
Teichert – Esparto	2020	4	3	92	9	0.472	± 0.027
Teichert – Esparto (Reiff)	2021	4	1	98	12	1.064	± 0.041
Historic/Baseline Data							
River Mile 28	2011	4	3-5	75	6	0.142	± 0.013
River Mile 15	2011	3	1	93	10	0.050	± 0.014

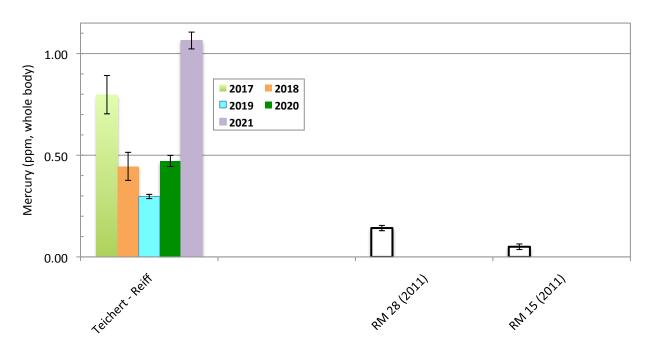


Figure 26. Juvenile Largemouth Bass summary data, and historic creek comparisons *(means of multiple whole-body, multi-individual composite samples)*

Table 27. Juvenile Green Sunfish summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Green Sunfish (juveni	les)						
Teichert-Esparto – Reiff	2015	1	1	68	5.1	0.241	
Teichert-Esparto – Reiff	2016	_	_				
Teichert-Esparto – Reiff	2017	_	_				
Teichert-Esparto – Reiff	2018	4	2	48	2.3	0.252	± 0.010
Teichert-Esparto – Reiff	2019	4	3-10	41	1.3	0.187	± 0.029
Teichert – Esparto	2020	4	3	35	0.7	0.230	± 0.018
Teichert – Esparto – (Reiff)	2021	3	1	39	1.2	0.479	± 0.016
Historic/Baseline Data							
River Mile 28	2011	4	4	53	2.8	0.139	± 0.014
River Mile 20	2011	4	4	58	3.4	0.084	± 0.004
River Mile 17	2000-2002	8	5-10	41-90	1-6	0.169	± 0.045
River Mile 15	2000-2002	8	4-8	40-87	1-6	0.117	± 0.028
River Mile 15	2011	4	4-5	56	3.1	0.086	± 0.018

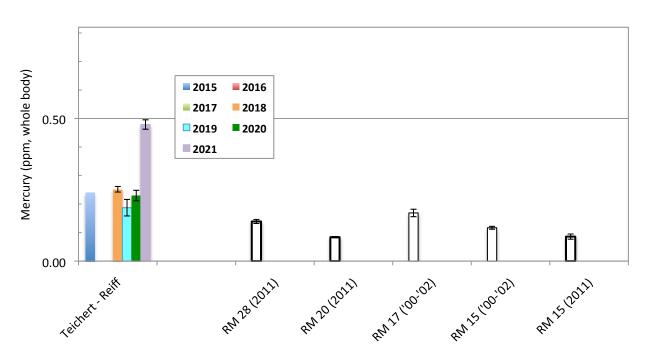


Figure 27. Juvenile Green Sunfish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)



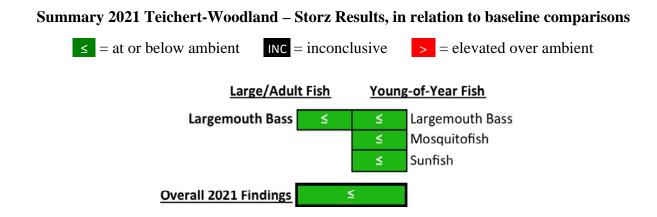
5. TEICHERT-WOODLAND – STORZ POND

(Google Earth 10/21/2020)

5. TEICHERT-WOODLAND – STORZ POND (Tables and Figures 28-33)

Summary

Drought conditions dried this pond down into two disconnected sub-basins in 2021. A sample of 18 small adult bass (the main size present) and 2 larger bass was taken. Small, young-of-year fish samples included multiple composites of Mosquitofish, juvenile Largemouth Bass, and juvenile Bluegill Sunfish. Mosquitofish were taken from both sub-basins to check for potential localized differences; no difference was found. The adult bass and all of the small fish samples remained in the lower range of mercury levels for this site, which has consistently been the first or second lowest mercury pond in the monitoring program. <u>Relative to Cache Creek comparison data, Storz Pond continued to rank as "not elevated over baseline" in 2021 and is not flagged for expanded analysis or management planning</u>. One-time-per-year routine water profiling was added to the monitoring in 2020, following recent revisions of the mining ordinance. With five years now with low mercury status, this site can shift to a schedule of fish and water testing once each two years rather than annually.



This pond is part of the Teichert–Woodland operations, located approximately 7 river miles downstream from the Reiff and Mast Ponds and Teichert–Esparto Plant. The Storz Pond is south of Cache Creek and just west of County Road 94b, near the Cache Creek Nature Preserve (which is located on the other, north, side of the creek). Our understanding is that it first become a wet pit in 2010-2011. The site has been relatively dormant and unmined over the course of mercury monitoring (since 2016); riparian and aquatic vegetation has colonized throughout, creating new habitat. The overall basin is approximately 150 m x 800 m (0.8 km) in size. Drought conditions dried this pond down into two disconnected, shallow sub-basins in 2021. Maximum depth in both parts was under 3 m (10 ft).

We began sampling this pond in 2016, but were unable to get our boat in at that time. By shore seining, we collected a good sample of Mosquitofish, (Gambusia affinis, 1-2") in 2016, but no additional species. In 2017, we were able to get our boat into the pond and sample more completely, making 2017 Year 1 of full sampling here. Since 2017, we have been able to collect Largemouth Bass (*Micropterus salmoides*) in addition to Mosquitofish. In 2021, 18 bass were taken in the small prevailing size range present of 200-235 mm (8-9"). An additional 2 bass were taken of a much larger size, 416-452 mm (16-18") and over ten times larger by weight. The 20 total bass were sampled for fillet muscle mercury. The bass were all taken in the Southwest Basin this year. Small fish were taken from both basins, to address the separation of the basins in the drought, with potentially different conditions in the separate ponds. Small fish collections were particularly difficult in 2021. Mosquitofish were sampled with 3 closely matching composites of 5-10 fish each from each side. The standard fourth size class (40-45 mm) was only available in the Northeast Basin. Three composite samples of juvenile Largemouth Bass were also taken in that basin. We were not able to collect juvenile bass from the Southwest Basin, despite extensive seining. In the Southwest Basin, in addition to the Mosquitofish, juvenile Bluegill Sunfish (Lepomis macrochirus) were found here for the first time and were sampled with 4 multiindividual composites.

In total, 20 large fish muscle samples and 14 small fish composite samples, or 34 separate mercury samples, were analyzed from the Teichert–Storz Pond(s) in the Fall 2021 monitoring. The fish metrics and analytical results from each individual large bass muscle sample can be seen in Table 28 and, graphically, in Figure 28. Then, the new data are shown in reduced form (means, error bars, etc) and compared to 2015-2020 results and the most closely comparable historic creek data (Table 29 and Figure 29). Results from the composite samples of small, young-of-year fish are similarly presented in Tables and Figures 30-33.

Large, Angling-sized Fish

Largemouth Bass (Tables/Figures 28 and 29)

As noted above, the bass samples included 18 fish from main cohort of fish present, in the small size range of 200-235 mm (8-9"), and two much larger individuals of 416 and 452 mm (16-18" and over ten times larger by weight). Fillet muscle mercury in the dominant size class ranged between 0.121 and 0.351 ppm, averaging 0.225 ppm, very low mercury levels for bass in this watershed. The two much larger fish had significantly higher mercury (0.534 and 0.809 ppm), averaging 0.672 ppm. Although the larger fish were higher in mercury than the smaller adults, the increase with size was muted, as compared to the trends at most of the other ponds. As can be seen in Figure 29, the average mercury levels in 2021 (0.269 ppm, including the two very large individuals), remained down significantly from the moderate levels of 2017 (0.657 ppm) and 2018 (0.611 ppm). As since 2019, the 2021 set of Storz fish was lower in mercury than bass samples from all the other monitored ponds. <u>Comparing with historic baseline Cache Creek samples, the Storz Pond 2021 adult largemouth bass remained unelevated</u>. They averaged lower mercury levels than all of the creek sample sets; the difference was statistically significant for five of seven comparisons.

Small, Young Fish

Mosquitofish (Tables/Figures 30 and 31)

Mosquitofish were the one species we could sample in both basins. Very similar samples were taken from each side in the first three standard size classes we use for them (25-30 mm, 30-35 mm, 35-40 mm). These ranged in mercury from 0.034-0.097 ppm, averaging 0.059 ppm in the Northeast Basin and 0.058 ppm in the Southwest, statistically identical and indicating no significant difference between the basins at this time. Including the standard fourth size class (40-45 mm) taken at the Northeast side, its average for inter-year comparisons rises to 0.082 ppm. This was in the low range for this generally low-mercury site, statistically similar to 2020 (0.059 ppm) and 2018 (0.087 ppm), and significantly lower than found in 2016, 2017, and 2019 (0.200-

0.282 ppm). It is interesting that there have been three years here with less than half the mercury levels found in the other three years. <u>As compared to baseline creek samples, the 2021 Storz</u> <u>Mosquitofish mercury levels were not elevated</u>; they were lower than all three of the creek data sets, which averaged 0.103-0.178 ppm. The differences were statistically significant for the Northeast Basin 4-size-class average in one of three baseline comparisons and, in for the Southwest Basin 3-size-class average, for all three.

Juvenile Largemouth Bass (Tables/Figures 30 and 32)

Juvenile bass were again very scarce, apparently due to cannibalism by larger bass, but we were able to collect 7 individuals from the Northeast Basin. These were analyzed in three whole-fish composite samples of 2-3 fish each. Mercury levels ranged from 0.100-0.129 ppm, averaging 0.118 ppm. This was within the range seen in the past two years (2019-2020: 0.097-0.131 ppm) and remained far below the levels found in the first full monitoring year (2017: 0.337 ppm). It was also significantly lower than at all but one of the other monitored ponds (Cemex–Phase 1, 0.096 ppm). As compared to baseline comparison samples from the creek, the 2021 Storz Pond juvenile Largemouth Bass were, on average, not elevated; they were significantly lower than the River Mile 28 set (0.142 ppm) and significantly higher than the River Mile 15 set (0.050 ppm).

Juvenile Bluegill Sunfish (Tables/Figures 30 and 33)

Sunfish were found here for the first time in 2021, in the Southwest Basin. At the small sizes collected for this monitoring, both of the sunfish species (Bluegill and Green) found across the region are functionally equivalent, accumulating mercury similarly and comparable to each other. Mercury in the multi-individual composites ranged from 0.015-0.024 ppm, averaging 0.020 ppm. <u>The 2021 juvenile Bluegill Sunfish from Storz Pond were not elevated over baseline;</u> they in fact had the lowest mercury levels found since 2015 in any of the seven ponds monitored or the Cache Creek baseline fish samples. All of the comparisons were statistically significant.

Fish	Fish Tot	al Length	Fish V	Weight	Muscle Mercury
Species	(mm)	(inches)	(g)	(lbs)	$(\mu g/g = ppm, wet wt)$
Largemouth Bass	200	7.9	78	0.2	0.229
Largemouth Bass	200	7.9	78	0.2	0.225
Largemouth Bass	203	8.0	68	0.1	0.263
Largemouth Bass	203	8.0	81	0.2	0.181
Largemouth Bass	204	8.0	90	0.2	0.191
Largemouth Bass	204	8.0	83	0.2	0.169
Largemouth Bass	205	8.1	85	0.2	0.214
Largemouth Bass	206	8.1	82	0.2	0.228
Largemouth Bass	206	8.1	92	0.2	0.278
Largemouth Bass	207	8.1	85	0.2	0.218
Largemouth Bass	211	8.3	92	0.2	0.209
Largemouth Bass	211	8.3	98	0.2	0.312
Largemouth Bass	214	8.4	93	0.2	0.121
Largemouth Bass	216	8.5	93	0.2	0.198
Largemouth Bass	217	8.5	105	0.2	0.198
Largemouth Bass	220	8.7	95	0.2	0.277
Largemouth Bass	233	9.2	122	0.3	0.184
Largemouth Bass	235	9.3	130	0.3	0.351
Largemouth Bass	416	16.4	1125	2.5	0.534
Largemouth Bass	452	17.8	1550	3.4	0.809

Table 28. Teichert-Woodland – Storz Pond: Individual large fish sampled, 2021

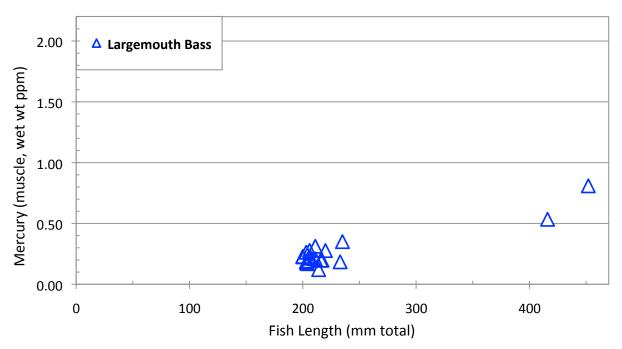
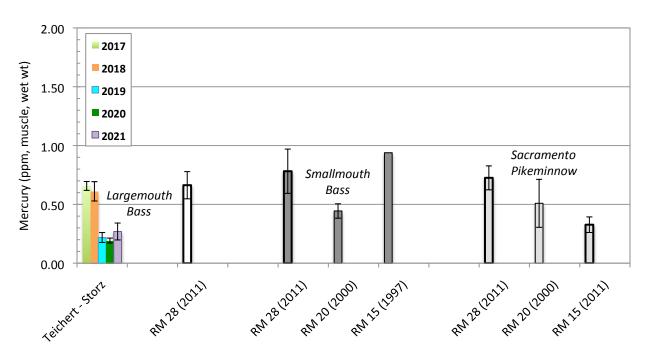


Figure 28. Teichert-Woodland – Storz Pond: Large Fish Sampled, 2021 (mean fillet muscle mercury, with 95% confidence intervals)

Table 29. Largemouth Bass summary data, and historic creek comparisons

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (μg/g ppm, wet wt)	
Teichert-Woodland – Storz	2017	20	245	203	0.657	± 0.038
Teichert-Woodland - Storz	2018	20	255	197	0.611	± 0.082
Teichert-Woodland – Storz	2019	12	222	196	0.218	± 0.042
Teichert-Woodland - Storz	2020	20	211	99	0.193	± 0.021
Teichert-Woodland - Storz	2021	20	233	216	0.269	± 0.072
Historic/Baseline Data (comparal	ble predatory	, species)			
Largemouth Bass River Mile 28	2011	9	199	137	0.663	± 0.116
0	2011	9	199	137	0.663	± 0.116
River Mile 28	2011 2011	9 7	199 265	137 326	0.663 0.782	± 0.116 ± 0.188
River Mile 28 Smallmouth Bass		-				
River Mile 28 Smallmouth Bass River Mile 28	2011	7	265	326	0.782	± 0.188
River Mile 28 Smallmouth Bass River Mile 28 River Mile 20	2011 2000	7 7	265 234	326 183	0.782 0.444	± 0.188
River Mile 28 Smallmouth Bass River Mile 28 River Mile 20 River Mile 15	2011 2000	7 7	265 234	326 183	0.782 0.444	± 0.188
River Mile 28 Smallmouth Bass River Mile 28 River Mile 20 River Mile 15 Sacramento Pikeminnow	2011 2000 1997	7 7 2	265 234 383	326 183 780	0.782 0.444 0.939	± 0.188 ± 0.061



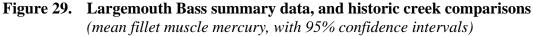
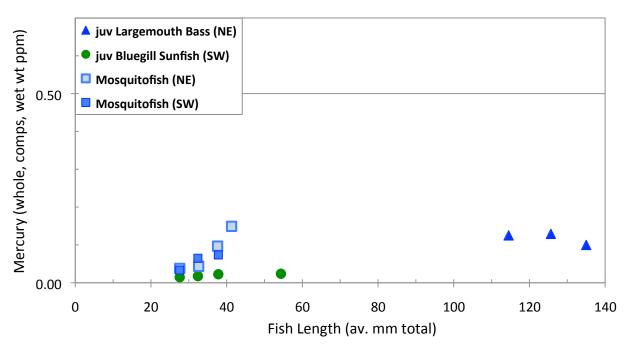


Table 30. Teichert-Woodland – Storz Ponds: Small Fish Sampled, 2021

(multi-individual, whole body composite samples) ' \mathbf{n} ' = number: number of individual fish per composite

Fish Species	n (indivs. in comp)	Av. Fish (mm)	Length (inches)	Av. Fis (g)	h Weight (oz)	Whole-Body Mercury $(\mu g/g = ppm, wet wt)$
Northeast Basin						
Largemouth Bass (juv)	2	115	4.5	22.3	0.79	0.125
Largemouth Bass (juv)	3	126	4.9	29.8	1.05	0.129
Largemouth Bass (juv)	2	135	5.3	34.7	1.22	0.100
Mosquitofish	10	28	1.1	0.22	0.008	0.038
Mosquitofish	10	33	1.3	0.41	0.014	0.043
Mosquitofish	10	38	1.5	0.64	0.023	0.097
Mosquitofish	10	41	1.6	1.02	0.036	0.149
Southwest Basin						
Bluegill Sunfish (juv)	10	28	1.1	0.49	0.017	0.015
Bluegill Sunfish (juv)	10	32	1.3	0.58	0.021	0.018
Bluegill Sunfish (juv)	5	38	1.5	1.34	0.047	0.023
Bluegill Sunfish (juv)	3	54	2.1	2.59	0.091	0.024
Mosquitofish	10	28	1.1	0.21	0.007	0.034
Mosquitofish	10	32	1.3	0.38	0.013	0.065
Mosquitofish	5	38	1.5	0.52	0.018	0.074



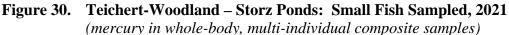


Table 31. Mosquitofish summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Mosquitofish							
Teichert-Woodland – Storz	2016	4	10	35	0.5	0.229	± 0.054
Teichert-Woodland – Storz	2017	4	8-10	29	0.2	0.282	± 0.011
Teichert-Woodland – Storz	2018	4	10	30	0.3	0.087	± 0.017
Teichert-Woodland – Storz	2019	4	6-10	33	0.4	0.200	± 0.018
Teichert-Woodland - Storz	2020	4	12	32	0.4	0.059	± 0.008
Teichert-Woodland – Storz N	<u>E</u> 2021	4	10	35	0.6	0.082	± 0.026
Teichert-Woodland – Storz SV	<u>v</u> 2021	3	5-10	33	0.4	0.057	± 0.012
Historic/Baseline Data							
River Mile 17	2000-2002	13	5-30	26-47	0.2-1.1	0.178	± 0.020
River Mile 15	2000-2002	10	5-30	26-47	0.2-1.0	0.100	± 0.018
River Mile 15	2011	4	1-10	37	0.7	0.103	± 0.024

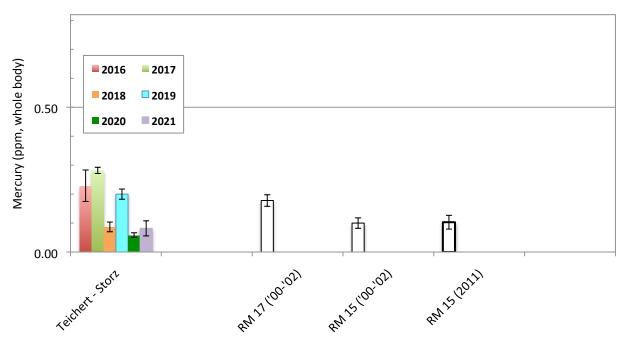


Figure 31. Mosquitofish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 32. Juvenile Largemouth Bass summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Largemouth Bass (juve	eniles)						
Teichert-Woodland – Storz	2017	4	1	143	35	0.337	± 0.030
Teichert-Woodland – Storz	2017	-	1	145	55	0.337	± 0.030
Teichert-Woodland – Storz	2010	4	1	130	29	0.131	± 0.036
Teichert-Woodland – Storz	2020	4	1	172	63	0.097	± 0.005 ± 0.005
Teichert-Woodland – Storz	2021	3	2-3	125	29	0.118	± 0.009
Historic/Baseline Data							
River Mile 28	2011	4	3-5	75	6	0.142	± 0.013
River Mile 15	2011	3	1	93	10	0.050	± 0.012 ± 0.014
		4 3			6 10		_

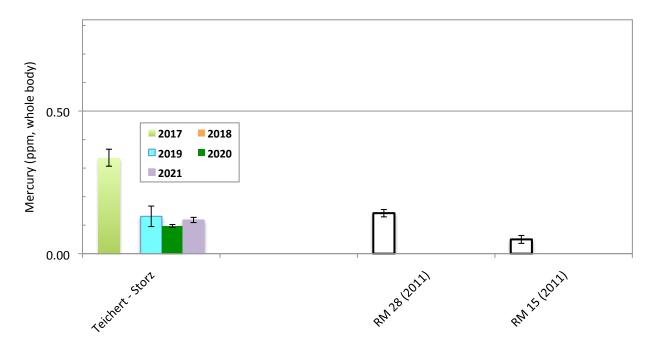


Figure 32. Juvenile Largemouth Bass summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 33. Juvenile Bluegill Sunfish summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	$\mathbf{Hg} \ (\mu g/g = ppm, wet wt)$	Std. Error
Bluegill Sunfish (juven	iles)						
0	,						
Teichert-Woodland – Storz	2017	—	—				
Teichert-Woodland – Storz	2018	—	—				
Teichert-Woodland – Storz	2019	—	_				
Teichert-Woodland – Storz	2020	_	_				
Teichert-Woodland - Storz SV	<u>w</u> 2021	4	3-10	38	1.3	0.020	± 0.002
Historic/Baseline Data	(juvenile C	Green Sun	fish)				
River Mile 28	2011	4	4	53	2.8	0.139	± 0.014
River Mile 20	2011	4	4	58	3.4	0.084	± 0.004
River Mile 17	2000-2002	8	5-10	41-90	1-6	0.169	± 0.045
	2000-2002	8	4-8	40-87	1-6	0.117	± 0.028
River Mile 15		0	.0	10 07	10	0.117	= 0.020

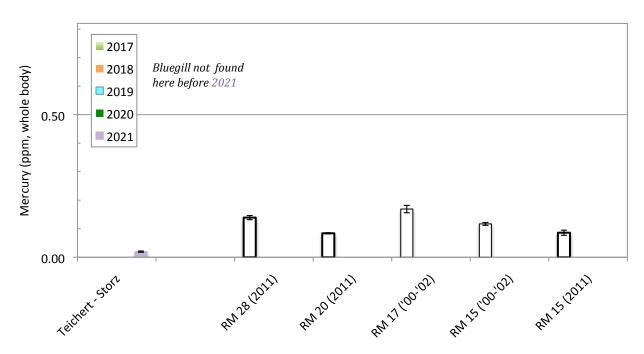


Figure 33. Juvenile Bluegill Sunfish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

6. SYAR – B1 POND

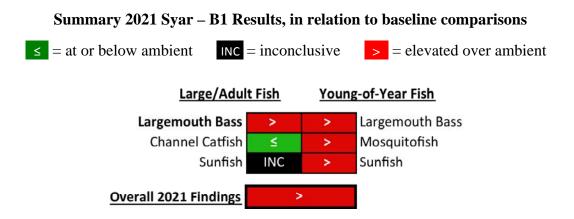


(Google Earth 10/21/2020)

6. SYAR-B1 POND (Tables and Figures 34-41)

Summary

Drought conditions affected both Syar Ponds significantly in 2020-2021, dropping water levels to far below normal. In addition to impacts on the ponds, it also made access and some fish collections difficult or not possible. Sixteen adult Largemouth Bass were sampled, one adult Bluegill Sunfish and for the first time here, four Channel Catfish. Young-of-year small fish were sampled with multiple composites of Mosquitofish, juvenile Largemouth Bass, and juvenile Bluegill Sunfish. Fish mercury was up in 2021 over the previous year for most sample types but remained significantly lower than the peak levels found here in 2015-2016. Despite the relatively lower levels in recent years, B1 Pond fish mercury in 2021 remained significantly higher on average than most baseline Cache Creek comparisons, most importantly in the Largemouth Bass. Because of the overall status of the B1 Pond as "elevated over baseline in three or more years of five" (all years since 2015), water column profiling and collection of bottom sediments was started here in 2018, in support of the development of a lake management plan. That work is detailed in accompanying reports.



The Syar Cache Creek mining operation, begun before 2002, has been idle since 2011 and remained inactive throughout the 7 years it has been monitored (2015-2021). The site is located south of Cache Creek and west of Highway 505, between 505 and County Road 87. There are

two mid-sized ponds at the site. We were provided access to the eastern pond of the two since 2015, and refer to that as the Syar–B1 Pond. It has an irregular shape about 500 m (0.5 km) long and 100-200 m wide. Beginning in 2017, we have also sampled the western pond (Syar–West), discussed in the next section. This (2021) was Year 7 of monitoring for the Syar–B1 Pond. The B1 Pond is located in a steep-sided surrounding depression. The shorelines are mostly steep, with the main area of the pond at a similar depth, within a meter or two of maximum depth. Drought conditions strongly impacted this site: maximum depth dropped from 7.9 m (26 feet) in October 2020 to 6.7 m (22 feet) in May 2021 to 3.0 m (10 feet) in September 2021.

As at the other sites, we sampled the B1 Pond on multiple days and with a range of techniques. The 2021 collections included a set of 16 Largemouth Bass (*Micropterus salmoides*), 4 Channel Catfish (*Ictalurus punctatus*) for the first time, and 1 adult Bluegill Sunfish (*Lepomis macrochirus*) for fillet muscle samples. The small, young fish present were juvenile Largemouth Bass (3-4"), juvenile Bluegill Sunfish (1-2") and Mosquitofish (*Gambusia affinis*, 1-2"). Each of these were sampled with 3-4 multi-individual composite samples as available, for 11 total.

In total, 21 large fish muscle samples and 11 young, small fish composite samples, or 32 separate mercury samples, were analyzed from the Syar–B1 Pond in the Fall 2021 monitoring. The fish metrics and analytical results from each individual large fish muscle sample can be seen in Table 34 and, graphically, in Figure 34. Then, for each large fish species taken, the new data are shown in reduced form (means, error bars, etc) and compared to 2015-2020 results and the most closely comparable historic creek data (Tables and Figures 35-37). Results from the composite samples of small, young-of-year fish are similarly presented in Tables and Figures 38-41.

Large, Angling-sized Fish

Largemouth Bass (Tables/Figures 34 and 35)

The B1 Pond adult Largemouth Bass samples included 16 fish across the range of adult sizes present: 279-435 mm (11-17") in length and 250-1050 g (0.5-2.3 lbs) in weight. In 15 of the 16 samples fillet muscle mercury ranged from 0.824-1.412 ppm, averaging 1.210 ppm. One fish was

notably higher, at 2.281 ppm. The overall average was 1.277 ppm. This was up from the previous four years (2017-2020: 0.904-1.095 ppm), though the difference was not statistically significant. It remained lower than the levels found in 2015-2016 when the bass averaged 1.628 and 1.640 ppm, extremely high fish mercury levels. <u>However, as compared to baseline samples</u> from Cache Creek, the 2021 B1 Pond adult Largemouth Bass remained clearly elevated in <u>mercury</u>; they were significantly higher than all comparison sets.

Green Sunfish / Bluegill Sunfish (Tables/Figures 34 and 36)

With the drought conditions in 2021 impacting boat access, we were not able to effectively sample sunfish in either of the Syar ponds. Only one adult Bluegill was taken. Muscle mercury was 0.335 ppm. This was lower for this site than all previous averages. <u>Relative to Cache Creek</u> comparisons, it was lower than one of four and higher than three. With just the single sample, we cannot make statistical comparisons; the 2021 Syar-B1 Sunfish mercury finding is 'inconclusive'.

Channel Catfish (Tables/Figures 34 and 37)

We found Channel Catfish here for the first time, while using baited set lines targeting bass. Four fairly large individuals were taken: 585-620 mm (23-25") in length and 1700-2300 g (3.7-5.1 lbs) in weight. Muscle mercury levels were surprisingly low at this site (for mid-high trophic level fish like channel catfish); the smallest fish had the highest concentration, at just 0.336 ppm. The three largest catfish had even lower concentrations of 0.164-0.191 ppm. Mercury concentration decreased with fish size, counter to typical trends. <u>The overall average was 0.220 ppm</u>, <u>statistically similar to all comparable Cache Creek baseline samples. Syar-B1 2021 Channel Catfish mercury was therefore in the "at or below baseline" category</u>. This was in contrast with the bass data from 2021 and all prior years, which were highly elevated. Possible explanations include: a) lower mercury diet items, (b) different (slower) metabolism, or (c) recent transfer/stocking from a lower mercury source by fishermen. This last would be consistent with the reverse size to mercury trend found.

Small, Young Fish

Juvenile Largemouth Bass (Tables/Figures 38 and 39)

The juvenile bass samples had whole-body mercury levels of 0.327-0.424 ppm, averaging 0.363 ppm. After five years of steady decline from 2015 (0.589 ppm) to 2020 (0.259 ppm), this was the first increase over the previous year. The increase above 2020 levels was significant, and was statistically similar to the three years before that (2017-2019: 0.338-0.461 ppm). It remained significantly lower than the highest levels seen here (2015-2016: 0.524-0.589 ppm). <u>Relative to baseline comparison data from Cache Creek, the 2021 BI Pond juvenile Largemouth Bass remained elevated; significantly higher than the two sample sets available: River Mile 28 (0.142 ppm) and River Mile 15 (0.050 ppm).</u>

Juvenile Sunfish (Tables/Figures 38 and 40)

Last year, Bluegill Sunfish became the dominant juvenile sunfish species present here, largely replacing Green Sunfish. Four extensive size-class composite samples were taken. As mentioned for some of the other sites, at these small sizes the two sunfish species are functionally equivalent, accumulating mercury in a comparable way. The 2021 juvenile Bluegill Sunfish composites had whole-body mercury of 0.203-0.232 ppm, averaging 0.222 ppm. This was up somewhat (significantly) from last year (2020: 0.181 ppm), though within the fairly consistent range seen in the last five years (2017-2021: 0.181-0.245 ppm). Similar to the juvenile bass, the juvenile sunfish here in 2017-2021 have had significantly lower mercury than in the initial monitoring years of 2015 and 2016 (0.325-0.414 ppm). <u>Relative to baseline juvenile Green Sunfish comparison numbers from Cache Creek though, the 2021 B1 Pond juvenile sunfish remained elevated</u>. The difference was statistically significant for all of the five comparisons.

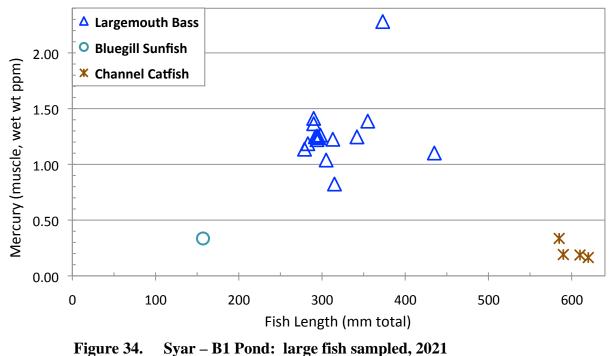
Mosquitofish (Tables/Figures 38 and 41)

The Mosquitofish samples had whole-body mercury ranging from 0.126-0.195 ppm, averaging 0.156 ppm. Like the other small fish species, this was up somewhat from 2020 (0.130 ppm); the difference was not statistically significant. Also like the other species, levels in recent years have remained significantly lower than in the initial monitoring years (2015-2017: 0.268-0.309 ppm). However, relative to the baseline Cache Creek data, the 2021 B1 Pond Mosquitofish mercury

<u>levels were once again elevated on average;</u> significantly elevated over two of three comparison sets and lower than, but statistically similar to, the third.

Fish Species	Fish Tot (mm)	(inches)	Fish (g)	Weight (lbs)	Muscle Mercury $(\mu g/g = ppm, wet wt)$
Largemouth Bass	279	11.0	250	0.6	1.136
Largemouth Bass	283	11.1	265	0.6	1.185
Largemouth Bass	290	11.4	245	0.5	1.412
Largemouth Bass	290	11.4	315	0.7	1.364
Largemouth Bass	292	11.5	290	0.6	1.247
Largemouth Bass	294	11.6	305	0.7	1.222
Largemouth Bass	294	11.6	295	0.7	1.254
Largemouth Bass	296	11.7	225	0.5	1.241
Largemouth Bass	298	11.7	340	0.7	1.263
Largemouth Bass	305	12.0	315	0.7	1.039
Largemouth Bass	313	12.3	385	0.8	1.225
Largemouth Bass	315	12.4	435	1.0	0.824
Largemouth Bass	342	13.5	435	1.0	1.246
Largemouth Bass	355	14.0	545	1.2	1.388
Largemouth Bass	373	14.7	505	1.1	2.281
Largemouth Bass	435	17.1	1,050	2.3	1.102
Channel Catfish	585	23.0	1,700	3.7	0.336
Channel Catfish	590	23.2	1,800	4.0	0.191
Channel Catfish	610	24.0	2,050	4.5	0.187
Channel Catfish	620	24.4	2,300	5.1	0.164
Bluegill Sunfish	157	6.2	65	0.1	0.335

Table 34. Syar – B1 Pond: Individual large fish sampled, 2021



(fillet muscle mercury in individual fish)

Table 35. Largemouth Bass summary data, and historic creek comparisons

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (μg/g ppm, wet w	
Syar – B1	2015	18	281	355	1.628	± 0.332
Syar – B1	2016	20	318	489	1.640	± 0.152
Syar – B1	2017	16	260	265	0.904	± 0.239
Syar – B1	2018	20	295	335	0.977	± 0.198
Syar – B1	2019	20	307	377	0.980	± 0.192
Syar – B1	2020	19	299	346	1.095	± 0.165
Syar – B1	2021	16	316	388	1.277	± 0.217

Largemouth Bass **River Mile 28** 2011 9 199 137 0.663 ± 0.116 Smallmouth Bass **River Mile 28** 7 0.782 2011 265 326 ± 0.188 River Mile 20 2000 7 234 183 0.444 ± 0.061 2 0.939 River Mile 15 1997 383 780 Sacramento Pikeminnow **River Mile 28** 2011 10 311 262 0.726 ± 0.102 River Mile 20 2000 8 269 147 0.509 ± 0.204 9 **River Mile 15** 2011 264 145 0.327 ± 0.066

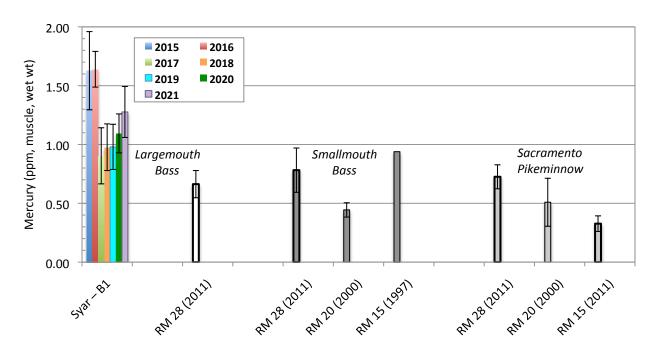


Figure 35. Largemouth Bass summary data, and historic creek comparisons (mean fillet muscle mercury, with 95% confidence intervals)

Table 36.	Gree	en and l	Bluegill Su	nfish summary	data, a	nd histo	ric cree	k comparisons
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Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (µg/g ppm, wet w	-
Green Sunfish						
Syar – B1	2015	10	118	25	0.777	± 0.086
Syar – B1	2016	1	83	12	1.446	
Syar – B1	2017	_				
Syar – B1	2018	_				
Syar – B1	2019	2	102	17	0.457	
Bluegill Sunfish						
Syar – B1	2020	10	157	63	0.602	± 0.051
Syar – B1	2021	1	157	65	0.335	
Historic/Baseline I	Data (Green Su	unfish)				
River Mile 28	2011	3	139	47	0.540	± 0.124
River Mile 20	2000	4	132	41	0.271	
River Mile 20	2011	10	122	31	0.138	± 0.029
River Mile 15	2011	10	133	41	0.195	± 0.031

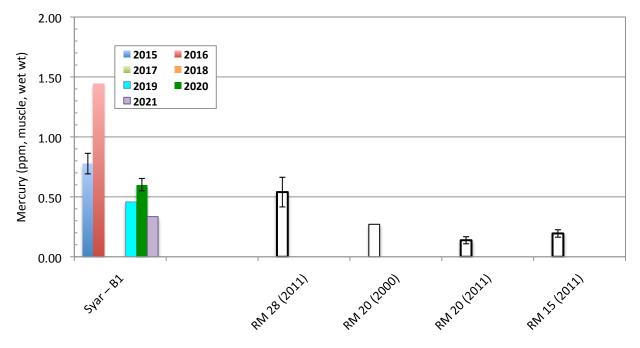


Figure 36. Sunfish summary data, and historic creek comparisons *(mean fillet muscle mercury, with 95% confidence intervals)*

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (μg/g ppm, wet w	-
Channel Catfish						
Syar – B1	2021	4	601	1,963	0.220	± 0.125
Historic/Baseline Data						
Channel Catfish						
Rumsey	2000	1	411	565	0.225	
River Mile 28	2011	5	239	102	0.229	± 0.102
River Mile 20	2000	1	368	380	0.225	
River Mile 03	1997	10	336	304	0.174	± 0.019

Table 37. Catfish summary data, and historic creek comparisons

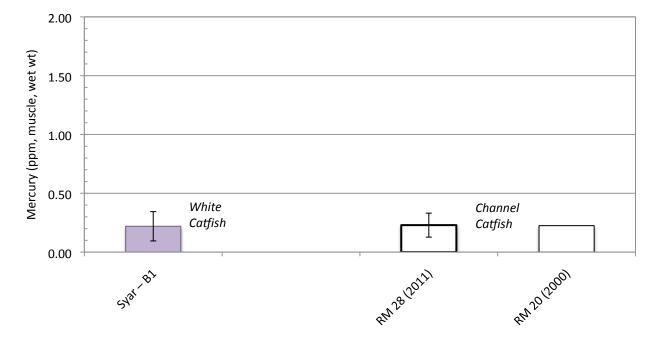


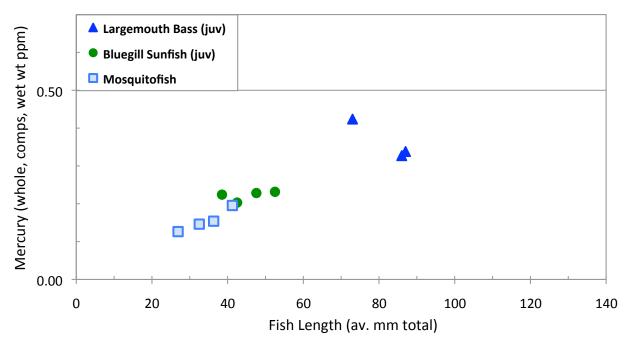
Figure 37. Catfish summary data, and historic creek comparisons (mean fillet muscle mercury, with 95% confidence intervals) (Rumsey and River Mile 3 similar, but outside mining/planning area)

Small, Young Fish Samples (note lower concentration scales)

Table 38. Syar – B1 Pond: Small Fish Sampled, 2021

Fish Species	n (indivs. in comp)	Av. Fis (mm)	h Length (inches)	Av. Fis (g)	h Weight (oz)	Whole-Body Mercury $(\mu g/g = ppm, wet wt)$
Largemouth Bass (juv)	1	73	2.9	4.77	0.17	0.424
Largemouth Bass (juv)	1	86	3.4	7.65	0.27	0.327
Largemouth Bass (juv)	1	87	3.4	8.30	0.29	0.338
Bluegill Sunfish (juv)	12	39	1.5	0.72	0.025	0.224
Bluegill Sunfish (juv)	12	43	1.7	1.11	0.039	0.203
Bluegill Sunfish (juv)	12	48	1.9	1.40	0.049	0.228
Bluegill Sunfish (juv)	10	53	2.1	2.07	0.073	0.232
Mosquitofish	10	27	1.1	0.20	0.007	0.126
Mosquitofish	10	33	1.3	0.34	0.012	0.146
Mosquitofish	10	36	1.4	0.52	0.018	0.154
Mosquitofish	4	41	1.6	0.85	0.030	0.195

(multi-individual, whole body composite samples) '**n**' = number: number of individual fish per composite



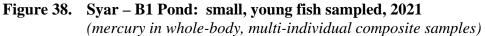


Table 39.Juvenile Largemouth Bass summary data, and historic creek comparisons
(means of multiple whole-body, multi-individual composite samples)
'n' = number: number of composite samples; number of individual fish per composite

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Largemouth Bass	s (juveniles)						
Syar – B1	2015	4	7	159	44	0.589	± 0.030
Syar – B1	2016	4	10	74	5	0.524	± 0.119
Syar – B1	2017	4	1-2	102	18	0.461	± 0.175
Syar – B1	2018	4	2	88	9	0.368	± 0.040
Syar – B1	2019	4	1	87	7	0.338	± 0.021
Syar – B1	2020	4	1	87	9	0.259	± 0.055
Syar – B1	2021	3	1	82	7	0.363	± 0.031
Historic/Baseline	Data						
River Mile 28	2011	4	3-5	75	6	0.142	± 0.013
River Mile 15	2011	3	1	93	10	0.050	± 0.012

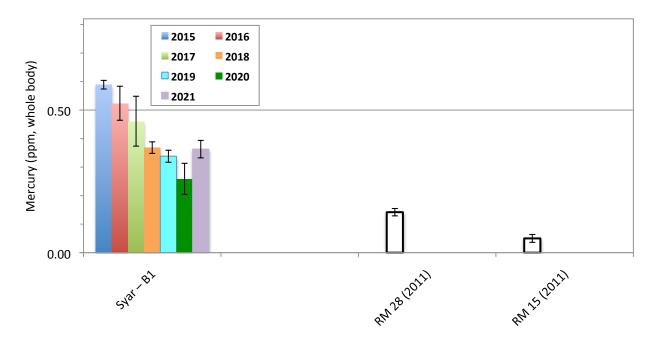


Figure 39. Juvenile Largemouth Bass summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 40. Juvenile Sunfish summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Green Sunfish (ju	veniles)						
Syar – B1	2015	4	8-9	47	1.7	0.325	± 0.097
Syar – B1	2016	4	4	50	1.9	0.414	± 0.076
Syar – B1	2017	4	6-7	40	1.0	0.225	± 0.069
Syar – B1	2018	4	10	37	0.8	0.231	± 0.044
Syar – B1	2019	4	8-10	45	1.5	0.245	± 0.016
Bluegill Sunfish (j	uveniles)						
Syar – B1	2020	4	12	44	1.3	0.181	± 0.018
Syar – B1	2021	4	10-12	45	1.3	0.222	± 0.006
Historic/Baseline I	Data						
River Mile 28	2011	4	4	53	2.8	0.139	± 0.007
River Mile 20	2011	4	4	58	3.4	0.084	± 0.002
River Mile 17	2000-2002	8	5-10	41-90	1-6	0.169	± 0.013
River Mile 15	2000-2002	8	4-8	40-87	1-6	0.117	± 0.005
River Mile 15	2011	4	4-5	56	3.1	0.086	± 0.009

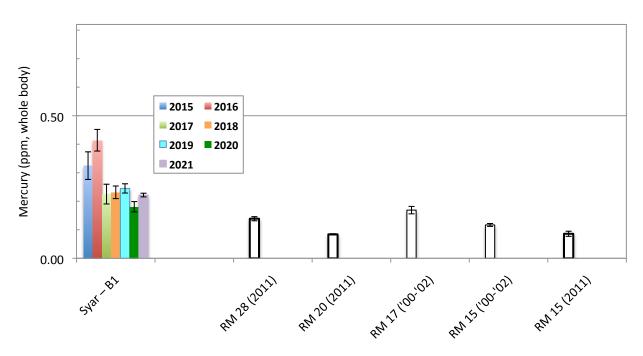


Figure 40. Juvenile Sunfish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 41. Mosquitofish summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Mosquitofish							
Syar – B1	2015	4	5-10	31	0.3	0.268	± 0.043
Syar – B1	2016	_	_	_	_	_	
Syar – B1	2017	4	9-10	35	0.4	0.309	± 0.110
Syar – B1	2018	4	6-9	31	0.4	0.163	± 0.056
Syar – B1	2019	3	1-3	38	0.7	0.214	± 0.011
Syar – B1	2020	4	6-12	34	0.5	0.130	± 0.026
Syar – B1	2021	4	4-10	34	0.5	0.156	± 0.015
Historic/Baseline	Data						
River Mile 17	2000-2002	13	5-30	26-47	0.2-1.1	0.178	± 0.020
River Mile 15	2000-2002	10	5-30	26-47	0.2-1.0	0.100	± 0.018
River Mile 15	2011	4	1-10	37	0.7	0.103	± 0.024

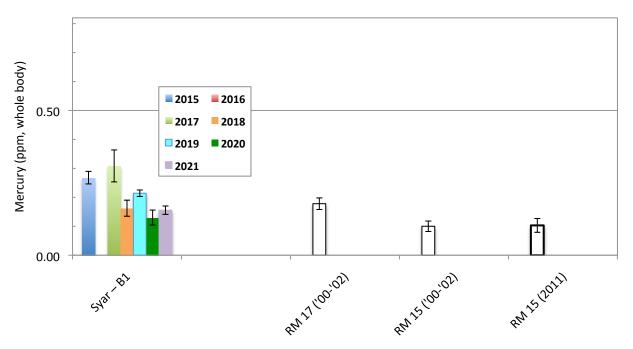


Figure 41. Mosquitofish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

7. SYAR – WEST POND

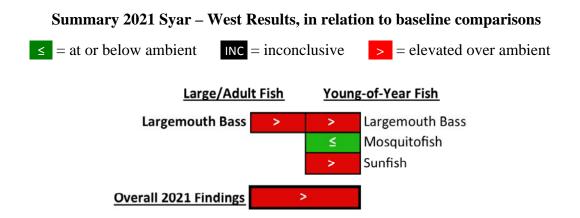


(Google Earth 10/21/2020)

7. SYAR – WEST POND (Tables and Figures 42-48)

Summary

Drought conditions affected both Syar Ponds significantly in 2020-2021, dropping water levels to far below normal. In addition to impacts on the ponds, it also made access and some fish collections difficult or not possible. Thirteen adult Largemouth Bass were sampled. Young-of-year small fish samples included multiple composites of Mosquitofish, juvenile Bluegill Sunfish, and juvenile Largemouth Bass. Fish mercury was up in 2021 over the previous year for most sample types; most importantly, the adult bass samples averaged higher than the bass from all previous years. Syar-West fish mercury in 2021 was significantly higher on average than most baseline Cache Creek comparisons, as in 2017, 2019, and 2020. Because of the overall status of the West Pond as "elevated over baseline in three or more years of five" as of 2020, expanded analysis and development of a lake management plan is required. Expanded analyses have in fact been conducted here since 2018, as a second control/reference site. This pond is far deeper than the other ponds currently, and is representative of the range of final depths projected at several of the sites. With elevated fish mercury status as of 2020, this work will help in the development of a lake management plan.



This pond is located about half a kilometer west of the B1 Pond; the overall Syar site and its history is described above in the section on the B1 Pond. The West Pond is approximately 300

m x 400 m in size. It has been dormant and unmined since 2011. The site was added to the monitoring program in 2017, in line with the Ordinance. This (2021) was Year 5 of monitoring for Syar-West. The basin is considerably deeper than all of the other ponds in the monitoring program at this time, with extensive areas more than 15 m (50 feet) deep – under normal conditions. As at the B1 Pond and the Teichert Ponds, the second year of drought strongly impacted Syar-West: maximum depth dropped from 16.5 m (54 feet) in September 2020 to 15.5 m (51 feet) in May 2021 to 10.1 m (33 feet) in October 2021.

Drought conditions impacted our ability to collect fish here. The dropping water level created a vertical, cliff edge around the pond. This made it impossible to access with a boat, or to do routine shoreline seining. We were able to obtain collect 13 Largemouth Bass (*Micropterus salmoides*) by spearfishing, free diving. These were sampled for filet muscle mercury. Small, young-of-year fish were taken by swim-seining; they included juvenile Bluegill Sunfish (*Lepomis macrochirus*, 1-2") and Mosquitofish (*Gambusia affinis*, 1-2"), each sampled with 4 multi-individual composites, plus 3 individual-fish composites of juvenile Largemouth Bass (3-4"), for 11 total composite samples.

In total, 13 large fish muscle samples and 11 small fish composite samples, or 24 separate mercury samples, were analyzed from the Syar–West Pond in the Fall 2021 monitoring. The fish metrics and analytical results from each individual large bass muscle sample can be seen in Table 42 and, graphically, in Figure 42. Then, the new data are shown in reduced form (means, error bars, etc) and compared to 2015-2020 results and the most closely comparable historic creek data (Table 43 and Figure 43); Sunfish data from previous years are given in Table/Figure 44. Results from the composite samples of small, young-of-year fish are similarly presented in Tables and Figures 45-48.

Large, Angling-sized Fish

Largemouth Bass (Tables/Figures 42 and 43)

Thirteen bass were sampled, across the size range of 284-440 mm (11-17") in length and 290-1470 g (0.6-3.2 lbs) in weight. The 2021 sample set averaged larger individuals than in other years here. Fillet muscle mercury ranged from 0.601-1.456 ppm, averaging 1.122 ppm. This was higher than in all prior monitoring years, and statistically elevated over the lowest levels seen here (2019: 0.672 ppm). Results may be at least partly a function of the larger average fish size/age, though levels were fairly flat across the sizes taken. Levels in 2021 were statistically similar to those found in similar bass from the nearby B1 Pond (1.277 ppm); bass data from both sites are plotted together in Figure 42(b). <u>Relative to historic/baseline creek comparisons, the 2021 West</u> <u>Pond adult Largemouth Bass sample were elevated in mercury</u>; they were higher than all of the seven comparison data sets; the elevation was statistically significant for all of the comparisons.

Green Sunfish / Bluegill Sunfish (Tables/Figures 42 and 44)

With no boat access in 2021, we were not able to sample sunfish effectively in either of the Syar ponds. Data from other years are presented.

Small, Young Fish

Juvenile Largemouth Bass (Tables/Figures 45 and 46)

We were able to collect just two, small juvenile bass from the Syar–West Pond. They had wholebody mercury of 0.168-0.185 ppm, averaging 0.176 ppm. This was higher statistically than the lowest levels found here to-date (2018: 0.153 ppm). But it was down significantly from the last samples available (2019: 0.273 ppm) and those from 2017 (0.418 ppm). As compared to corresponding samples from the adjacent B1 Pond (2021: 0.363 ppm), levels were significantly lower. <u>Relative to baseline juvenile bass comparison data from Cache Creek, the 2021 West Pond</u> juvenile bass remained elevated; significantly higher than both sets.

Juvenile Sunfish (Tables/Figures 45 and 47)

As found in the two Syar ponds since 2020, juvenile Bluegill Sunfish have mostly replaced Green Sunfish. As noted earlier, at the small sizes used for this monitoring, the two sunfish species are functionally equivalent and inter-comparable in their mercury accumulation. We collected

multiple-fish composite samples in the same size ranges used for the other sunfish. The 2021 fish had whole-body mercury ranging narrowly from 0.244-0.277 ppm, averaging 0.254 ppm. This was the highest average found to-date here; the difference was statistically significant over the prior three years (2018-2020: 0.102-0.187 ppm) and statistically similar to 2017 (0.237 ppm). In comparison to matching 2021 samples from the adjacent B1 Pond (0.222 ppm), West Pond mercury levels were significantly higher in 2021. <u>Relative to baseline/historic juvenile Green</u> <u>Sunfish from Cache Creek, the 2021 West Pond samples were significantly elevated in mercury levels over all five comparison sets</u>.

Mosquitofish (Tables/Figures 45 and 48)

We were able to collect four size-class composite samples of 5-10 Mosquitofish each. The composites had whole-body mercury ranging from 0.097-0.175 ppm, averaging 0.121 ppm. This was down significantly from 2019 (0.165 ppm) and 2017 (0.236 ppm), and was statistically similar to the lowest levels found here to-date (2018: 0.088 ppm). In comparison to matching 2021 samples from the adjacent B1 Pond (0.156 ppm), West Pond mercury levels were lower; the difference was significant statistically. <u>As compared to baseline Cache Creek sampling, the 2021</u> <u>West Pond Mosquitofish mercury levels were not elevated</u>; they were statistically similar to two of the baseline sets and significantly lower than one.

Fish	Fish Tot	al Length	Fish \	Weight	Muscle Mercury
Species	(mm)	(inches)	(g)	(lbs)	$(\mu g/g = ppm, wet wt)$
Largemouth Bass	284	11.2	290	0.6	1.051
Largemouth Bass	300	11.8	380	0.8	0.881
Largemouth Bass	307	12.1	340	0.7	1.221
Largemouth Bass	318	12.5	385	0.8	1.209
Largemouth Bass	328	12.9	430	0.9	0.601
Largemouth Bass	330	13.0	490	1.1	1.241
Largemouth Bass	343	13.5	465	1.0	1.228
Largemouth Bass	347	13.7	470	1.0	1.256
Largemouth Bass	377	14.8	690	1.5	1.151
Largemouth Bass	378	14.9	675	1.5	1.096
Largemouth Bass	417	16.4	1,100	2.4	1.003
Largemouth Bass	425	16.7	1,010	2.2	1.191
Largemouth Bass	440	17.3	1,470	3.2	1.456

Table 42. Syar – West Pond: Individual large fish sampled, 2021

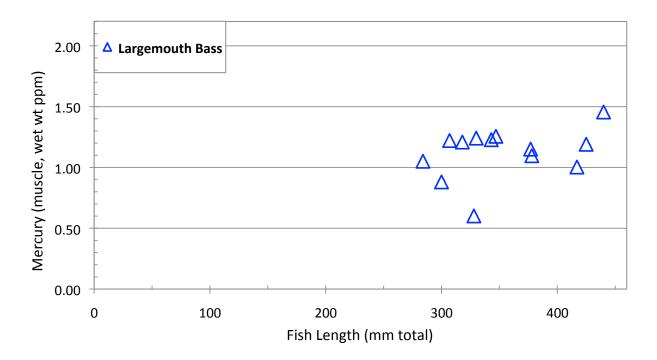


Figure 42.Syar – West Pond: large fish sampled, 2021
(fillet muscle mercury in individual fish)

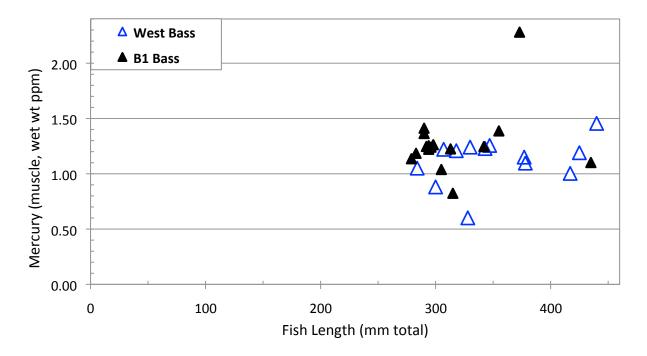


Figure 42(b). Syar – West Pond AND B1 Pond Bass comparison, 2021 (*fillet muscle mercury in individual fish*)

Table 43. Largemouth Bass summary data, and historic creek comparisons

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	$\mathbf{Av Hg} (\mu g/g = ppm, wet wt)$	95% C.I.
Syar – West	2017	17	283	320	0.925	± 0.205
Syar – West	2018	20	278	292	0.798	± 0.229
Syar – West	2019	20	275	271	0.672	± 0.105
Syar – West	2020	19	295	304	0.902	± 0.159
Svar – West	2021	13	353	630	1.122	± 0.161

	· •	1 2	1 /			
Largemouth Bass						
River Mile 28	2011	9	199	137	0.663	± 0.116
Smallmouth Bass						
River Mile 28	2011	7	265	326	0.782	± 0.188
River Mile 20	2000	7	234	183	0.444	± 0.061
River Mile 15	1997	2	383	780	0.939	
Sacramento Pikeminnow	,					
River Mile 28	2011	10	311	262	0.726	± 0.102
River Mile 20	2000	8	269	147	0.509	± 0.204
River Mile 15	2011	9	264	145	0.327	± 0.066

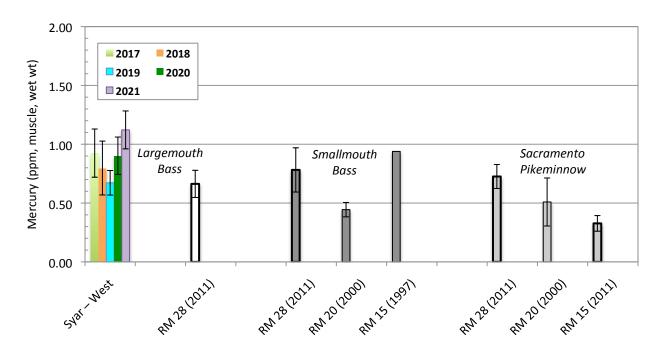


Figure 43. Largemouth Bass summary data, and historic creek comparisons (mean fillet muscle mercury, with 95% confidence intervals)

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (µg/g = ppm, wet wt)	
Green Sunfish						
Syar – West	2017	4	93	12	0.579	± 0.089
Syar – West	2018	_				
Syar – West	2019	1	126	41	0.238	
Bluegill Sunfish						
Syar – West	2020	10	185	121	0.612	± 0.068
Syar – West	2021	_				
Historic/Baseline Dat	a					
River Mile 28	2011	3	139	47	0.540	± 0.124
River Mile 20	2000	4	132	41	0.271	
River Mile 20	2011	10	122	31	0.138	± 0.029
River Mile 15	2011	10	133	41	0.195	± 0.031

Table 44. Green and Bluegill Sunfish summary data, and historic creek comparisons

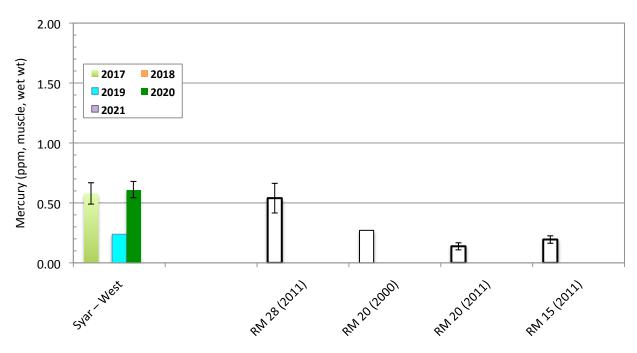


Figure 44.Sunfish summary data, and historic creek comparisons
(mean fillet muscle mercury, with 95% confidence intervals)

Small, Young Fish Samples (note lower concentration scales)

Table 45. Syar – West Pond: Small Fish Sampled, 2021

(*multi-individual*, *whole body composite samples*) '**n**' = *number: number of individual fish per composite*

Fish Species	n (indivs. in comp)	Av. Fish (mm)	Length (inches)	Av. Fisl (g)	n Weight (oz)	Whole-Body Mercury $(\mu g/g = ppm, wet wt)$
Largemouth Bass (juv)	1	53	2.1	2.04	0.072	0.168
Largemouth Bass (juv)	1	56	2.2	2.89	0.102	0.185
Bluegill Sunfish (juv)	10	37	1.4	0.53	0.019	0.244
Bluegill Sunfish (juv)	10	41	1.6	0.81	0.029	0.251
Bluegill Sunfish (juv)	10	47	1.9	1.15	0.041	0.245
Bluegill Sunfish (juv)	10	54	2.1	2.11	0.074	0.277
Mosquitofish	10	28	1.1	0.21	0.007	0.097
Mosquitofish	10	33	1.3	0.36	0.013	0.097
Mosquitofish	10	36	1.4	0.50	0.018	0.116
Mosquitofish	5	40	1.6	0.71	0.025	0.175

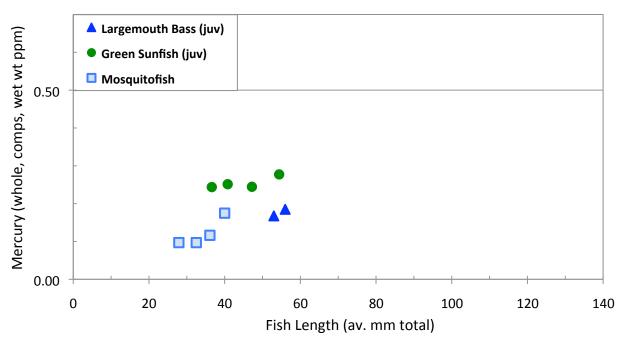


Figure 45. Syar – West Pond: small, young fish sampled, 2021

Table 46. Juvenile Largemouth Bass summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Largemouth Bass	(juveniles)						
Syar – West	2017	2	1	123	27	0.418	± 0.030
Syar – West	2018	4	2	77	6	0.153	± 0.024
Syar – West	2019	2	1	96	11	0.273	± 0.006
Syar – West	2020	(none take	en)				
Syar – West	2021	2	1	55	2	0.176	± 0.007
Historic/Baseline I	Data						
River Mile 28	2011	4	3-5	75	6	0.142	± 0.013
River Mile 15	2011	3	1	93	10	0.050	± 0.014

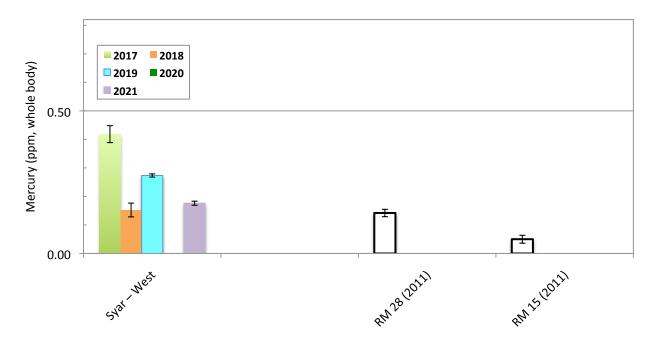


Figure 46. Juvenile Largemouth Bass summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 47. Juvenile Sunfish summary data, and historic creek comparisons

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
	•••						
Green Sunfish (ju							
Syar – West	2017	4	5-10	45	1.7	0.237	± 0.077
Syar – West	2018	4	2-4	34	0.6	0.102	± 0.017
Syar – West	2019	4	8-10	46	1.5	0.177	± 0.010
Bluegill Sunfish (j	uveniles)						
Syar – West	2020	4	10-12	42	1.2	0.187	± 0.024
Syar – West	2021	4	10	45	1.1	0.254	± 0.008
Historic/Baseline I	Data						
River Mile 28	2011	4	4	53	2.8	0.139	± 0.007
River Mile 20	2011	4	4	58	3.4	0.084	± 0.002
River Mile 17	2000-2002	8	5-10	41-90	1-6	0.169	± 0.013
River Mile 15	2000-2002	8	4-8	40-87	1-6	0.117	± 0.005
River Mile 15	2000 2002 2011	4	4-5	-0 0/ 56	3.1	0.086	± 0.009

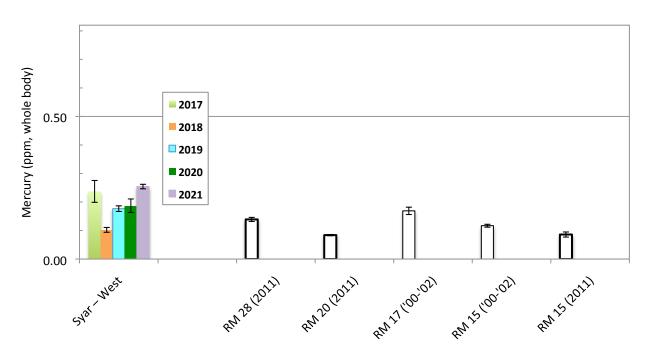


Figure 47. Juvenile Sunfish summary data, and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 48. Mosquitofish summary data, and historic creek comparisons

(means of multiple whole-body, multi-individual composite samples) '**n**' = number: number of composite samples; number of individual fish per composite

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Mosquitofish							
Syar – West	2017	4	10	34	0.4	0.236	± 0.034
Syar – West	2018	4	6-7	29	0.3	0.088	± 0.012
Syar – West	2019	3	2-3	36	0.6	0.165	± 0.032
Syar – West	2020	4	8-12	35	0.5	0.109	± 0.018
Syar – West	2021	4	5-10	34	0.4	0.121	± 0.018
Historic/Baseline I	Data						
River Mile 17	2000-2002	13	5-30	26-47	0.2-1.1	0.178	± 0.020
River Mile 15	2000-2002	10	5-30	26-47	0.2-1.0	0.100	± 0.018
River Mile 15	2011	4	1-10	37	0.7	0.103	± 0.024

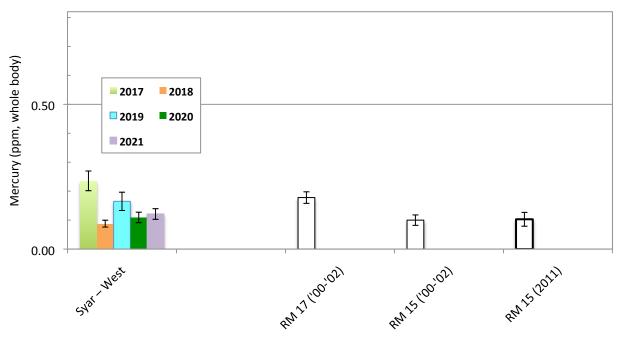


Figure 48.Mosquitofish summary data, and historic creek comparisons
(means of multiple whole-body, multi-individual composite samples)

8. COMPARISON OF ALL THE MONITORED SITES AND HISTORICAL DATA, BY FISH SPECIES

This section is presented to consolidate the monitoring data and place the various findings into relative context. For each sample type, data are first presented in a table and then graphically with an accompanying figure. These presentations allow the reader (and these researchers) to assess overall trends, across all of the monitored ponds and over time.

 ± 0.217

 ± 0.205

 ± 0.229

 ± 0.105

 ± 0.159

 ± 0.161

1.277

0.925

0.798

0.672

0.902

1.122

Table 49. Largemouth Bass summary data (all sites) and historic creek comparisons

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (μ g/g = ppm, wet wt)	95% C.I.
Largemouth Bass						
Cemex – Phase 1	2015	18	305	393	0.278	± 0.055
Cemex – Phase 1	2016	20	313	383	0.350	± 0.066
Cemex – Phase 1	2017	17	299	357	0.393	± 0.079
Cemex – Phase 1	2018	20	298	331	0.481	± 0.131
Cemex – Phase 1	2019	20	280	247	0.404	± 0.085
Cemex – Phase 1	2020	20	267	253	0.352	± 0.075
Cemex – Phase 1	2021	20	274	262	0.381	± 0.079
Cemex – Phase 3-4	2015	20	344	526	0.840	± 0.113
Cemex – Phase 3-4	2016	20	344	557	0.858	± 0.139
Cemex – Phase 3-4	2017	20	334	479	1.093	± 0.172
Cemex – Phase 3-4	2018	20	331	463	0.918	± 0.119
Cemex – Phase 3-4	2019	20	312	402	0.819	± 0.195
Cemex – Phase 3-4	2020	20	310	399	0.656	± 0.094
Cemex – <u>Phase 3</u>	2021	20	358	575	0.971	± 0.092
Cemex – <u>Phase 4</u>	2021	20	325	418	0.782	± 0.105
Teichert-Esparto – Reiff	2017	5	189	78	1.679	± 0.180
Teichert-Esparto – Reiff	2018	10	251	181	1.997	± 0.170
Teichert-Esparto – Reiff	2019	10	295	353	1.183	± 0.314
Teichert – <u>Esparto</u>	2020	13	311	453	1.238	± 0.204
Teichert – <u>Esparto</u>	2021	20	275	214	1.648	± 0.215
Teichert-Woodland – Storz	2017	20	245	203	0.657	± 0.038
Teichert-Woodland – Storz	2018	20	255	197	0.611	± 0.082
Teichert-Woodland – Storz	2019	12	222	196	0.218	± 0.042
Teichert-Woodland – Storz	2020	20	211	99	0.193	± 0.021
Teichert-Woodland - Storz	2021	20	233	216	0.269	± 0.072
Syar – B1	2015	18	281	355	1.628	± 0.332
Syar – B1	2016	20	318	489		± 0.152
Syar – B1	2017	16	260	265		± 0.239
Syar – B1	2018	20	295	335		± 0.198
Syar – B1	2019	20	307	377	0.980	± 0.192
Syar – B1	2020	19	299	346	1.095	± 0.165
Course D1	2021	16	216	200	1 077	0.017

(mean fillet muscle mercury, with 95% confidence intervals)

(continued next page)

316

283

278

275

295

353

388

320

292

271

304

630

2021

2017

2018

2019

2020

2021

16

17

20

20

19

13

Syar - B1

Syar-West

Syar – West

Syar – West

Syar - West

Syar-West

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (μg/g ppm, wet w	
Historic/Baseline L	Data (comparal	ble predatory	v species)			
Largemouth Bass						
River Mile 28	2011	9	199	137	0.663	± 0.116
Smallmouth Bass						
River Mile 28	2011	7	265	326	0.782	± 0.188
River Mile 20	2000	7	234	183	0.444	± 0.061
River Mile 15	1997	2	383	780	0.939	
Sacramento Pikeminn	ow					
River Mile 28	2011	10	311	262	0.726	± 0.102
River Mile 20	2000	8	269	147	0.509	± 0.204
River Mile 15	2011	9	264	145	0.327	± 0.066

(Table 49, continued)

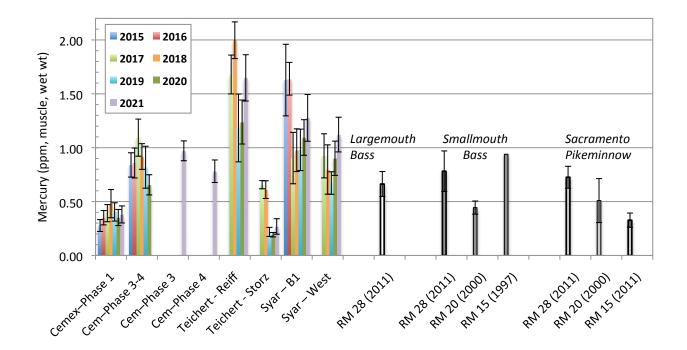


Figure 49. Largemouth Bass summary data, and historic creek comparisons (mean fillet muscle mercury, with 95% confidence intervals)

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (μ g/g = ppm, wet wt)	95% C.I.
Channel Catfish						
Cemex – Phase 1	2015	2	595	2,130	0.198	
Cemex – Phase 1	2016	2	412	1,150	0.100	
Cemex – Phase 1	2017	2	531	1,440	0.236	
Cemex – Phase 1	2018	3	533	1,973	0.337	± 0.58
	(Catfish -	– both species	– not found at Ce	mex–Phase 1 sind	ce 2018)	
Syar – B1	2021	4	601	1,963	0.220	± 0.125
White Catfish						
Cemex – Phase 1	2016	3	661	2,900	0.372	
Cemex – Phase 1	2017	6	615	2,120	0.448	± 0.134
Cemex – Phase 1	2018	1	398	1115	0.571	
Teichert-Esparto – Reiff	2015	20	347	658	0.737	± 0.156
Teichert-Esparto – Reiff	2016	20	297	341	0.996	± 0.153
Teichert-Esparto – Reiff	2017	16	355	677	1.287	± 0.197
Teichert-Esparto – Reiff	2018	(unable to	sample in 2018)			
Teichert-Esparto – Reiff	2019	10	337	535	0.637	± 0.134
Teichert – Esparto	2020	10	369	742	0.408	± 0.123
Teichert – <u>Esparto</u>	2021	6	431	1,043	0.501	± 0.253
Historic/Baseline Data						
Channel Catfish						
Rumsey	2000	1	411	565	0.225	
River Mile 28	2011	5	239	102	0.229	± 0.102
River Mile 20	2000	1	368	380	0.225	
River Mile 03	1997	10	336	304	0.174	± 0.019

Table 50. Catfish summary data (all sites) and historic creek comparisons

(mean fillet muscle mercury, with 95% confidence intervals)

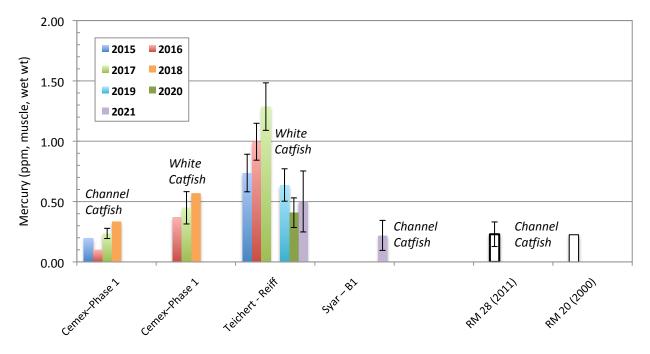


Figure 50. Catfish summary data (all sites) and historic creek comparisons (mean fillet muscle mercury, with 95% confidence intervals) (Rumsey and River Mile 3 were outside the mining/planning area)

Table 51. Sunfish summary data (all sites) and historic creek comparisons

(mean fillet muscle mercury, with 95% confidence intervals)

Site	Year	Number of Fish	Av Length (mm total)	Av Weight (grams)	Av Hg (μ g/g = ppm, wet wt)	95% C.I.
Green Sunfish (unless	noted Blue	egill)				
Cemex – Phase 1	2017	5	105	35	0.273	± 0.094
Cemex – Phase 1	2018	1	200	165	0.227	
Cemex – Phase 3-4	2015	10	133	67	0.534	± 0.076
Cemex – Phase 3-4	2016	1	101	16	0.382	
Cemex – Phase 3-4	2017	_				
Cemex – Phase 3-4	2018	-				
Teichert-Esparto – Reiff	2015	1	140	40	0.328	
Teichert-Esparto – Reiff	2016	_				
Teichert-Esparto – Reiff	2017	—				
Teichert-Esparto – Reiff	2018	—				
Teichert-Esparto – Reiff	2019	1	106	23	0.373	
Syar – B1	2015	10	118	25	0.777	± 0.086
Syar – B1	2016	1	83	12	1.446	
Syar – B1	2017	—				
Syar – B1	2018	—				
Syar – B1	2019	2	102	17	0.457	
Syar – B1 *Bluegill*	2020	10	157	63	0.602	±0.072
Syar – B1 *Bluegill*	2021	1	157	65	0.335	
Syar – West	2017	4	93	12	0.579	± 0.089
Syar – West	2018	—				
Syar – West	2019	1	126	41	0.238	
Syar – West *Bluegill*	2020	10	185	121	0.612	± 0.095
Historic/Baseline Data						
River Mile 28	2011	3	139	47	0.540	± 0.124
River Mile 20	2000	4	132	41	0.271	. 0.020
River Mile 20 River Mile 15	2011 2011	10 10	122 133	31 41	0.138 0.195	± 0.029 ± 0.031
NIVEL WHIE 15	2011	10	155	41	0.195	± 0.031

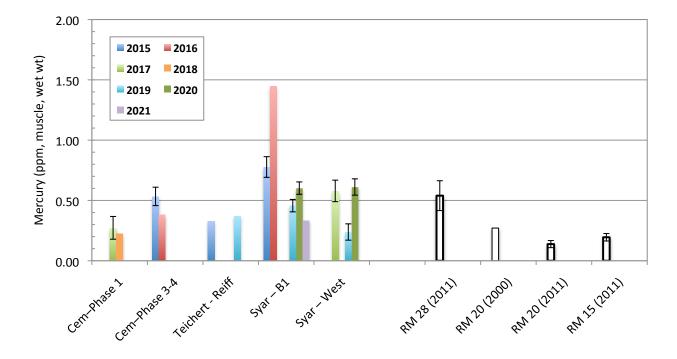


Figure 51. Sunfish summary data (all sites) and historic creek comparisons (mean fillet muscle mercury, with 95% confidence intervals)

Table 52. Carp summary data (all sites) and historic creek comparisons

Year	Number of Fish	Av Length (mm total)	Av Weight (grams)		
2015	2	421	918	0.351	
2016	5	430	975	0.854	± 0.387
2017	9	481	1,499	1.122	± 0.321
2018	(unable to sam	ple)			
2019	9	483	1,475	0.988	± 0.279
2020	7	381	1,086	0.636	± 0.334
2021	2	568	2,575	1.602	
nost cor	nparable spec	ies available)			
2000	6	328	396	0.198	± 0.113
2000	5	253	174	0.154	± 0.034
2011	8	276	231	0.143	± 0.011
2000	4	319	336	0.339	
1997	5	343	402	0.263	± 0.068
	2015 2016 2017 2018 2019 2020 2021 most cor 2000 2000 2000 2011 2000	of Fish 2015 2 2016 5 2017 9 2018 (unable to sam) 2019 9 2020 7 2021 2 nost comparable spec 2000 6 2000 5 2011 8 2000 4	of Fish (mm total) 2015 2 421 2016 5 430 2017 9 481 2018 (unable to sample) 2019 2019 9 483 2020 7 381 2021 2 568 most comparable species available) 2000 6 328 2000 5 253 2011 8 276 2000 4 319	of Fish(mm total)(grams)2015242191820165430975201794811,4992018(unable to sample)20199201994831,475202073811,086202125682,575most comparable species available)20006328396200052531742011827623120004319336	of Fish(mm total)(grams)ppm, wet wt)20152421918 0.351 20165430975 0.854 20179481 $1,499$ 1.122 2018(unable to sample)2019948320199483 $1,475$ 0.988 20207381 $1,086$ 0.636 20212568 $2,575$ 1.602 most comparable species available)200062000525317420118276231 0.143 20004319336 0.339

(mean fillet muscle mercury, with 95% confidence intervals)

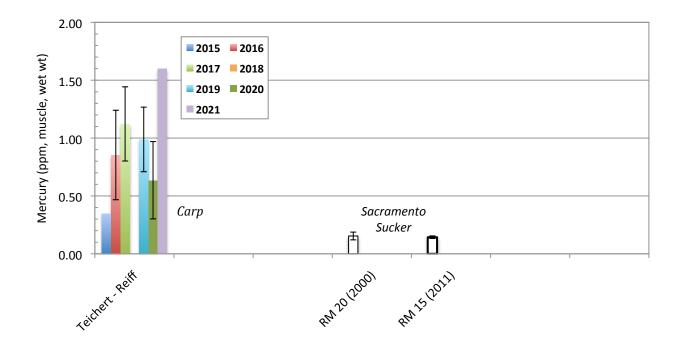


Figure 52. Carp summary data (all sites) and historic creek comparisons (mean fillet muscle mercury, with 95% confidence intervals)

Small, Young Fish Samples (note lower concentration scales)

Table 53. Juvenile Bass summary data (all sites) and historic creek comparisons

(means of multiple whole-body, multi-individual composite samples) 'n' = number: number of composite samples; number of individual fish per composite

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	$\mathbf{Hg} \ (\mu g/g = ppm, wet wt)$	Std. Error
Cemex – Phase 1	2015	4	8	109	17	0.044	± 0.004
Cemex – Phase 1	2016	4	3	102	17	0.094	± 0.006
Cemex – Phase 1	2017	4	2	117	22	0.146	± 0.011
Cemex – Phase 1	2018	1	1	78	6	0.068	
Cemex – Phase 1	2019	4	4-5	106	17	0.114	± 0.007
Cemex – Phase 1	2020	5	2-4	100	13	0.104	± 0.008
Cemex – Phase 1	2021	4	2-3	79	7	0.096	± 0.006
Cemex – Phase 3-4	2015	4	7	108	16	0.334	± 0.026
Cemex – Phase 3-4	2016	4	2	114	18	0.372	± 0.026
Cemex – Phase 3-4	2017	4	2-3	108	16	0.249	± 0.016
Cemex – Phase 3-4	2018	_	_				
Cemex – Phase 3-4	2019	1	1	125	23	0.336	
Cemex – Phase 3-4	2020	1	1	124	23	0.144	
Cemex – Phase 3	2021	3	1	110	12	0.302	± 0.016
Cemex – <u>Phase 4</u>	2021	4	2-4	94	12	0.160	± 0.015
Teichert-Esparto – Reiff	2017	4	1-2	137	32	0.798	± 0.094
Teichert-Esparto – Reiff	2018	4	4-6	111	17	0.445	± 0.069
Teichert-Esparto – Reiff	2019	4	5	107	15	0.297	± 0.010
Teichert – <u>Esparto</u>	2020	4	3	92	9	0.472	± 0.027
Teichert – Esparto	2021	4	1	98	12	1.064	± 0.041
Teichert-Woodland – Storz	2017	4	1	143	35	0.337	± 0.030
Teichert-Woodland – Storz	2018	_	-	100	20	0.101	0.000
Teichert-Woodland – Storz	2019	4	1	130	29	0.131	± 0.036
Teichert-Woodland – Storz	2020	4	1	172	63	0.097	± 0.005
Teichert-Woodland – Storz	2021	3	2-3	125	29	0.118	± 0.009
Syar – B1	2015	4	7	159	44	0.589	± 0.015
Syar – B1	2016	4	10	74	5	0.524	± 0.060
Syar – B1	2017	4	1-2	102	18	0.461	± 0.087
Syar – B1	2018	4	2	88	9	0.368	± 0.020
Syar – B1	2019	4	1	87	7	0.338	± 0.021
Syar – B1	2020	4	1	87	9	0.259	± 0.055
Syar – B1	2021	3	1	82	7	0.363	± 0.031
		(cont	inued next [page)			

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Syar – West	2017	2	1	123	27	0.418	± 0.030
Syar – West	2018	4	2	77	6	0.153	± 0.024
Syar – West	2019	2	1	96	11	0.273	± 0.006
Syar – West	2020	4	4-5	106	17	0.114	± 0.007
Syar – West	2021	2	1	55	2	0.176	± 0.007
Historic/Baseline Da	ta						
River Mile 28 River Mile 15	2011 2011	4 3	3-5 1	75 93	6 10	0.142 0.050	± 0.013 ± 0.014

(Table 53, continued)

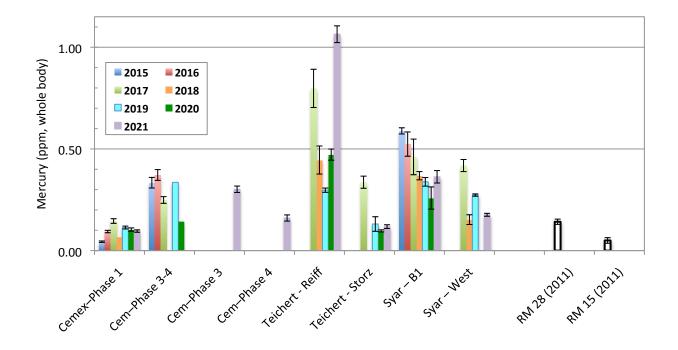


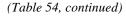
Figure 53. Juvenile Bass summary data (all sites) and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 54. Juvenile Sunfish summary data (all sites) and historic creek comparisons

(means of multiple whole-body, multi-individual composite samples) '**n**' = number: number of composite samples; number of individual fish per composite

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Erro
Green Sunfish (unless n	oted Blu	egill) – juv	eniles				
Cemex – Phase 1	2017	4	8-10	47	1.9	0.118	± 0.01
Cemex – Phase 1	2018	4	2	51	2.1	0.035	± 0.00
Cemex – Phase 1	2019	4	2-10	44	1.7	0.089	± 0.01
Cemex – Phase 1	2020	3	1-3	50	2.7	0.089	± 0.00
Cemex – Phase 1	2021	3	1-9	33	2.3	0.090	± 0.01
Cemex – Phase 3-4	2015	4	10	47	1.8	0.275	± 0.01
Cemex – Phase 3-4	2016	4	4-5	49	2.0	0.233	± 0.01
Cemex – Phase 3-4	2017	4	2-6	36	0.7	0.150	± 0.02
Cemex – Phase 3-4	2018	4	1	34	0.5	0.112	± 0.01
Cemex – Phase 3-4	2019	4	10	43	1.6	0.185	± 0.01
Cemex – Phase 3-4	2020	4	1-12	38	0.9	0.117	± 0.02
Cemex – Phase 3	2021	3	2-5	33	0.6	0.159	± 0.02
Cemex – <u>Phase 4</u>	2021	4	3-10	44	0.5	0.160	± 0.02
Teichert-Esparto – Reiff	2015	_	1	68	2.7	0.241	
Teichert-Esparto – Reiff	2016	_	_				
Teichert-Esparto – Reiff	2017	_	_				
Teichert-Esparto – Reiff	2018	4	2	48	2.3	0.252	± 0.01
Teichert-Esparto – Reiff	2019	4	3-10	41	1.3	0.187	± 0.02
Teichert – <u>Esparto</u>	2020	4	3	35	0.7	0.230	± 0.01
Teichert – <u>Esparto</u>	2021	3	1	39	1.2	0.479	± 0.01
Teichert-Woodland – Storz *Bluegill*	2021	4	3-10	38	1.3	0.020	± 0.00
Syar – B1	2015	4	8-9	47	1.7	0.325	± 0.04
Syar – B1	2016	4	4	50	1.9	0.414	± 0.03
Syar – B1	2017	4	6-7	40	1.0	0.225	± 0.03
Syar – B1	2018	4	10	37	0.8	0.231	± 0.02
Syar – B1	2019	4	8-10	45	1.5	0.245	± 0.01
Syar – B1 *Bluegill*	2020	4	12	44	1.3	0.181	± 0.01
Syar – B1 *Bluegill*	2021	4	10-12	45	1.3	0.222	± 0.00
Syar – West	2017	4	5-10	45	1.7	0.237	± 0.03
Syar – West	2018	4	2-4	34	0.6	0.102	± 0.00
Syar – West	2019	4	8-10	46	1.5	0.177	± 0.01
Syar – West *Bluegill*	2020	4	10-12	42	1.2	0.187	± 0.02
Syar – West *Bluegill*	2021	4	10	45	1.1	0.254	± 0.00
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Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Historic/Baseline	Data						
River Mile 28	2011	4	4	53	2.8	0.139	± 0.007
River Mile 20	2011	4	4	58	3.4	0.084	± 0.002
River Mile 17	2000-2002	8	5-10	41-90	1-6	0.169	± 0.013
River Mile 15	2000-2002	8	4-8	40-87	1-6	0.117	± 0.005
River Mile 15	2011	4	4-5	56	3.1	0.086	± 0.009



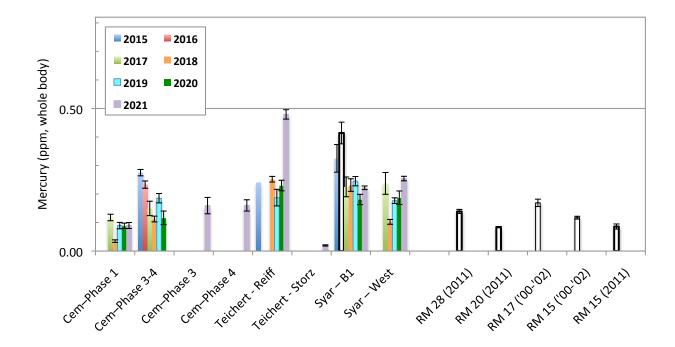


Figure 54. Juvenile Sunfish summary data (all sites) and historic creek comparisons (means of multiple whole-body, multi-individual composite samples)

Table 55. Mosquitofish summary data (all sites) and historic creek comparisons

(means of multiple whole-body, multi-individual composite samples) '**n**' = number: number of composite samples; number of individual fish per composite

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Cemex – Phase 1	2015	4	10	39	0.6	0.075	± 0.008
Cemex – Phase 1	2016	4	10	34	0.4	0.093	± 0.019
Cemex – Phase 1	2017	4	10	33	0.4	0.135	± 0.019
Cemex – Phase 1	2018	4	6-10	34	0.5	0.083	± 0.016
Cemex – Phase 1	2019	4	10	34	0.5	0.096	± 0.024
Cemex – Phase 1	2020	4	12	35	0.5	0.102	± 0.021
Cemex – Phase 1	2021	4	8-10	34	0.5	0.115	± 0.018
Cemex – Phase 3-4	2015	4	10	37	0.6	0.228	± 0.029
Cemex – Phase 3-4	2016	4	10	37	0.6	0.157	± 0.019
Cemex – Phase 3-4	2017	4	6-10	34	0.5	0.286	± 0.035
Cemex – Phase 3-4	2018	4	3-10	34	0.5	0.203	± 0.021
Cemex – Phase 3-4	2019	4	10	35	0.6	0.183	± 0.029
Cemex – Phase 3-4	2020	4	3-12	33	0.4	0.112	± 0.018
Cemex – <u>Phase 3</u>	2021	4	10	35	0.5	0.137	± 0.022
Cemex – <u>Phase 4</u>	2021	4	6-10	35	0.5	0.122	± 0.015
Teichert-Esparto – Reiff	2015	4	12	38	0.6	0.094	± 0.005
Teichert-Esparto – Reiff	2016	4	10	36	0.5	0.212	± 0.021
Teichert-Esparto – Reiff	2017	—	_				
Teichert-Esparto – Reiff	2018	4	10	35	0.5	0.262	± 0.026
Teichert-Esparto – Reiff	2019	4	5-10	33	0.46	0.222	± 0.041
Teichert – <u>Esparto</u>	2020	4	1-12	37	0.7	0.267	± 0.033
Teichert – <u>Esparto</u>	2021	3	12	37	0.7	0.579	± 0.033
Teichert-Esparto – Mast	2017	8	10	35	0.5	0.312	± 0.046
Teichert-Esparto – Mast	2018	8	10	34	0.5	0.182	± 0.015
Teichert-Esparto – Mast	2019	8	10	34	0.5	0.287	± 0.058
Teichert-Woodland – Storz	2016	4	10	35	0.5	0.229	± 0.054
Teichert-Woodland – Storz	2017	4	8-10	29	0.2	0.282	± 0.011
Teichert-Woodland – Storz	2018	4	10	30	0.3	0.087	± 0.017
Teichert-Woodland – Storz	2019	4	6-10	33	0.4	0.200	± 0.018
Teichert-Woodland - Storz	2020	4	12	32	0.4	0.059	± 0.008
Teichert-Woodland - Storz	2021	4	10	35	0.6	0.082	± 0.026
Syar – B1	2015	4	5-10	31	0.3	0.268	± 0.022
Syar – B1	2016	_	_				
Syar – B1	2017	4	9-10	35	0.4	0.309	± 0.055
Syar – B1	2018	4	6-9	31	0.4	0.163	± 0.028
Syar – B1	2019	3	1-3	38	0.7	0.214	± 0.011
Syar – B1	2020	4	6-12	34	0.5	0.130	± 0.026
Syar – B1	2021	4	4-10	34	0.5	0.156	± 0.015
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Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Syar – West	2017	4	10	34	0.4	0.236	± 0.034
Syar – West	2018	4	6-7	29	0.3	0.088	± 0.012
Syar – West	2019	3	2-3	36	0.6	0.165	± 0.032
Syar – West	2020	4	8-12	35	0.5	0.109	± 0.018
Syar – West	2021	4	5-10	34	0.4	0.121	± 0.018
Historic/Baseline D	ata						
River Mile 17	2000-2002	13	5-30	26-47	0.2-1.1	0.178	± 0.020
River Mile 15	2000-2002	10	5-30	26-47	0.2-1.0	0.100	± 0.01
River Mile 15	2011	4	1-10	37	0.7	0.103	± 0.02

(Table 55, continued)

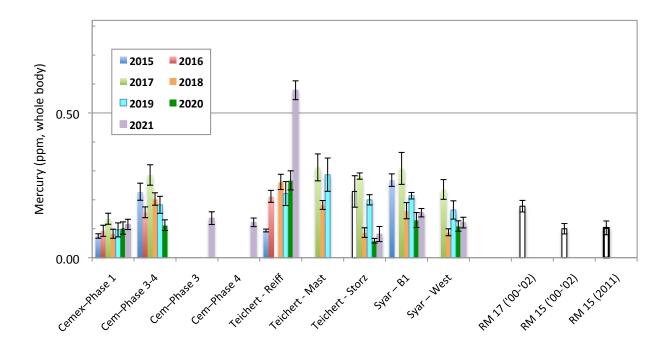


Figure 55. Mosquitofish summary data (all sites), and historic creek comparisons *(means of multiple whole-body, multi-individual composite samples)*

Table 56. Red Shiner summary data (all sites), and historic creek comparisons

(means of multiple whole-body, multi-individual composite samples) '**n**' = number: number of composite samples; number of individual fish per composite

Site	Year	n (comps)	n (inds/ (comp)	Av Lgth (mm total)	Av Wt (grams)	Hg (μ g/g = ppm, wet wt)	Std. Error
Red Shiners							
Teichert-Esparto – Reiff	2015	4	10	50	1.3	0.152	± 0.009
Teichert-Esparto – Reiff	2016	4	10	47	1.1	0.412	± 0.042
Teichert-Esparto – Reiff	2017	4	10	49	1.1	0.695	± 0.070
Teichert-Esparto – Reiff	2018	4	10	45	0.8	0.556	± 0.031
Teichert-Esparto – Reiff	2019 (Shi	ners not fou	und in 2019	or 2020)			
Teichert – Esparto (Reiff)	2021	3	4	54	1.8	1.310	± 0.034
Teichert – <u>Esparto</u> (Mast NW	() 2021	3	1-5	51	1.4	0.951	± 0.259
Historic/Baseline Data							
River Mile 28	2011	4	10	48	1.0	0.242	± 0.018
River Mile 20	2000	3	9	42	0.6	0.166	± 0.002
River Mile 17	2000-2002	11	6-15	27-58	0.2-1.8	0.225	± 0.023
River Mile 15	1997	3	19	37	0.5	0.159	± 0.014
River Mile 15	2000-2002	13	6-12	30-60	0.2-2.0	0.131	± 0.005

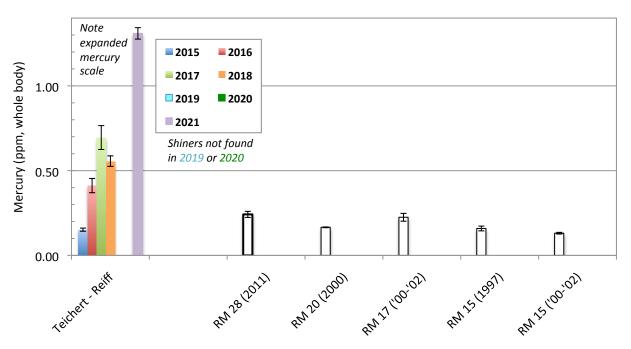


Figure 56. Red Shiner summary data (all sites), and historic creek comparisons *(means of multiple whole-body, multi-individual composite samples)*

DISCUSSION AND CONCLUSIONS

The Yolo County Ordinance for mercury in aggregate mining ponds was revised and updated in December 2019 (Sec. 10-4.420.1 – 10-5.517 Mercury Bioaccumulation in Fish). The full, updated text is attached below as Appendix A. Fish monitoring results have been assessed, since 2019, in relation to the updated Ordinance measures. The updated Ordinance calls for action based on three to five years of data, as follows:

If, during the mining phase of monitoring, the pit lake's average fish tissue mercury concentration exceeds the ambient mercury level for any three of five monitoring years, annual monitoring shall continue for an additional five years, and the operator shall undertake expanded analysis pursuant to subsection (f) and preparation of a lake management plan pursuant to subsection (g). Sec. 10-5.517(e)(2).

The "exceeds the ambient mercury level" above refers to whether pond fish mercury levels are found to be significantly elevated above corresponding Cache Creek Baseline samples – in three of five monitoring years. The summary table below shows overall annual results of fish mercury testing in the monitored ponds, in relation to ambient fish mercury levels.

Pit Sites	2015	2016	2017	2018	2019	2020	2021
Cemex – Phase 1	≦	≦ _	1	٤	≦ _	≤	
Cemex – Phase 3				^			>
Cemex – Phase 4	>	>	>	1	>	i≦.	INC
Taishant Fananta Daiff	INIC						
Teichert-Esparto – Reiff	INC	>	>	>	>	>	>
Teichert-Esparto – Mast			>	INC	>		
Teichert-Woodland – Storz			INC	×.	×.	×.	≤
6 B4		200			100		
Syar – B1	>	>	>	>	>	>	>
Syar – West			>	VI	>	>	>
\leq = at or below ambient INC = inconclusive > = elevated over ambient							

Fish Mercury Monitoring Summary – All Sites, 2015-2021

Annual monitoring of fish mercury levels began in 2015 at four aggregate mining ponds: Cemex–Phase 1, Cemex–Phase 3-4, Teichert-Esparto–Reiff, and Syar–B1. By 2018, with four years of data from the initial four monitored ponds, three were found to be elevated in fish mercury in three or more years: Cemex–Phase 3-4, Teichert-Esparto–Reiff, and Syar–B1. These three ponds have remained elevated above baseline through nearly all of the monitoring. The Cemex–Phase 1 Pond, in contrast, has been consistently low in fish mercury (relatively). It was chosen as a control/reference pond, as specified in the ordinance. <u>Beginning in 2018, "expanded analyses" were added to the program at these four ponds and routine fish monitoring was extended by five years</u>.

Three other ponds were added to the fish monitoring in 2017: Teichert-Woodland–Storz, Syar–West, and Teichert-Esparto–Mast. Teichert-Esparto Mast was later combined with Reiff Pond into the current Esparto Pond, which continues to be monitored for fish mercury and expanded analyses. There are now five years of fish data for Teichert-Woodland–Storz and Syar–West. Storz has been identified as another consistently lower mercury site, not requiring expanded analyses other than routine water profiling once per year in fish monitoring years. <u>With five years of non-elevated</u> status, fish monitoring at Storz can shift to once every two years. Syar–West was flagged in 2020 for required expanded analyses beginning this year and an additional five years of fish testing. Expanded analyses actually began here earlier; Syar–West was tested as a second control/reference pond, important for its depth which more closely matches projected final post-reclamation pond depths at some of the sites. The timelines of water profiling and other project components are summarized in tables at the end of this section.

For the ponds flagged as significantly elevated over ambient in three of five years, the Ordinance states:

... the operator shall undertake expanded analysis pursuant to subsection (f) and preparation of a lake management plan pursuant to subsection (g). Sec. 10-5.517(e)(2).

The "expanded analysis" task is meant to precede and provide guidance for the "preparation of a lake management plan". Because of the complexities of the methylmercury cycle and the unique configuration, depth, chemistry, and biology of each individual pond, additional information is

needed to help craft site-specific management approaches that are likely to be effective. The first steps are to 1) broadly characterize the bottom sediments of the pond and 2) initiate seasonal water column profiling of a range of potentially relevant water quality parameters.

1. Characterize pond bottom sediment

For the ponds that have been flagged for expanded analysis and development of lake management plans, and the required control/reference pond, some basic information about the bottom sediments is essential, to see if there are any large differences between the ponds that could help account for the mercury bioaccumulation patterns. Sediment sampling was conducted in Fall 2018 at the 3 ponds identified as elevated in fish mercury at that time, plus the identified control site Cemex–Phase 1. The Syar–West pond was also sampled, making five ponds in total for initial sediment characterization. As specified in the Ordinance, for each pond, six independent bottom samples were taken from locations distributed across the pond, specifically of fine-grained surficial sediments (top 2 cm). These were analyzed for total mercury and organic matter content.

The bottom sediment mercury data ranged between mean levels of 0.266 and 0.518 ppm, across all five ponds tested. These levels were similar to the 0.390 ppm average from the USGS Settling Basin studies. There was a small, approximate two-fold range between lowest and highest concentrations. The ponds were elevated above 'clean/background' levels, as is to be expected for this watershed. Sediment mercury around upstream source areas ranges into the hundreds of parts per million. The report for the 2018 sediment work concluded:

"... But the two lowest sediment mercury sites, Cemex–Phase 1 and Teichert-Esparto–Reiff, included both the lowest and the highest fish mercury conditions. Clearly, the ranges of sediment mercury levels present in these ponds are <u>all</u> more than enough to potentially lead to elevated fish mercury levels. The low fish mercury at the Cemex–Phase 1 pond and very high fish mercury at Teichert-Esparto–Reiff, with nearly identical sediment mercury at both, strongly suggests that <u>other conditions</u> of the ponds are more important. This is an advance that will help guide potential management directions. These initial sediment characterization tests were looking for potentially dramatic sediment mercury trends that were much higher than baseline and/or vastly different between ponds. That has been ruled

out. This points management ideas more toward modification of other pond conditions that may lead to differences in methylmercury production and transfer, and to the large differences seen in fish mercury levels. The accompanying water column profiling work seeks to identify some of these possible factors."

It is possible that additional or different sediment analyses may be warranted in the future to help determine appropriate management approaches.

2. Initiate water column profiling

For the ponds that have been flagged for expanded analysis and development of lake management plans, and the required control/reference pond, the Ordinance outlines:

The analysis shall include expanded lake water column profiling (a minimum of five profiles per affected wet pit lake plus one or more nonaffected lakes for control purposes) conducted during the warm season (generally May through October) in an appropriate deep profiling location for each pit lake. The following water quality parameters shall be collected at regular depth intervals, from surface to bottom of each lake, following protocols identified in subsection (a): temperature, dissolved oxygen, conductivity, pH and oxidation-reduction potential (ORP), turbidity or total suspended solids, dissolved organic matter, and algal density by Chlorophyll or Phycocyanin.

Water column profiling began in 2018, as described above. The three identified elevated-mercury ponds and the lower-mercury control/reference pond have been tested seasonally, five times per year between May and October. The Syar-West Pond was also studied as a deep pond control, added in 2019. Profiling continued at these sites in 2021. Results are presented in accompanying water reports. Excerpting from the conclusions of recent water profiling reports:

"Some of the greatest accumulations or changes were found in the lower water of ponds that stratified thermally. Most of the monitored ponds were too shallow to stratify completely (isolate water layers from each other) in the warm season but two, Teichert-Esparto–Reiff and Syar–B1, stratified enough for many of the measured water parameters to shift significantly, including oxygen, pH, and ORP, with deep accumulations of turbidity and algal cells."

"Among the three ponds identified as elevated in fish mercury – Syar–B1, Teichert-Esparto– Reiff, and Cemex–Phase 3-4 – there was not a single, consistent trend. While the two most elevated ponds, Syar–B1 and Teichert-Esparto–Reiff have consistently shown evidence of seasonal water column anoxia, that was not the case at Cemex-Phase 3-4. The new data from the much deeper Syar–West pond confirmed it as a site of strong seasonal water stratification and bottom water anoxia (loss of oxygen)."

"At this point with the new water profiling data, seasonal bottom water anoxia – or its absence – appears to be an important link to the observed fish mercury trends. Since seasonal anoxia is known to enhance the production of methylmercury and its movement into fish, management approaches that disrupt that pattern may reduce the problem. This is something to consider for ponds identified as elevated in fish mercury and requiring management. The profiling results to-date support management approaches that could provide summer mixing and the disruption of bottom water anoxia – specifically for ponds that require mercury management and that have seasonal anoxia. The case of Cemex–Phase 3-4 though, with high fish mercury but no seasonal anoxia, is a reminder that there may not be any single 'magic bullet' management approach; different approaches may be needed at different sites. Many different physical, chemical, and biological factors can influence the mercury cycle in each pond. Seasonal anoxia is the most straightforward one to tackle – when it is present. When it isn't, and fish mercury is still elevated, other mechanisms will need to be identified for possible alternate management approaches. This water column profiling is an important step to better understand the options."

The fish monitoring itself has also highlighted factors that could be significant for lake management. In particular, the bass mercury trend at Teichert–Esparto, in relation to changing prey species, supports the idea that food web structure may significantly impact mercury accumulation in the top predator fish. Additionally, fish mercury trends have been observed over

the years in relation to the presence or absence of active mining or processing plant slurry flows. These processes that suspend sediment particles into the water have been associated with declines in fish mercury uptake rather than increases, presumably by placing alternate binding sites into the water for methylmercury, deflecting some of it from foodweb pathways. And, in contrast, clear water conditions with low suspended solids have tended to be associated with relative increases in fish mercury uptake.

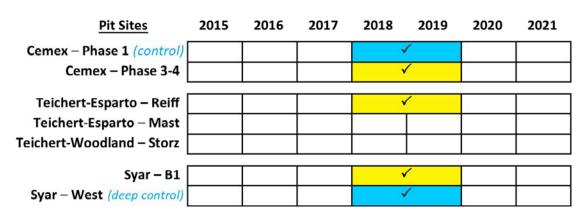
Fish monitoring and seasonal water column testing will continue at the designated ponds. Ongoing findings will continue to narrow down management options for the sites requiring lake management plans and action. <u>At this point (late 2022 at the time of this reporting), with 3-4</u> <u>years of additional information at Cemex–Phase 3 and 4, Teichert–Esparto, Syar–B1, and Syar–</u> <u>West, it is time to develop Lake Management Plans for those sites</u>.

Status of Other Components of the Mercury Monitoring Program

Pit Sites 2015 2016 2017 2018 2019 2020 2021 Cemex - Phase 1 (control) √ √ ~ √ Cemex – Phase 3 ~ ~ 1 √ Cemex – Phase 4 √ \checkmark Teichert-Esparto – Reiff ~ ~ 1 Teichert-Esparto – Mast Teichert-Woodland – Storz (1x)Syar – B1 \checkmark \checkmark ~ \checkmark Syar – West (control to 2020) 1 ~ \checkmark 1

<u>Water Column Profiling</u> (elevated sites and controls)

Bottom Sediment Collections (single event, elevated sites and controls)



Reports Completed

Report	2015	2016	2017	2018	2019	2020	2021
Fish Mercury Monitoring	Final	Final	Final	Final	Final	Draft	Draft
Water Column Profiling			10	Final	Final	Draft	Draft
Bottom Sediments (1x)				Final			

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- Yolo County Code, Title 10. Chapter 5 (Surface Mining Reclamation), Section 10.5.517 Mercury Bioaccumulation in Fish. 2019 Revision.

APPENDIX A

Yolo County, CA Code of Ordinances

Sec. 10-4.420.1 – 10-5.517 Mercury Bioaccumulation in Fish – December 2019 Update and Revision – <u>Yolo County Mining Ordinance, Sec.10-4.420.1 Mercury Bioaccumulation in Fish.</u> Each mining area to be reclaimed to a permanent lake as part of each approved longrange mining plan shall be evaluated annually by the operator for five years after the pit fills with groundwater with an intensive fish mercury monitoring program described in Section 10-5.517 of the Reclamation Ordinance.

Reclamation Ordinance, Sec. 10-5.517. Mercury bioaccumulation in Fish.

As part of each approved long-term mining plan involving wet pit mining to be reclaimed to a permanent pond, lake, or water feature, the operator shall maintain, monitor, and report to the Director according to the standards given in this section. Requirements and restrictions are distinguished by phase of operation as described below.

(a) <u>Mercury Protocols.</u> The Director shall issue and update as needed "Lower Cache Creek Off-Channel Pits Mercury Monitoring Protocols" (Protocols), which shall provide detailed requirements for mercury monitoring activities. The Protocols shall include procedures for monitoring conditions in each pit lake, and for monitoring ambient mercury level in the lower Cache Creek channel within the CCAP planning area, as described below. The Protocols shall be developed and implemented by a qualified aquatic scientist or equivalent professional acceptable to the Director. The Protocols shall identify minimum laboratory analytical reporting limits, which may not exceed the applicable response threshold identified in subsection (e) below. Data produced from implementing the Protocols shall meet or exceed applicable standards in the industry.

(b) <u>Ambient Mercury Level</u>. The determination of the ambient or "baseline" fish mercury level shall be undertaken by the County every ten years in years ending in 0. This analysis shall be undertaken by the County for use as a baseline of comparison for fish mercury testing conducted in individual wet mining pits. The work to establish this baseline every ten years shall be conducted by a qualified aquatic systems scientist acceptable to the Director and provided in the form of a report to the Director. It shall be paid for by the mining permit operators on a fair-share basis. The results of monitoring and evaluation of available data shall be provided in the report to substantiate the conclusions regarding ambient concentrations of mercury in fish within the lower Cache Creek channel within the CCAP planning area.

(c) <u>Pit Monitoring.</u>

(1) <u>Mining Phase</u> (including during idle periods as defined in SMARA). The operator shall monitor fish and water column profiles in each pit lake once every year during the period generally between September and November for the first five years after a pit lake is created. Fish monitoring should include sport fish where possible, together with other representative species that have comparison samples from the creek and/or other monitored ponds. Sport fish are defined as predatory, trophic level four fish such as bass, which are likely to be primary angling targets and have the highest relative mercury levels. The requirements of this subsection apply to any pit lake that is permanently wet and navigable by a monitoring vessel. If, in the initial five years after the pit lake is created, the applicable response threshold identified in subsection (e) is exceeded in any three of five monitoring years, the operator shall, solely at their own expense, undertake expanded analysis pursuant to subsection (f) and preparation of a lake management plan pursuant to subsection (g).

- (2) <u>Reclamation Phase</u>. No monitoring is required after mining has concluded, during the period that an approved reclamation plan is being implemented, provided reclamation is completed within the time specified by SMARA or the project approval, whichever is sooner.
- (3) <u>Post-Reclamation Phase</u>. After reclamation is completed, the operator shall monitor fish and water column profiles in each pit lake at least once every two years during the period of September-November for ten years following reclamation. Monitoring shall commence in the first calendar year following completion of reclamation activities. If fish monitoring results from the post-reclamation period exceed the applicable response threshold described in subsection (e) or, for ponds that have implemented mitigation management, results do not exhibit a general decline in mercury levels, the operator shall, solely at their own expense, undertake expanded analysis pursuant to subsection (f) and preparation of a lake management plan pursuant to subsection (g).
- (4) <u>Other Monitoring Obligation</u>. If monitoring conducted during both the mining and post-reclamation phase did not identify any exceedances of the ambient mercury level for a particular pit lake, and at the sole discretion of the Director no other relevant factors substantially support that continued monitoring is merited, the operator shall have no further obligations.

(d) <u>Reporting</u>.

- (1) <u>Pit Monitoring Results</u>. Reporting and evaluating of subsection (c) pit monitoring results shall be conducted by a qualified aquatic scientist or equivalent professional acceptable to the Director. Monitoring activities and results shall be summarized in a single report (addressing all wet pit lakes) and submitted to the Director within six months following each annual monitoring event. The report shall include, at a minimum: (1) results from subsection (b) (pit monitoring), in relation to subsection (a) (ambient mercury levels).
- (2) <u>Expanded Analysis Results</u>. Reporting and evaluation of subsection (f) expanded analysis shall be conducted by a qualified aquatic scientist or equivalent professional acceptable to the Director. Results shall be summarized in a single report (addressing all affected wet pit lakes) and submitted to the Director within six months following

each annual monitoring event. The report shall include, at a minimum, the results of the expanded analysis undertaken pursuant subsection (f).

- (3) <u>Data Sharing</u>. For pit lakes open to the public, the Director may submit the data on mercury concentrations in pit lake fish to the state Office of Environmental Health Hazard Assessment (or its successor) for developing site-specific fish consumption advisories.
- (e) <u>Response Thresholds</u>.
 - (1) <u>Fish Consumption Advisory</u>. If at any time during any phase of monitoring the pit lake's average sport fish tissue mercury concentration exceeds the Sport Fish Water Quality Objective, as it may be modified by the state over time (as of 2019, the level was 0.2 mg/kg), the operator shall post fish consumption advisory signs at access points around the lake and around the lake perimeter. Catch-and-release fishing may still be allowed. Unless site-specific guidance has been developed by the state's Office of Health Hazard Assessment or the County, statewide fish consumption guidance shall be provided.
 - (2) <u>Mining Phase Results</u>. If, during the mining phase of monitoring, the pit lake's average fish tissue mercury concentration exceeds the ambient mercury level for any three of five monitoring years, annual monitoring shall continue for an additional five years, and the operator shall undertake expanded analysis pursuant to subsection (f) and preparation of a lake management plan pursuant to subsection (g).
 - (3) <u>Post-Reclamation Phase Results</u>. If during the first ten years of the post-reclamation phase of monitoring, the pit lake's average fish tissue mercury concentration exceeds the ambient mercury level for any three of five monitoring years, biennial monitoring shall continue for an additional ten years, and the operator shall undertake expanded analysis pursuant to subsection(f) and preparation of a lake management plan pursuant to subsection (g).
- (f) Expanded Analysis.
 - (1) <u>General</u>. If during the mining or post-reclamation phase, any pit lake's average fish tissue mercury concentration exceeds the ambient mercury level for any three years, the operator shall undertake expanded analyses. The analysis shall include expanded lake water column profiling (a minimum of five profiles per affected wet pit lake plus one or more nonaffected lakes for control purposes) conducted during the warm season (generally May through October) in an appropriate deep profiling location for each pit lake. The following water quality parameters shall be collected at regular depth intervals, from surface to bottom of each lake, following protocols identified in subsection (a): temperature, dissolved oxygen, conductivity, pH and oxidation-reduction potential (ORP), turbidity or total suspended solids, dissolved organic

matter, and algal density by Chlorophyll or Phycocyanin. The initial analysis shall also include one-time collections of fine grained (clay/silt) bottom sediments from a minimum of six well distributed locations for each affected lake, and from one or more non-affected lakes for control purposes, to be analyzed for mercury and organic content.

- (2) Scope of Analysis. The purpose of the expanded analyses is to identify and assess potential factors linked to elevated methylmercury production and/or bioaccumulation in each pit lake. The scope of the expanded analyses shall include monitoring and analysis appropriate to fulfill this purpose, invoking best practices in the industry. In addition to the analyses described in subsection (f)(1) above, the analysis should also consider such factors as: electrical conductivity, bathymetry (maximum and average depths, depth-to-surface area ratios, etc.), and trophic status indicators (concentrations, Secchi depth, chlorophyll a, fish assemblages, etc.). Additional types of testing may be indicated and appropriate if initial results are inconclusive.
- (3) <u>Use of Results</u>. The results of the expanded analyses undertaken pursuant to this subsection shall be used to inform the preparation of a lake management plan described below under subsection (g).

(g) Lake Management Activities

- (1) <u>General</u>. If monitoring conducted during the mining or post-reclamation phases triggers the requirement to undertake expanded analysis and prepare and implement a lake management plan, the operator shall implement lake management activities designed by a qualified aquatic scientist or equivalent professional acceptable to the Director, informed by the results of subsection (f). Options for addressing elevated mercury levels may include (A) and/or (B) below at the Director's sole discretion and at the operator's sole expense.
 - (A) <u>Lake Management Plan</u>. Prepare a lake management plan that provides a feasible, adaptive management approach to reducing fish tissue mercury concentrations to at or below the ambient mercury level. Potential mercury control methods could include, for example: addition of oxygen to or physical mixing of anoxic bottom waters; alteration of water chemistry (modify pH or organic carbon concentration); and/or removal or replacement of affected fish populations. The lake management plan may be subject to external peer review at the discretion of the Director. Lake management activities shall be appropriate to the phase of the operation (e.g., during mining or post-reclamation). The Lake Management Plan shall include a recommendation for continued monitoring and reporting. All costs associated with preparation and implementation of the lake management plan shall be solely those of the operator. Upon acceptance by the Director, the operator shall immediately implement the plan. The lake management plan shall generally be implemented within three years of reported results from the expanded analyses resulting from subsection (f). If lake management does not achieve

acceptable results and/or demonstrate declining mercury levels after a maximum of three years of implementation, at the sole discretion of the Director, the operator may prepare an alternate management plan with reasonable likelihood of mitigating the conditions.

(B) <u>Revised Reclamation Plan</u>. As an alternative to (A), or if (A) does not achieve acceptable results and/or demonstrate declining mercury levels after a maximum of three years of implementation, at the sole discretion of the Director, the operator shall prepare and submit revisions to the reclamation plan (including appropriate applications and information for permit amendment) to fill the pit lake with suitable fill material to a level no less than five (5) feet above the average seasonal high groundwater level, and modify the end use to agriculture, habitat, or open space at the discretion of the Director, subject to Article 6 of the Mining Ordinance and/or Article 8 of the Reclamation Ordinance as may be applicable.

(2) Implementation Obligations.

- (A) If a lake management plan is triggered during the mining or post-reclamation phase and the subsequent lake management activities do not achieve acceptable results and/or demonstrate declining mercury levels, the operator may propose different or additional measures for consideration by the Director and implementation by the operator, or the Director may direct the operator to proceed to modify the reclamation plan as described in subsection (g)(1)(B).
- (B) Notwithstanding the results of monitoring and/or lake management activities during the mining phase, the operator shall, during the post-reclamation phase, conduct the required ten years of biennial monitoring.
- (C) If monitoring conducted during the post-reclamation phase identifies three monitoring years of mercury concentrations exceeding the ambient mercury level, the operator shall implement expanded analyses as in subsection (f), to help prepare and implement a lake management plan and associated monitoring.
- (D) If subsequent monitoring after implementation of lake management activities, during the post-reclamation phase, demonstrates levels of fish tissue mercury at or below the ambient mercury level for any three monitoring years (i.e., the management plan is effective), the operator shall be obligated to continue implementation of the plan and continue monitoring, or provide adequate funding for the County to do both, in perpetuity.

APPENDIX B

PHOTOS FROM THE 2021 FISH MONITORING

GENERAL FIELD WORK, AND EXAMPLES OF MAIN ADULT FISH



Seining for small fish



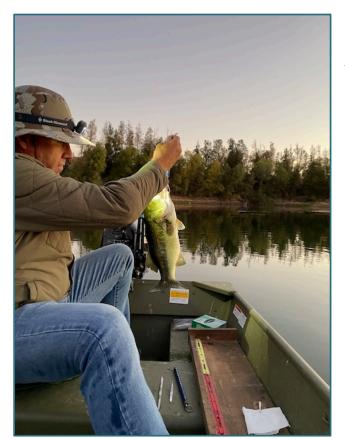
Baited set-lines for Catfish (White Catfish)



Angling for bass at key times



Spearfishing as last resort; 2021 Carp



Largemouth Bass

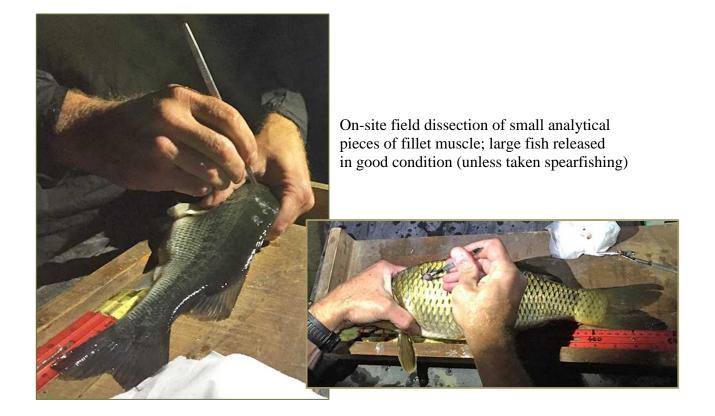
Measuring weight and length





(Channel) Catfish

Carp



(photos this page from earlier reports)

Muscle sample into preweighed vial; stored on ice in field.

On return, careful re-weigh of vial with sample, to get exact sample weight.

Then into lab freezer, until mercury analysis



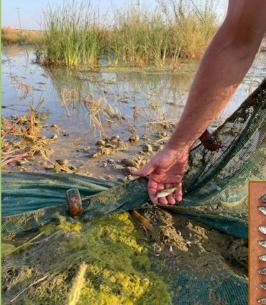
CEMEX – PHASE 1 POND



Water levels were maintained during drought in the Cemex ponds

Adult Largemouth Bass





Seining for small fish



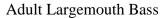


Juvenile bass and sunfish

Mosquitofish

CEMEX – PHASE 3 POND







Seining for small fish

Mosquitofish (top) Juvenile Bass and juv. Green Sunfish (right)

CEMEX – PHASE 4 POND



The actively mined Cemex pond



Small fish in seine



Adult Largemouth Bass



Juvenile Bass



Juvenile Green Sunfish

TEICHERT – ESPARTO POND



By September, drought (photo from north)



Adult Largemouth Bass



White Catfish

(Esparto Pond, continued)





Mosquitofish

Reiff Basin

Juvenile Bass, Green Sunfish, Red Shiners





Mosquitofish

Red Shiners

Mast Northwest (Central) Basin

TEICHERT-WOODLAND – STORZ POND



Adult Largemouth Bass



Mosquitofish

Seining for small fish



(Storz Pond, continued)



Northeast Basin





Seining for small fish





Juvenile Bass

Mosquitofish

SYAR – B1 POND



Drought Year 2, water level dropped deeply; see former water line



Channel Catfish, Largemouth Bass



Swim-seining for small fish

Juvenile Bluegill



Juvenile Bass, Mosquitofish

SYAR-WEST POND



Drought-lowered water level created cliff-edge shores



Adult Bass taken spearfishing



Juvenile Bass, Mosquitofish

Swim-seining for small fish



